

1 Ex situ cultivation impacts on plant traits and drought stress response in a
2 multi-species experiment

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16 **Abstract**

17 Ex situ collections of wild plants of conservation value are supposed to preserve the
18 phenotypic variability of their wild source populations as well as their plastic responses to
19 environmental stress. However, genetic erosion and adaptive evolutionary changes during
20 cultivation are likely to impose strong constraints to those aims. To date, is it not known
21 whether cultivation affects plant trait variability and stress response of species in botanic
22 garden collections. We studied the effects of cultivation on the trait expression, genetic
23 trait variability, and drought stress response in 12 plant species cultivated in Meise Botanic
24 Garden, Belgium. We found that cultivation increased germination rate and leaf length
25 across all species, while it decreased flowering and delayed the beginning of the flowering
26 period in six drought-tolerant species. We furthermore found indications that plant
27 response to drought was reduced by cultivation in some performance variables, while
28 mortality due to drought slightly increased with cultivation. In three out of 10 traits
29 measured, genetic variability decreased with increasing cultivation time in the botanic
30 garden indicating a loss of evolutionary potential. Our results suggest that the preservation
31 of the phenotypic status and evolutionary potential in ex situ collections is a challenging
32 task and that the application of up-to-date protocols is decisive to achieve meaningful
33 conservation collections.

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36 Key words: ex situ conservation, botanic gardens, trait variability, evolutionary potential,
37 unconscious selection, adaptation

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41 1. Introduction

42 Preserving threatened plant species *ex situ* (i.e. outside their natural habitats) has grown
43 to a fundamental pillar in global conservation efforts (Donaldson, 2009; Maunder et al.,
44 2004; Mounce et al., 2017). In this context, botanic garden collections of seeds and living
45 plants of rare and endangered plants have become a more and more important resource
46 and now 40 % of the world's threatened species are somehow backed-up in botanic
47 gardens (Mounce et al., 2017). This demonstrates the potential of botanic garden's *ex situ*
48 collections in preventing the extinction of endangered plants, and even a resource of
49 returning species extinct in the wild (Abeli et al., 2019).

50 However, *ex situ* collections of plants in botanic gardens have also been associated with
51 several problematic processes, i.e. genetic erosion, inbreeding depression or adaptation to
52 the artificial cultivation conditions (Ensslin et al., 2015; Ensslin and Godefroid, 2019;
53 Havens et al., 2004; Schoen and Brown, 2001). These concerns arise from the fact that
54 most populations *ex situ* are very small, increasing adverse effects of genetic drift and the
55 likelihood of inbreeding (Ellstrand and Elam, 1993). Moreover, as plants in *ex situ*
56 collections are grown out of their natural conditions, selection pressures are expected to
57 change considerably, increasing the risk of a genetic adaptation to cultivation and a loss of
58 adaptations to the wild conditions over time (Ensslin et al., 2015; Husband and Campbell,
59 2004). In general, the combination of stochastic processes (genetic and demographic
60 changes due to small population sizes) and selection pressures (relaxed selection
61 pressures and unconscious selection by gardeners) may result in a reduction of genetic
62 variability (genetic erosion) and a differentiation away from their original wild population
63 (Ensslin et al., 2015).

64 Genetic variability of the plant's phenotype is a crucial feature of a population as it is what
65 natural selection acts upon (Falconer and Mackay, 1996; Vitt and Havens, 2004). Hence,
66 variation in phenotypic traits in a population can be directly connected to its adaptability to
67 cope with environmental changes (Lande and Shannon, 1996; Shaw and Etterson, 2012;
68 Vitt and Havens, 2004). In *ex situ* populations, these back-up populations are expected to
69 preserve all evolutionary relevant variability, and thereby the potential to evolve to
70 environmental changes. It has been shown in several studies that plants cultivated *ex situ*
71 can be genetically depauperate and may not represent the genetic diversity of their wild
72 populations (Brütting et al., 2013; Christe et al., 2014; Rucińska and Puchalski, 2011).
73 However, whether this also translates into a change in morphological traits or a reduction
74 of trait variability has not been studied. Neutral genetic markers used to infer genetic
75 diversity of a population (i.e. microsatellites, AFLPs or RAPDs) do not necessarily reflect
76 the variability in morphological characters and therefore may not be an appropriate
77 assessment of morphological genetic diversity (Reed and Frankham, 2001; Vitt and
78 Havens, 2004). Only very few studies so far investigated shifts in phenotypic traits
79 resulting from *ex situ* cultivation. Those studies found an increase in the germination rate,
80 advanced phenological timing and a change in plant architecture (Ensslin et al., 2018,
81 2015; Enßlin et al., 2011; Espeland et al., 2017; Rauschkolb et al., 2019). These studies
82 suggest that phenological (including germination) and life-history traits associated with
83 plant fitness might be most affected by evolutionary changes due to cultivation. However, it
84 has not been studied whether this could translate into reduced evolutionary potential or a
85 change in stress response. In recent years, the plant growth strategy has received
86 considerable interest due to its link to plant's responses to environmental change (Funk et
87 al., 2017). Specific leaf area (SLA) is an important trait for distinguishing fast-growing from
88 slow-growing individuals and can describe the effects of plant domestication (Matesanz
89 and Milla, 2018). Whether plants in *ex situ* collections show a shift towards a more fast-
90 growing strategy (Matesanz and Milla, 2018), has not been studied yet.

91 A population's evolutionary potential is composed of the genetic variation in quantitative
92 traits as well as the plants' response to stresses (i.e. phenotypic plasticity; Hoffmann &
93 Sgrò 2011; Franks, Weber & Aitken 2014). If plastic responses are costly, the selection will
94 remove them if they are not needed anymore (DeWitt et al., 1998). In the relaxed selection
95 of an ex situ environment, stresses such as lack of water and nutrient availability or
96 competition are typically reduced (Ensslin et al., 2015; Havens et al., 2004). Hence, there
97 is concern that the reduced environmental stresses may result in a loss of specific
98 adaptations to those stresses. As the magnitude of adaptation to a new environment
99 depends on the difference between the previous and the new habitat (Husband and
100 Campbell, 2004), species from more stressful habitats (e.g. dry and nutrient-poor
101 grasslands) may be more prone to genetic changes than species from mild or stable
102 habitats, which may manifest itself in a stronger change in their responses to stress when
103 cultivated.

104 In this study, we measured plant traits and their responses to drought stress in 12 plant
105 species that varied in years in cultivation from 1 to 30 years from the Meise Botanic
106 Garden, Belgium. We hypothesized that the longer ex situ cultivation will result in a change
107 in trait means, a reduction in trait variability and a reduced response to drought stress.
108 Specifically, we were interested in the following questions: 1. Do cultivated plants show a
109 shift in traits associated with phenology, fitness, performance and growth-strategy traits
110 compared to wild-collected plants? 2. Do cultivated plants show reduced responses to
111 drought compared to wild-collected plants, and is the change in response greater in
112 drought-tolerant than in drought-intolerant species? 3. Do cultivated plants have a reduced
113 genetic variability in traits compared to wild-collected plants?

114

115 **2. Methods**

116 **2.1 Species selection**

117 We conducted a multi-species common garden experiment with 12 herbaceous plant
118 species of the European flora (Table 1). All species are common species in Belgium,
119 meaning that they are not threatened and most of them widely distributed (van der Meijden
120 et al., 2016). Seeds of the 12 species were collected in 2015 from the garden beds of
121 Meise Botanic Garden and wild populations close to the original collection site. In the wild
122 populations, seeds were collected randomly from 7-10 individuals across the whole
123 population. Plants were sampled several meters apart to ensure that we collected from
124 separate genets. Garden populations were less extensive, so it was not possible to
125 maintain a distance of several meters between plants, but we were able to sample 8-10
126 individuals, unless the population was so small that all plants had to be sampled (e.g., in
127 *Helianthemum nummularium*, *Bromus erectus* and *Brachypodium pinnatum*, Table 1). The
128 cultivated species had been grown in the beds of the botanic garden between 1 and 30
129 years (Table 1). Due to logistical reasons and the disappearance of the some of the
130 original populations, the wild populations did not exactly match the populations originally
131 sampled for the ex situ collections. Most of the populations however, came from the same
132 bio-climatic region (within 1 and 53 km), except for three species, which originally came
133 from France and were substituted from a greater distance (300-400 km) from the wild
134 population. Because the species differ naturally in their sensitivity to drought, we used their
135 Ellenberg values (Ellenberg et al., 1992) to discriminate drought-tolerant (Ellenberg F ≤ 3)
136 from drought-intolerant (Ellenberg F > 3) species (Table 1).

137 **2.2 Germination and common garden experiment**

138 Seeds of all species were germinated in March 2016 in Petri dishes filled with 1% agar
139 (10g/l). Species that required cold stratification had been pre-placed in a fridge at 5°C for
140 two months (also on agar). Seeds of *Trifolium pratense* and *Helianthemum nummularium*
141 were scarified prior to germination with a scalpel. For the germination, we placed all petri
142 dishes into incubation chambers (LMS Cooled Incubator A280; LMS Ltd, Sevenoaks, UK)
143 and recorded germination twice a week. Germination treatments (temperature and day
144 length) for each species followed the recommendations of the Seed Information Database
145 (Royal Botanic Gardens Kew, 2017). In April 2016, we planted the seedlings into 0.8 litre
146 pots filled with a mixture of standard potting soil and sand (3:1), and placed in an open
147 tunnel greenhouse (4 m width, 2.2 m high, 21 m long) on the compound of Meise BG (see
148 Ensslin & Godefroid, 2019). Temperatures in the greenhouse ranged between 10 and 30
149 °C over the measurement period (recorded with a temperature logger in the greenhouse,
150 Model EL-USB 2, Dataq Instruments Inc, USA). We recorded the date of planting and the
151 size of each plant at the start as a covariate for the analyses. We randomly placed the
152 plants into four blocks, of which two blocks were treated with drought treatments in July
153 and August 2016. Blocks and individual positions within the blocks were randomized twice
154 each season. We measured the plants across two years in 2016 and 2017. In 2016, we
155 measured plant height, number of flowering stems, length of the longest leaf (hereafter
156 referred to as 'leaf length') and total aboveground biomass at the end of the first season. In
157 2017, we measured survival, number of flowering stems, leaf length, SLA and biomass at
158 the end of the second season. SLA was measured as the area of three freshly cut and
159 scanned leaves, divided by their dry weight. After scanning, leaves were dried at 29 °C for
160 three days in a drying oven and the weight measured with a precision scale. We used the
161 programme imageJ to determine the leaf area from scanned leaves (<https://imagej.net>).
162 Furthermore, we recorded the beginning of flowering period in every individual in both
163 years, and the duration of flower production (i.e. the time that open flowers were visible on
164 the plant) in 2017. We recorded flowering as a binary variable (yes or no) throughout the
165 experiment. Because only four of the 12 species flowered in the first year, we merged the
166 variables related to the flowering, i.e. flowering (yes/no), beginning of the flowering period
167 and number of flowering stems from both years, while all other variables were analysed
168 separately in the two years (Table 2). This means that flowering meant whether each
169 individual flowered within the duration of the whole experiment (two years), and the
170 beginning of the flowering period was counted from the day when each individual was
171 potted until it produced the first flower. The number of flowering stems was averaged
172 across both years if species flowered in both years.

173 **2.3 Drought treatment**

174 In summer 2016, we subjected half of the plants to two periods of intensive drought. The
175 first drought treatment started on the 12th of June and lasted until the 30th of June; while
176 the second drought treatment started on the 11th of July and lasted until the 20th of July. In
177 both drought treatments, plants subjected to drought were not watered until they showed
178 strong signs of suffering (wilting of leaves). This meant that within one treatment, we only
179 gave enough water to the suffering plants to prevent their death, but left the ones that did
180 not show signs of suffering untouched. The drought treatments were stopped after 10 and
181 19 days, respectively, in order not to risk the complete death of the more drought-sensitive
182 species. The control plants were watered either every day or two times per week
183 depending on the season to prevent a shortage of water availability. Drought stress
184 response was analysed as the difference in measured traits between plants under drought
185 compared to control plants. An increase in drought stress response means a greater
186 difference between drought and control plants.

187 **2.4 Quantitative genetic design**

188 We used a maternal half-sib design in our experiment meaning that all individuals had a
189 known mother (maternal lineage), but an unknown father. Genetic variation in quantitative
190 traits can be inferred from maternal lineage as genetically controlled traits are significantly
191 more similar when they are from individuals of the same maternal lineage, than when they
192 are taken randomly (Falconer and Mackay, 1996). In the analyses, we used the maternal
193 lineage as an indication of genetic variation in traits (Ensslin and Fischer, 2015; Falconer
194 and Mackay, 1996). We grew four individuals per maternal lineage, for each treatment and
195 species combination. The only exception was the wild origin seed of *Bromus erectus*,
196 which could not be collected separately by maternal lineage (Table 1) and therefore, did
197 not contribute to the maternal lineage effect. In this species, we planted 25 wild and 24
198 garden origin individuals per treatment. Due to mortality of young seedlings in some
199 species, this resulted in 1451 individuals for the experiment of the possible 1650 (Table A1
200 in the appendix). To further test genetic trait variability between garden-cultivated and wild
201 plants across all species, we calculated standardized coefficients of variation across
202 maternal lineages with all continuous traits and analysed those in separate mixed effect
203 models as described below (Table 3).

204 **2.5 Statistical analyses**

205 We used linear and generalized linear mixed models (Bates et al., 2015) to test whether
206 cultivated plants differed in morphological traits and their response to drought stress
207 compared to wild plants. To explore this, we fitted the full models including main effects
208 and interactions and determined significance by omitting one factor from the model and
209 comparing the two models with a likelihood ratio test. Significances of all predictors were
210 retained in the model and used for a subsequent FDR-Analysis (graphically sharpened
211 method) to account for multiple testing (Pike, 2011). Because our plant species had been
212 cultivated for very different periods in the botanic garden (Table 1), we used the time the
213 plants had spent in cultivation (in years) as continuous explanatory variable. Wild collected
214 plants were defined as having zero cultivation time. Moreover, we included species'
215 drought-tolerance (Ellenberg values) and drought treatment as explanatory variables. We
216 tested these variables and all their interactions on the continuous plant traits (plant
217 performance, phenology and growth strategy) and binary response traits (survival,
218 flowering). We included species identity, maternal lineage and block as random variables,
219 while all others as fixed. For the species identity effect, we accounted for a random
220 intercept and a random slope (Drought/Species) in our model assuming trait means as
221 well as their response to drought varied among species. We also tried to fit a random
222 slope for the maternal lineage, but these models did not converge. We therefore included
223 the random intercept (maternal lineage) in the final model. We included two covariates, the
224 starting plant size to account for the maternal environment and the geographic distance
225 between wild and original source population to account for the difference in wild and
226 original population identity (see section 2.1). We discuss effects only across all 12 species
227 in order to minimize a potential confounding effect of cultivation time and population
228 identity. Some variables were log- or square root transformed to meet assumptions of
229 normality and heteroscedasticity (Table 2).

230

231 We used the effects package (Fox, 2003) to display regression lines and confidence
232 intervals (CI). Partial residuals were fitted with the remef package (Hohenstein and Kliegl,
233 2020). As CIs made by the effects package may not always be reliable, we calculated
234 credible intervals (a Bayesian analogue of confidence intervals) for a subsample of the
235 tests with the sim function of the arm package (Gelman and Yu-Sung, 2018). As the
236 credible intervals were almost indistinguishable from the CIs, we decided to keep the CIs
237 for simplicity. All analyses were performed with R version 3.4.2 (R Core Team, 2020).

239 3. Results

240 3.1 Trait means and drought stress response

241 We found significant effects of drought-tolerance, drought treatment and cultivation and
242 their interactions in all response variables except for the duration of flowering and the
243 biomass in the second year (Table 2).

244 Drought treatment and drought-tolerance: The drought treatment affected all species
245 negatively in three performance traits (survival, plant height, number of flowering stems)
246 and reduced flowering (Survival: $\chi^2=4.5$, $P=0.046$; height: $\chi^2=5.1$, $P=0.036$; flower stems:
247 $\chi^2=6.4$, $P=0.01$; Fig. A1a-d, Appendix). However, drought-tolerant species responded
248 much less to the drought treatment in terms of biomass and flowering than drought-
249 intolerant species (Drought x drought-tolerant interaction, biomass: $\chi^2=5.4$, $P=0.032$;
250 flowering: $\chi^2=5.5$, $P=0.034$; Fig A2a-b, Appendix). One year after the drought, drought-
251 intolerant species, which had experienced drought, re-grew stronger (leaf length), while
252 drought-tolerant species produced slightly smaller leaves than their control plants ($\chi^2=4.7$,
253 $P=0.034$; Fig A2c, Appendix).

254 Ex situ cultivation: Cultivation time affected the species' trait means in eight out of 12
255 cases, but this often depended on the other factors (Table 2). Germination rate as well as
256 leaf length increased with increasing cultivation time in the garden, while flowering was
257 reduced (Germination: $\chi^2=26.8$, $P<0.001$, leaf length: $\chi^2=8.5$, $P=0.009$, flowering: $\chi^2=8.3$,
258 $P=0.01$ Fig. 1a-c). While for drought-tolerant species, cultivation in the garden delayed the
259 beginning of the flowering period, reduced the number of flowering stems and slightly
260 increased SLA, this was not the case for drought-intolerant species, where the flowering
261 time and SLA remained unaffected and stem number even increased (Drought-tolerance x
262 cultivation time interaction, flowering start: $\chi^2=24.5$, $P<0.001$, flowering stems: $\chi^2=8.5$,
263 $P=0.009$, SLA: $\chi^2=5.1$, $P=0.034$; Fig. 2a-c). The effect of cultivation time also depended on
264 the drought treatment of the plants (Table 2). Here, cultivation resulted in a higher mortality
265 under drought (Drought x cultivation time interaction, $\chi^2=4.3$, $P=0.049$; Fig. 2d). However,
266 plants that were cultivated for longer times slightly decreased in leaf length and in SLA
267 when they had experienced drought, while this was not the case for the control plants
268 (Leaf length: $\chi^2=5.9$, $P=0.034$, SLA: $\chi^2=4.7$, $P=0.042$; Fig. 2e-f).

269 For the plant height, drought-intolerant species responded less to drought the longer they
270 were cultivated, while it was opposite for drought-tolerant species (Fig. A3, Appendix).
271 However, only one drought-tolerant species (*Origanum vulgare*) produced measurable
272 tillers in the first year (only rosettes for all other drought-tolerant species), so for the
273 drought-tolerant species, this result may also be a population identity effect (see section
274 2.5).

275 3.2 Trait variability

276 We found genetic variation in quantitative traits in 10 out of 12 traits studied (Maternal
277 lineage effect; Table 2). We found an effect of cultivation time on trait variability in three
278 out of the 12 analysed traits. In all these traits, i.e. the germination rate, the leaf length and
279 the aboveground biomass in the first year, the coefficient of variation decreased with
280 increasing cultivation time (Germination: $\chi^2=8.6$, $P=0.022$, leaf length: $\chi^2=11.3$, $P=0.006$,
281 biomass: $\chi^2=10.4$, $P=0.001$; Fig. 1d-f). The variability in leaf length increased also with the
282 drought treatment, but only for the drought-intolerant species (Drought-tolerance x drought
283 interaction, $\chi^2=6.8$, $P=0.043$; Fig A4, Appendix).

285 4. Discussion

286 Cultivation of plants in botanic gardens may influence the nature and variability of traits via
287 stochastic effects and selection during the cultivation process, including loss of selective
288 advantages to environmental stressors. We found indications that not only traits had
289 changed during cultivation, but also trait variability and the response of the plants to
290 drought stress, i.e. phenotypic plasticity.

291 4.1 Changes in trait means and drought stress response

292 Altered environmental conditions and management practices have been considered as key
293 drivers of a rapid evolutionary response in cultivated species (Ensslin et al., 2015;
294 Husband and Campbell, 2004). Known as the domestication syndrome (Hammer, 1984),
295 this involves genetic changes in trait means towards an increased plant vigour, the
296 homogenisation of traits (reduction of trait variability) and the reduction of costly
297 adaptations to natural abiotic stressors (Milla et al., 2015). While those changes directly or
298 indirectly contributed to the creation of crop species and hence to the aim of better making
299 use of the plants, they would mostly be unwanted and potentially detrimental in the context
300 of conservation (Havens *et al.* 2004; Ensslin & Godefroid 2019, but see Chivers et al.,
301 2016).

302 Potential changes due to the cultivation of wild plants have received increasing interest in
303 recent years (Basey et al., 2015; Ensslin et al., 2015; Espeland et al., 2017), but in-depth
304 studies are still rare. In our study of 12 common herbaceous species from a Belgian
305 botanic garden, we found indications that garden-cultivation resulted in an increase in
306 germination rate, a decrease in flowering probability, and for drought-tolerant species also
307 a delay in flowering time. The increase of germination rate in botanic garden plants has
308 been shown and discussed in several studies before (Ensslin et al., 2018; Enßlin et al.,
309 2011; Rauschkolb et al., 2019) and has also been found in commercially produced seeds
310 for grassland restoration and medicinal plants (Qu et al., 2005; Schröder and Prasse,
311 2013).

312 The decrease in flowering probability and the delay in flowering observed in our study are
313 at first glimpse puzzling as we could have expected acceleration of phenology and life
314 cycle together with the increase in germination rate (Donohue, 2002). However, later
315 flowering of botanic garden cultivated plants has also been found by Rauschkolb *et al.*,
316 (2019), who explained it with a potential relaxation of the flowering threshold due to
317 missing competition under garden conditions. Flowering threshold sizes are highly
318 heritable and can evolve as a consequence of habitat stability (Wesselingh et al., 1997).
319 As all species of our experiment were cultivated in ordinary cultivation beds with no
320 competition and good nutrient provision, an evolution of increased threshold size in the
321 garden collections seems a possible explanation for the observed pattern. However, a
322 study with seed of the same origin as the longest cultivated species in our experiment
323 (*Digitalis lutea*), showed an earlier flowering in the garden plants (T. Sandner, unpublished
324 data). This indicates that selection on earlier or later flowering may depend on the
325 cultivation, seed collection and regeneration method of the species in the garden (Ensslin
326 and Godefroid, 2019). In general, our study shows that phenology is a sensitive trait, which
327 may undergo strong and rapid shifts in ex situ collections. Interestingly, ex situ cultivation
328 did not affect aboveground biomass production in our experiment. Similarly, Ensslin et al.,
329 (2015) and Rauschkolb et al., (2019) did not find significant differences in biomass
330 production between ex situ and wild plants, while they did, like us, find changes in traits
331 referring to phenology and plant architecture (e.g. number of flowering stems).

332 Aboveground biomass is a rather broad trait that encompasses productivity and
333 summarizes many fine-tuned traits of plant architecture, life cycle strategy and resource
334 use. We assume that changes due to ex situ cultivation may happen more quickly in
335 narrow-defined traits (because of direct selection on them) and may take more time in
336 more broad traits such as biomass.

337 Changes in trait means varied by drought-tolerance of the species. Delay in the beginning
338 of the flowering period and the increase in SLA in drought-tolerant species point to a
339 change in the growth strategy towards increased resource acquisition (as shown for an
340 increase in SLA, see e.g Milla and Matesanz, 2017) and increase in threshold size of the
341 flowering rosette (see discussion above). A fertilization treatment would give more insights
342 here. An experiment with a dry grassland plant in Germany, Enßlin *et al.*, (2011) found that
343 garden-cultivated plants responded to fertilization by producing one big flowering stem
344 rather than the several smaller ones found in wild plants. Adaptive changes in species due
345 to cultivation are expected to increase with the relative difference between the new ex situ
346 environment and the old natural environment (Husband and Campbell, 2004). As we
347 assume that the change in environment when growing the plants in the garden beds was
348 greater in drought-tolerant than in drought-intolerant species, it might explain their greater
349 changes in leaf and phenological traits during cultivation. The number of flowering stems
350 slightly decreased in drought-tolerant species during cultivation (as also found in other
351 studies, see Ensslin *et al.*, 2015; Enßlin *et al.*, 2011), but increased in drought-intolerant
352 species. As there were only six species in each group and flowering probability was not
353 very high in some species, longer studies with more habitat generalists and specialists
354 could give further insights into this matter.

355 We found that the effect of cultivation on plant traits was often dependent on the drought
356 treatment. Here, cultivation resulted in both an increase and decrease of response to
357 drought, depending on the trait in question. For instance, there was an increase in
358 mortality when cultivated plants experienced drought, but those who survived also showed
359 a reduced response performance-related traits (plant height, leaf length, SLA). As the
360 plants in the botanic garden beds are usually watered when drought stress occurs, they do
361 not experience strong selection as in nature. If drought responses are costly, e.g.
362 producing stress-resistant enzymes or proteins, or by altering the architecture of the plant
363 (DeWitt *et al.*, 1998; van Kleunen and Fischer, 2007), then those responses could be
364 selected against in the optimized conditions of a garden. Moreover, as most of the
365 populations in the gardens are very small, stress responses can be lost by demographic
366 stochasticity and random genetic drift (Ellstrand and Elam, 1993), or changed by
367 inbreeding (Sandner and Matthies, 2018). We interpret our results that garden-cultivated
368 plants may have reduced some of their natural responses to drought, such as to reduce
369 transpiration by growing smaller and producing smaller leaves. The increased mortality of
370 cultivated plants under drought may have negative consequences if these plants were
371 used in reintroductions.

372 We note that the drought treatment may have been perceived differently by drought-
373 tolerant and intolerant species, as shown by the stronger biomass reduction in drought-
374 intolerant species with drought (Fig. A2a, Appendix). However, this potential difference in
375 stress perception did not influence the cultivation effects on plant response to drought, as
376 shown by the non-significant three-way interactions of drought-tolerance, drought
377 treatment and cultivation (Table 2, see section 3.1 for the plant height effect). While we are
378 confident that our results show real changes in the response to drought stress during
379 cultivation, the effect sizes (i.e. magnitude of the change in response) were rather small.
380 Field transplantation studies with ex situ cultivated material would be needed to test the
381 consequences of those changes under natural conditions.

383 4.2 Genetic variation in traits during cultivation

384 The domestication syndrome predicts a reduction in variability, particularly in phenological
385 traits, as a consequence of selection by agricultural growth and harvesting methods
386 (Meyer et al., 2012). Similar predictions can be made from stochastic processes, which
387 may also result in a reduction in phenotypic trait variability and hence, evolutionary
388 potential (Vitt and Havens, 2004). In our analyses with 12 species, we found a reduction in
389 trait variability in three performance traits (germination rate, biomass, leaf length). Genetic
390 variation in traits is considered important for adaptation to rapid changes in environmental
391 and biotic conditions (Jump et al., 2009). The reduction of variability could compromise the
392 ability of cultivated species to adapt to rapid changes in their natural habitat. To our
393 knowledge, this is the first study that demonstrates a reduction of evolutionary potential, a
394 real concern in ex situ-cultivated plants.

395 We were surprised to find only a weak impact of the drought treatment on the trait
396 variability as it may be expected to increase under stress (Stanton et al., 2000) and hence,
397 a reduction of variability could have been more strongly visible in drought-stressed plants.
398 While we found this for one trait (leaf length first year), it was only true for drought-
399 intolerant species and did not change with cultivation time. As the drought-tolerant species
400 were potentially less stressed by our drought treatment (see discussion above in 4.1),
401 further studies with more and different kinds of stresses are needed to reveal how trait
402 variability and stress interact in ex situ collections.

403 We acknowledge that the species in our experiment were common plants in Belgium and
404 were not collected with a conservation purpose for cultivation in the botanic garden.
405 Therefore, those collections were not sampled with a particular protocol and may not have
406 been treated with special care in the garden to prevent genetic changes. Moreover, we do
407 not know the number of generations that the plants have passed in the ex situ environment
408 and as our plants were all perennials, there is a possibility that some individuals still
409 represent the first generation planted in the garden (especially the shortly cultivated
410 species). We hence cannot clearly say to which degree adaptive evolution or genetic drift
411 by founder effects were responsible for the observed effects. Also maternal effects might
412 have influenced our results as we used the F1 generation for our study, especially in the
413 germination where we did not account for it. However, several other studies showed that
414 ex situ cultivation can result in genetic changes in phenotypic traits within only very few
415 generations (Ensslin et al., 2015; Espeland et al., 2017; Nagel et al., 2019; Rauschkolb et
416 al., 2019). We therefore believe that the patterns observed in our study reflect well the
417 risks that are threatening the genetic and phenotypic representation of conservation
418 collections of endangered plants.

419

420 4.3 Conclusions

421 We conclude that changes in traits, a reduction of trait variability and drought stress
422 response are realistic concerns in ex situ collections, especially when species are
423 cultivated over longer time scales. Our results moreover indicate that drought-tolerant
424 species respond differently to cultivation than drought-intolerant ones, with some hints to
425 an increased resource acquisition strategy in drought-tolerant species. Hence, we
426 reinforce our call to cultivate conservation-dedicated collections with a long-term
427 perspective (in contrast to a short-term multiplication for direct reintroduction measures) in

428 near-to-nature habitat rather than under optimal growth conditions (Ensslin and Godefroid,
429 2019).

430 We found a reduced trait variability and a slight reduction of response to drought stress in
431 cultivated plant species, indicating a reduced evolutionary potential in the ex situ
432 collections. This could compromise their ability to survive and evolve under natural
433 conditions, particularly with current rapid change of climate and land use. We therefore
434 strongly recommend that the cultivation of wild plant species follows the latest protocols to
435 avoid the above-mentioned risks. While those protocols are readily available (Basey et al.,
436 2015; Ensslin and Godefroid, 2019; Maschinski and Albrecht, 2017), they still need to be
437 tested for their effectiveness. A number of botanic gardens are already offering specific
438 trainings in how to support rare and endangered species (Ensslin and Godefroid, 2019;
439 Sharrock and Chavez, 2013). We suggest that all institutions dealing with ex situ
440 conservation have their gardeners follow a training in conservation horticulture that
441 addresses the implications of genetic and evolutionary processes on their collections.

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608 Table 1 : Overview of the species used in the experiment, the plant family they belong to, the
 609 number of mother plants from which the seeds were collected (ML = Maternal lineages; G/W
 610 stands for garden and wild origin), and the time (years) the garden plants had been cultivated at
 611 Meise Botanic Garden. Drought-tol = drought-tolerant (Yes when $F \leq 3$, No when $F > 3$, according
 612 to Ellenberg et al., 1992).

Species	Family	ML (G/W)	Cult. time	Drought-tol
<i>Brachypodium pinnatum</i>	Poaceae	6/10	17	Yes
<i>Bromus erectus</i>	Poaceae	4/1	15	Yes
<i>Dianthus carthusianorum</i>	Caryophyllaceae	8/8	1	Yes
<i>Digitalis lutea</i>	Plantaginaceae	10/9	30	No
<i>Geum urbanum</i>	Rosaceae	10/10	19	No
<i>Globularia bisnagarica</i>	Plantaginaceae	10/10	10	Yes
<i>Helianthemum nummularium</i>	Cistaceae	3/7	20	Yes
<i>Linaria vulgaris</i>	Plantaginaceae	8/10	10	No
<i>Origanum vulgare</i>	Lamiaceae	9/10	20	Yes
<i>Pulicaria dysenterica</i>	Asteraceae	10/8	15	No
<i>Teucrium scorodonia</i>	Lamiaceae	10/10	20	No
<i>Trifolium pratense</i>	Fabaceae	10/10	1	No

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615 Table 2 : Overview of the glmm and lmm models testing the effects of drought-tolerant and drought-intolerant species type (Drought-tol), drought
616 treatment (Drought), cultivation time (Cult. Time) and all their interactions on 12 performance and growth variables in a multi-species common
617 garden experiment. The variables: geographic distance (Geogr dist) and size of the plants at the start of the experiment (Startsize) were treated as
618 co-variables. Transformation indicates when a transformation of a dependent variable was necessary to fit model assumptions. Dependent
619 variables : Germination rate (Germ), beginning of the flowering period (FI start), leaf length (Leaf l), plant height (Height), aboveground biomass
620 (Biom), survival, flowering, flowering duration (FI dur), specific leaf area (SLA), number of flowering stems (FI stems). / = Random effects. N=
621 sample size for each trait measurement. Significant effects are bold faced. P-values were corrected for false discovery rates using the method
622 proposed by Pike (2011).

Year of measurement	2016	both	2016	2016	2016	both	both	2017	2017	2017	both	2017
Trait	Germ	FI start	Leaf l	Height	Biom	Survival	Flowering	FI dur	Leaf l	SLA	FI stems	Biom
Transformation	-	-	sqrt	-	sqrt	glmer	glmer	sqrt	log+1	log+1	sqrt	sqrt
Sample size (N)	410	591	1429	634	1350	1447	1447	328	1205	1200	617	1176
Geogr dist	0.111	0.412	0.465	0.034	0.272	0.166	0.388	0.085	0.141	0.290	0.037	0.044
Startsize	-	0.037	0.000	0.001	0.000	0.430	0.002	0.096	0.006	0.430	0.128	0.000
Drought-tol	0.045	0.020	0.436	0.034	0.027	0.360	0.037	0.456	0.228	0.017	0.465	0.070
Drought	-	0.238	0.194	0.036	0.096	0.046	0.030	0.432	0.431	0.046	0.011	0.178
Cult. time	0.000	0.000	0.009	0.440	0.510	0.135	0.010	0.465	0.225	0.456	0.128	0.298
Drought x Cult. time	-	0.111	0.238	0.418	0.418	0.049	0.349	0.430	0.034	0.042	0.096	0.426
Drought-tol x Drought	-	0.272	0.070	0.415	0.034	0.054	0.034	0.436	0.034	0.456	0.085	0.070
Drought-tol x Cult. time	0.096	0.000	0.418	0.298	0.298	0.381	0.363	0.190	0.272	0.034	0.009	0.314
Drought-tol x Drought x Cult. time	-	0.125	0.358	0.013	0.326	0.418	0.230	0.430	0.194	0.430	0.456	0.266
1/Block	-	-	0.090	0.076	0.076	0.272	0.456	-	0.019	0.096	0.497	0.036
1/Species	0.000	-	-	-	-	-	0.000	0.000	-	-	0.000	-
1/Maternal lineage	0.000	0.000	0.000	0.001	0.000	0.418	0.000	0.272	0.000	0.000	0.000	0.000
Drought/Species		0.128	0.000	0.000	0.000	0.000	-	-	0.034	0.020	-	0.042

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628 Table 3 : Overview of the glmm and lmm models testing the effects of drought-tolerant and drought-intolerant species type (Drought-tol), drought
629 treatment (Drought), cultivation time (Cult. Time) and all their interactions on the coefficient of variation (CV) of 12 performance and growth
630 variables in a multi-species common garden experiment. The variables: geographic distance (Geogr dist) and size of the plants at the start of the
631 experiment (Startsize) were treated as co-variables. Transformation indicates when a transformation of a dependent variable was necessary to fit
632 model assumptions. Dependent variables: Germination rate (Germination), beginning of the flowering period (FI start), leaf length (Leaf l), plant
633 height (Height), aboveground biomass (Biom), flowering duration (FI dur), specific leaf area (SLA), number of flowering stems (FI stems). / =
634 Random effects. N= sample size for each trait measurement. Significant effects are bold faced. P-values were corrected for false discovery rates
635 using the method proposed by Pike (2011).

Year of measurement	2016	both	2016	2016	2016	2017	2017	2017	both	2017
CV of Trait	Germination	FI start	Leaf l	Height	Biom	FI dur	Leaf l	SLA	FI stems	Biom
Transformation	sqrt	sqrt	sqrt	sqrt	sqrt	-	-	-	-	-
Sample size (N)	193	160	363	183	356	84	321	321	166	322
Geogr Dist	0.427	0.200	0.728	0.039	0.783	0.092	0.427	0.448	0.223	0.162
Drought-tol	0.525	0.723	0.724	0.092	0.385	0.723	0.728	0.728	0.110	0.456
Drought	-	0.728	0.299	0.000	0.724	0.581	0.728	0.581	0.783	0.728
Cult. time	0.022	0.728	0.006	0.385	0.001	0.427	0.724	0.783	0.623	0.378
Drought x Cult. time	-	0.514	0.092	0.162	0.728	0.579	0.724	0.630	0.724	0.630
Drought-tol x Drought	-	0.724	0.043	0.783	0.581	0.728	0.484	0.728	0.728	0.092
Drought-tol x Cult. time	0.783	0.783	0.514	0.120	0.448	0.756	0.120	0.385	0.378	0.724
Drought-tol x Drought x Cult. time	-	0.350	0.783	0.043	0.728	0.728	0.378	0.514	0.484	0.427
1/Species	0.000	0.000	0.000	0.728	0.000	0.783	0.023	0.000	0.000	0.043

636 Figure legends

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638 Figure 1 : Influence of cultivation time (years) on trait means (a-c) and trait variability (d-f) of 12 ex-
639 situ cultivated and wild-collected species in a multi-species experiment at Meise Botanic Garden,
640 Belgium. (a) Germination rate ($P < 0.001$), (b) leaf length of the first year ($P = 0.009$) and (c) flowering
641 ($P = 0.01$). Trait variability was measured as the coefficient of variation (CV, see section 2.5). (d) CV
642 of germination rate ($P = 0.022$), (e) CV of leaf length of the first year ($P = 0.006$) and (f) CV of the
643 aboveground biomass in the first year ($P = 0.001$). The graphs show regression lines (solid), partial
644 residuals (points, except c) and confidence intervals (dotted lines; see section 2.5).

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646 Figure 2 : Interaction plots of the effects of cultivation time (years) and drought tolerant (blue)
647 versus drought intolerant (orange) species (a-c), as well as drought stressed (brown) and control
648 (green) plants (d-f) on survival, phenology and performance of 12 ex-situ-cultivated and wild-
649 collected species in a multi-species experiment at Meise Botanic Garden, Belgium. (a) Day of first
650 flowering ($P < 0.001$), (b) number of flowering stems ($P = 0.009$) and (c) specific leaf area ($P = 0.034$).
651 (d) Survival (0.049), (e) leaf length of the 2nd year ($P = 0.034$) and (f) specific leaf area ($P = 0.043$).
652 The graphs show regression lines (solid), partial residuals (points, except d) and confidence
653 intervals (dotted lines; see section 2.5). The specific leaf area (c, f) is plotted on a log scale to
654 increase readability of the graph due to high spread of the data.

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