

Improving ecological insights from dendroecological studies of Arctic shrub dynamics: research gaps and potential solutions

Candice C. Power^{1,2*}, Jakob J. Assmann^{1,2}, Angela L. Prendin^{1,2}, Urs A. Treier^{1,2}, Jeffrey T. Kerby^{2,3}, Signe Normand^{1,2*}

¹Department of Biology, Aarhus University, Ny Munkegade 114-116, DK-8000, Aarhus C, Denmark

²Center for Biodiversity Dynamics in a Changing world, Aarhus University, Ny Munkegade 114-116, DK-8000, Aarhus C, Denmark

³Aarhus Institute of Advanced Studies, Aarhus University, Høegh-Guldbergs Gade 6B, DK-8000, Aarhus C, Denmark

*Corresponding authors, shared

ORCIDiDs:

Candice C. Power 0000-0002-9677-2587

Jakob J. Assmann 0000-0002-3492-8419

Angela L. Prendin 0000-0002-5809-7314

Urs A. Treier 0000-0003-4027-739X

Jeffrey T. Kerby 0000-0002-2739-9096

Signe Normand 0000-0002-8782-4154

FUNDING

CCP, JJA, UAT, and SN have been supported by the Independent Research Fund of Denmark (grant 7027-00133B to SN). ALP has been supported by a Marie Skłodowska-Curie Individual fellowship (IF) under contract number 895233. JTK has been supported by the Aarhus University Research Foundation and the European Union's Horizon 2020 Research and Innovation Programme under Marie Skłodowska-Curie Grant Agreement 754513.

DECLARATIONS

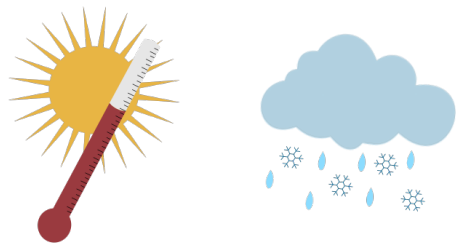
The authors declare no conflicts of interests.

ACKNOWLEDGEMENTS

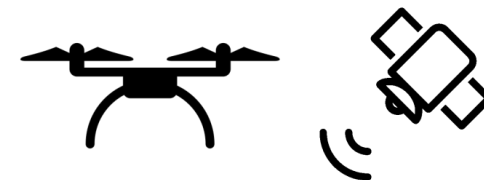
We acknowledge the analysis and graphical representation of Figure 1 in Babst et al. (2017) that inspired our Figure 3 and 4. The drone icon in the graphical abstract is by gzz from the Noun Project. We thank Liv Normand-Treier for the shrub illustrations and Jonas Ravn Jensen for the shrub photographs in Figure 5. We thank Ashley Percy-Buitenwerf for helpful comments on the manuscript and three anonymous reviewers for their feedback on an earlier version of this manuscript.

More comprehensive understanding of Arctic shrub dynamics

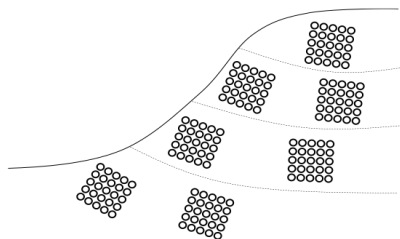
Investigation of more potential drivers



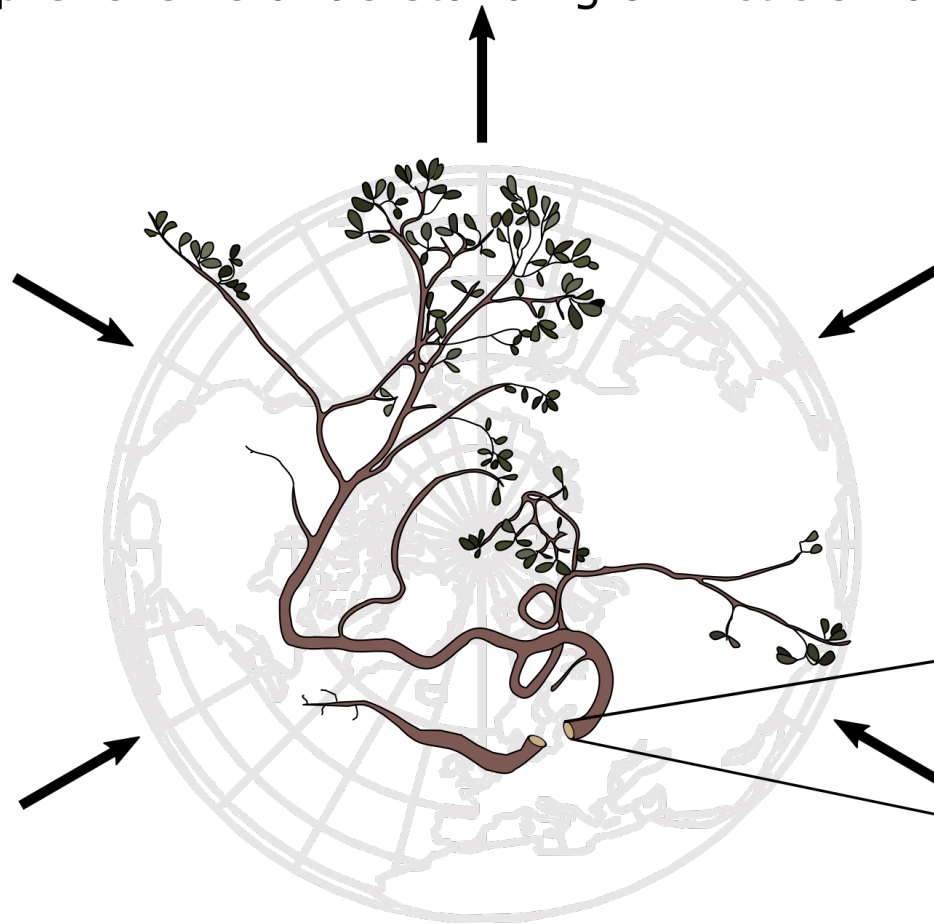
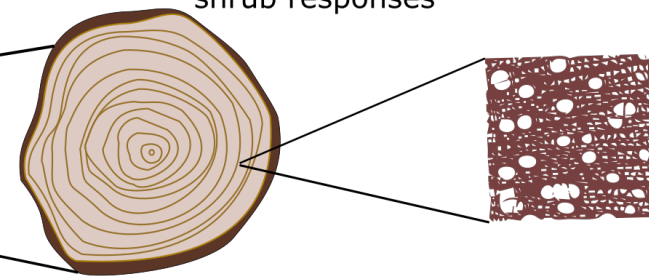
Multidisciplinary Approaches



Systematic sampling methods



Incorporation of more shrub responses



1 **Improving ecological insights from dendroecological studies of**

2 **Arctic shrub dynamics: research gaps and potential solutions**

3 **Research Highlights**

- 4 1. We provide a perspective on how to improve ecological insights using dendroecology
- 5 2. We highlight research gaps based on recent Arctic shrub dendroecological studies
- 6 3. More attention should be given to inter- and intra- specific demographic responses
- 7 4. More drivers and responses should be assessed at relevant spatial and temporal scales
- 8 5. An interdisciplinary approach is suggested to improve cross-scale ecological insights

9 **ABSTRACT**

10 Rapid climate change has been driving changes in Arctic vegetation in recent decades, with
11 increased shrub dominance in many tundra ecosystems. Dendroecological observations of
12 tundra shrubs can provide insight into current and past growth and recruitment patterns, both
13 key components for understanding and predicting ongoing and future Arctic shrub dynamics.
14 However, generalizing these dynamics is challenging as they are highly scale-dependent and
15 vary among sites, species, and individuals. Here, we provide a perspective on how some of
16 these challenges can be overcome. Based on a targeted literature search of dendroecological
17 studies from 2005-2022, we highlight five research gaps that currently limit dendroecological
18 studies from revealing cross-scale ecological insight into shrub dynamics across the Arctic
19 biome. We further discuss the related research priorities, suggesting that future studies could
20 consider: 1) increasing focus on intra- and interspecific variation, 2) including demographic
21 responses other than radial growth, 3) incorporating drivers, in addition to warming, at
22 different spatial and temporal scales, 4) implementing systematic and unbiased sampling
23 approaches, and 5) investigating the cellular mechanisms behind the observed responses.

24 Focusing on these aspects in dendroecological studies could improve the value of the field for
25 addressing cross-scale and plant community-framed ecological questions. We outline how
26 this could be facilitated through the integration of community-based dendroecology and
27 dendroanatomy with remote sensing approaches. Integrating new technologies and a more
28 multidisciplinary approach in dendroecological research could provide key opportunities to
29 close important knowledge gaps in our understanding of scale-dependencies, as well as intra-
30 and inter-specific variation, in vegetation community dynamics across the Arctic tundra.

31 **Keywords:** demographic responses, shrub ring-width, spatial-temporal scales, earth
32 observations, sampling approach, plant community-based dendroecological sampling

33

34 **1. INTRODUCTION**

35 Rising temperatures have been associated with increases in plant productivity across the
36 Arctic over recent decades (Bjorkman et al., 2020; IPCC Working Group II, 2022). Shrubs, a
37 dominant feature of Arctic tundra, have increased in height, biomass, cover, and abundance
38 in many areas of the biome, a process referred to as “shrubification” (Martin et al., 2017;
39 Myers-Smith et al., 2011; Tape et al., 2006). These shrub responses to warming have
40 potentially wide-reaching impacts such as altering herbivory interactions (Le Moullec et al.,
41 2020; Post et al., 2021; Skarin et al., 2020), albedo (Belke-Brea et al., 2020; Blok et al., 2011;
42 Bonfils et al., 2012; Loranty et al., 2011), soil temperatures (Lawrence and Swenson, 2011),
43 and land-atmosphere exchanges of carbon (CO₂ and CH₄) (Bonfils et al., 2012; Mekonnen et
44 al., 2021). Thus, changes in shrub communities can contribute to large-scale climatic
45 feedbacks (Pearson et al., 2013), although the net effect of the combined feedbacks remains
46 uncertain (AMAP, 2021, 2017).

47 Despite many areas showing increases in shrub productivity, these changes are occurring
48 heterogeneously throughout the landscape (Berner et al., 2020; Gamm et al., 2018; Guay et
49 al., 2014), with evidence of substantial differences in local and regional scale trends (Bhatt et
50 al., 2013; Gamm et al., 2018; Myers-Smith et al., 2015a; Phoenix and Bjerke, 2016; Post et
51 al., 2021; Reichle et al., 2018). Currently, the mechanisms driving these different vegetation
52 responses are not fully understood. In particular, we lack knowledge on how the responses of
53 individuals scale to the responses of communities and vary across species' geographic ranges
54 (Martin et al., 2017) (Fig. 1a). This contributes uncertainty to predictions of the magnitude
55 and extent of ongoing local changes and their implications for climate feedbacks, as well as
56 the structure and function of Arctic ecosystems (Kemppinen et al., 2021; Myers-Smith et al.,
57 2011; Vowles and Björk, 2019). Therefore, to better understand and predict future global
58 climate and ecosystem changes, it is necessary to understand the complex dynamics and
59 interactions within and among Arctic shrubs and their responses to non-climatic and climatic
60 factors from local to broad scales. This requires quantifying responses and their drivers across
61 temporal and spatial scales, including where along the plant structure or life stage the driver
62 acts, in which season it acts, and if the effect is immediate or gradual (Körner and
63 Hiltbrunner, 2018). Advancements in technology and techniques across a spectrum of
64 disciplines (e.g., remote sensing, dendroanatomy, dendrometers, and phenocams) now allow
65 for investigations ranging from quantifying very fine cellular responses to broad scale
66 patterns in both time and space (Anderson and Gaston, 2013; Dobbert et al., 2021a; Myers-
67 Smith et al., 2020; Parmentier et al., 2021; Prendin et al., 2020).

68 Dendroecology is a branch of dendrochronology that examines annual and sub-annual growth
69 patterns of woody plants in relation to ecological drivers (Fritts, 1976, 1971; Fritts and
70 Swetnam, 1989; Schweingruber, 1996). Although long recognized for having
71 dendroecological potential (Johnstone and Henry, 1997; Wilson, 1964; Woodcock and

72 Bradley, 1994), the widespread adoption of dendroecology to study dwarf shrubs growing
73 above the tree line has been relatively recent. Notably, dendroecology, has been successfully
74 applied to understand long-term temporal variation in shrub growth (Bär et al., 2008, 2006;
75 Rayback and Henry, 2005; Schweingruber and Poschlod, 2005) and recruitment (Büntgen et
76 al., 2015; Myers-Smith and Hik, 2018). Additionally, dendroanatomy has been successfully
77 used to gain knowledge on structure–function responses of plants to climate and
78 environmental variability (Fonti et al., 2010; Schweingruber et al., 2013). This discipline
79 focuses on the quantitative assessment of wood anatomical traits (e.g., lumen dimensions and
80 wall thickness of conducting cells, fibers, and several ray properties (von Arx et al., 2016)),
81 and their link to the specific xylem functions (e.g., water and nutrient transport from the
82 plant–soil interface to stems and leaves, mechanical support and storage (Myburg et al.,
83 2013)), using a dendrochronological approach (Carrer et al., 2016; von Arx et al., 2016).
84 Studies combining multiple approaches, e.g., ring-width measurements with time series of
85 anatomical traits, have the potential to provide insight into current and past shrub dynamics
86 and mechanistic responses (Prendin et al., 2020; Wilmking et al., 2018), and to inform
87 predictions about future changes (Babst et al., 2018; McMahon et al., 2011). These
88 techniques can provide information on an intra-annual to annual time scale. However, as
89 these studies typically provide data from discrete geographic locations, the findings can be
90 difficult to up-scale and generalize (Babst et al., 2018).

91 Geospatial analyses and remote sensing data on abiotic and biotic variables have become
92 more accessible to ecologists in recent years due to improvements in sensor design and
93 deployment, a push towards open access, cloud-computing, storage and software, and the
94 availability of ready-to-use data products (Kwok, 2018). These emerging geospatial
95 techniques and technologies could supplement and strengthen dendroecological studies to
96 better link individual and plot-based observations with landscape scale patterns (Fig. 1a). For

97 example, dendroanatomical (Prendin et al., 2020) and tree-ring (Babst et al., 2010) analyses
98 can be combined with satellite-based remote sensing to reveal insights into impacts of insect
99 outbreaks at scales not possible with *in-situ* observations alone. Such multi-proxy approaches
100 have the potential to provide cross-scale ecological insight into how climate and
101 environmental drivers affect plant growth and xylem structure at different temporal (i.e.,
102 seasonal, annual, intra-annual) and spatial (e.g., local, regional) scales (Fig. 1b), but they
103 have not yet been widely adopted. Although possible with current technology, integrating
104 these techniques requires new perspectives in dendroecological studies (Manzanedo and
105 Pederson, 2019; Pearl et al., 2020) with increased attention on study design that allows for
106 scaling responses from individuals to communities and ecosystems (Babst et al., 2018).

107 To evaluate the prevalence and potential for cross-scale perspectives in Arctic
108 dendroecological research, we conducted a targeted literature review to highlight key gaps in
109 contemporary Arctic dendroecological approaches that may limit cross-scale generalization.
110 Building on these gaps, we outlined potential research priorities to improve cross-scale
111 ecological insights of Arctic shrub dynamics. Specifically, we: (1) reviewed current
112 methodologies and approaches; (2) highlighted research gaps and discussed the related
113 priorities; and (3) outlined potential ways forward for integrating plant community
114 perspectives and spatio-temporal approaches spanning from dendroanatomy to geospatial
115 analyses and remote sensing. The highlighted research gaps and suggested ways forward are
116 defined with respect to how dendroecological studies could contribute to the understanding of
117 cross-scale Arctic plant community dynamics, and do not reflect any criticism of the
118 individual studies included in the present review.

119

120 **2. A TARGETED LITERATURE REVIEW**

121 **2.1 METHODS**

122 We conducted a targeted literature review of studies using dendroecological methods and
123 extracted spatial and temporal information on the investigated samples, responses, and
124 drivers to obtain an overview of the research gaps that might limit cross-scale generalization
125 of dendroecological research. We included studies published between January 1st 2005 and
126 June 1st 2022. We searched the Scopus database for published articles including the
127 following words in the title, abstract, or key words: ‘Arctic or Oro-Arctic or tundra’ and
128 ‘shrub or woody’ and ‘growth or recruitment or establishment or advancement or shrubline or
129 anatomy or xylem or lumen or fiber or vessel or dendroecology or dendrochronology or ring’.
130 Five-hundred and seventy articles were identified with this search (Appendix A).

131 **2.1.1 Inclusion criteria**

132 Of the 570 studies, 82 were included in our synthesis (Appendix B) based on the following
133 geographic and methodological criteria: The study 1) has at least one site located in the
134 Arctic, as defined by (Walker et al., 2005), or the ‘Oro-Arctic’ alpine tundra (Virtanen et al.,
135 2016), including sites located within a 50 km buffer outside of these boundaries, to be
136 consistent with previous literature (e.g. (Berner et al., 2020; Martin et al., 2017)); 2)
137 quantifies shrub growth or age using annual growth-rings or stem increments, and/or 3)
138 analyzes xylem anatomical traits. We organized the data both by study and by site (Appendix
139 A). If a study included multiple sites, we considered the sites as unique if: a) they were within
140 distinct geographic locations as defined by the authors, and b) when site-specific data was
141 readily available. If multiple studies included the same site but with different analyses, the
142 site was considered more than once. When articles included sites that were inside and outside
143 of the Arctic, we only report on sites meeting our Arctic criteria.

144 We included studies that stated “shrubs” were sampled, regardless of how the authors define
145 a shrub, but only if methods of dendrochronology adjusted for shrubs were used (i.e., not tree
146 cores). The shrub species included range from prostrate (e.g., *Salix arctica* in high Arctic
147 Canada and Greenland (Boulanger-Lapointe et al., 2014) to tall shrubs (e.g., >3m *Alnus*
148 *viridis* in West Greenland (Wilmking et al., 2018)).

149 **2.1.2 Extracted information**

150 For each included study, we recorded: 1) geographic coordinates of the study location(s); 2)
151 sampling strategy; 3) shrub part(s) sampled; 4) timespan of shrub growth and/or recruitment
152 records; 5) responses tested; and 6) drivers tested. When not given, geographic coordinates
153 were estimated based on the site description using Google Maps (maps.google.com).

154 We separated the sampling strategy into two main categories: 1) site selection (i.e., the
155 strategy used to select the specific sampling site in the landscape), distinguishing whether the
156 site selection was systematic or not; 2) shrub selection (i.e., the strategy used to select the
157 specific individual shrubs to be sampled) distinguishing whether individuals were selected
158 based on specific characteristics or using a systematic/unbiased approach. Here, we consider
159 systematically placed plots those placed purposefully within a landscape, for example, to
160 capture heterogeneity across an elevation gradient. We consider unbiased samples those
161 selected in a systematic way (e.g., the closest individual to a transect), and not based on plant
162 characteristics (e.g., the largest individual). We acknowledge that a systematic sampling
163 design may also create biases, but, compared to the frequent dendroecological practice of
164 targeted selection based on plant characteristics, these biases are more appropriate for
165 upscaling. In this manuscript, we use the term “unbiased” to refer to this reduction in bias
166 when comparing targeted selection and systematic sampling approaches.

167 For each study we extracted the maximum timespan of shrub growth and/or recruitment
168 based on the length of the chronology for all species combined. When the study reported the
169 age of the oldest shrub, we included that; otherwise, we included the time span of the final
170 chronology. When not stated, we estimated the time span from the relevant figures provided
171 in the manuscripts.

172 We identified responses measured using dendrochronological methods, grouping them as: 1)
173 growth; 2) recruitment; 3) quantitative wood anatomy; 4) qualitative wood anatomy and 5)
174 mortality. Growth was quantified either by 1) measuring radial ring-width increments; 2)
175 measuring annual stem-increments (for example, measuring distances between
176 wintermarksepta (dark bands of meristem tissue) (Rozema et al., 2009) of *Cassiope*
177 *tetragona*) (Myers-Smith et al., 2015b); or 3) dividing the radius or diameter of the stem by
178 the number of rings to obtain the average annual growth (e.g. (Schmidt et al., 2010))
179 Recruitment and mortality were included only if quantified using age estimates from
180 dendroecological methods (i.e. quantifying representative age dynamics based on ring counts
181 (e.g. (Büntgen et al., 2015))).

182 We extracted information on drivers of shrub dynamics that were explicitly tested in the
183 studies. Tested drivers were grouped into nine categories: 1) air temperature, 2) precipitation,
184 3) large scale climate systems, 4) biotic factors, interactions, and disturbances, 5) geophysical
185 factors (including soil moisture), processes, and disturbances, 6) growing season length and
186 timing, 7) snow cover and depth, 8) ice dynamics, e.g., sea ice, glacial retreats or indirect
187 influences of ice caps, and 9) other (including solar-related variables) (Table 1). While some
188 of these are direct mechanistic drivers (e.g., temperature and precipitation), others are indirect
189 and proxies for mechanistic drivers (e.g., ice dynamics and climate systems).

190 Below we report our findings based on the extracted information from the publications
191 identified in our targeted literature search that meet our selection criteria. Since we did not
192 perform a systematic and comprehensive review the specific percentages relate specifically to
193 our criteria and may not be representative of the entire literature.

194 **2.1.3 Illustrations and analyses**

195 We used bar graphs to visualize the research gaps identified in our review. Additionally, we
196 mapped the extracted study locations in relation to the bioclimate subzones from the
197 Circumpolar Arctic Vegetation Map (CAVM) (Raynolds et al., 2019) with the addition of the
198 Oro-Arctic (Virtanen et al., 2016) to visualize the spatial distribution of where drivers are
199 tested (Martin et al., 2017). Hereafter, we refer to the combined area of the CAVM and the
200 Oro-Arctic simply as “Arctic”, “tundra”, or “Arctic tundra”. We also assessed how
201 representative the study locations are with respect to the climate space of the tundra. For this
202 analysis, we retrieved mean annual temperature and annual precipitation for all study
203 locations, as well as 14,000 random locations across the Arctic tundra during the period from
204 1979-2013 using the CHELSA v1.2 bioclim dataset (Karger et al., 2017).

205 **2.2 HIGHLIGHTED GAPS, RESEARCH PRIORITIES, AND POTENTIAL WAYS FORWARD**

206 Based on our targeted literature review and the extracted information, we highlighted, what
207 we see, as five key research gaps that limit cross-scale generalizations in the current Arctic
208 dendroecological literature and discuss research priorities that address each gap. The key
209 gaps and related research priorities are: 1) Knowledge of variation in shrub responses across
210 and within species is limited. Including more species and individuals would enhance our
211 understanding of inter- and intraspecific variation, 2) Shrub demographic responses to Arctic
212 change are not yet fully understood. Including more life history responses (such as
213 recruitment and mortality, and their relationship to growth) would provide a more

214 comprehensive understanding, 3) Other potential drivers of shrub change aside from warming
215 have not been adequately explored. Increased focus on multiple drivers, their spatial and
216 temporal variation, and events could reveal insights into complex relationships, 4) Cross-site
217 and -study comparisons are limited by varying sampling methods. Systematic and unbiased
218 sampling of plant parts and individuals could reveal more cross-scale responses, and 5)
219 Insight on the structural-functional relationships behind growth trends is limited.
220 Incorporation of dendroanatomy could enhance this understanding.

221 **2.2.1 Gap 1: Knowledge of variation in responses across and within species is limited.**
222 **Including more species and individuals would enhance understanding of inter- and**
223 **intraspecific variation.**

224 *Finding: Single species studies are the norm*

225 Most studies assessed a dendroecological-derived response of just one species (67%), while
226 13% analyzed two species, 10% analyzed three to four species, and 11% analyzed over five
227 species (Fig. 2a). While 26% (21) analyzed multiple species at the same site, only five studies
228 used an unbiased sample from systematically placed plots. Twenty-seven studies (33%)
229 analyzed the same species across multiple sites, though only five of these used systematic and
230 unbiased sampling.

231 A total of 39 species from 12 genera were sampled across all studies. The three most studied
232 species were *Betula nana* (in 19 studies), *Cassiope tetragona* (in 17 studies) and *Salix glauca*
233 (in 15 studies). *Salix* was the most represented genus; eighteen different species were
234 included, and 44 studies considered at least one *Salix* species.

235 In summary, the included literature represents the most widely distributed genera of Arctic
236 shrubs. However, most studies focused on the same shrub taxa and investigated the response

237 of just one species. Relatively few (6%) used an unbiased sample to compare responses of
238 different species within the same sites or for the same species across multiple sites. This
239 leaves a gap in our ecological understanding of the variation in responses among individuals
240 of the same species across populations and species ranges, and of particular importance, in
241 responses among species within local plant communities.

242 *Research priority: Focus on inter- and intraspecific variation across sites*

243 Overcoming this gap requires studying multiple species at individual sites, as well as
244 studying the same species across multiple sites. Improved quantification of intra- and
245 interspecific responses is important for several reasons: First, different species might respond
246 to different drivers. For example, Weijers et al. (2018a), found that *Betula nana* growth in a
247 Norwegian alpine site was best explained by summer precipitation, while the growth of three
248 other species was best explained by summer temperature. Second, different species might
249 respond differently, and with different rates, to the same drivers, e.g., deciduous species have
250 been found to respond more rapidly to warming than evergreen ones (Demarco et al., 2014;
251 Elmendorf et al., 2012; Gough et al., 2014). Similarly, individuals of the same species are
252 expected to respond differently to the same driver across the geographic range, e.g.,
253 depending on local plant-plant interactions, micro-environmental conditions (Ackerman et
254 al., 2017) or limiting resources (Post et al., 2021). Third, different species might have
255 contrasting effects on ecosystems (Cahoon et al., 2012; Post et al., 2021). For example, it has
256 been hypothesized that deciduous shrub expansion will have positive feedback on global
257 warming, in contrast to overall negative feedback caused by evergreen shrub expansion, due
258 in part to their lower stature and production of more recalcitrant litter (Vowles and Björk,
259 2019).

260 *Steps forward: Analyzing more species and individuals, facilitated by data sharing*

261 Future studies should represent as many individuals and species as feasible within and among
262 sites. Although this intensive sampling approach is costly, data sharing can alleviate some of
263 the additional work by providing site- and species-specific chronologies to assist chronology
264 building and facilitating cross-species analyses. Myers-Smith et al. (2015a) provide an
265 example of such a cross-species and cross-site approach by compiling and analyzing growth
266 data from published and unpublished shrub chronologies of 25 species sampled across the
267 Arctic-Alpine tundra. However, to use the full potential of shrub data, e.g., to study
268 demographic responses other than growth (see Gap 2) and the underlying cellular responses
269 (see Gap 5), unbiased sampling (see Gap 4, cf. (Klesse et al., 2018)) and detailed metadata
270 across sites would be necessary (discussed further in section 3).

271 **2.2.2 Gap 2: Shrub demographic responses to Arctic change are not completely**
272 **understood. Including more life history responses (such as recruitment and mortality,**
273 **and their relationship to growth) would provide a more comprehensive understanding.**

274 *Finding: Growth was the most common response investigated*

275 In total, 89% of the studies analyzed growth, 27% recruitment, and 6% mortality. Only 18%
276 included both growth and recruitment (Fig. 2b, 2c), one of which also included mortality. The
277 shrubs in each study covered an average of 90 years (median: 84, range: 7-337; years: 1675-
278 2018) (Fig. 3).

279 Ultimately, recruitment and mortality are understudied aspects of shrub dynamics, in terms of
280 responses quantified using dendrochronological methods. Additionally, there is underutilized
281 potential of applying these existing shrub ring time-series to gain retrospective insights into
282 demographic parameters.

283 *Research priority: Quantify growth, recruitment, and mortality*

284 Recruitment is the driving mechanism behind shrub expansion at a landscape to global scale,
285 as it is what drives new individuals to fill in gaps in the current range and to advance range
286 limits, e.g., the northernmost shrubline (Büntgen et al., 2015; Myers-Smith et al., 2011).

287 While studies of recruitment are especially important for understanding range expansion,
288 increased insight on what drives mortality is important for understanding range contraction
289 (Hampe and Petit, 2005; Lesica and Crone, 2017). Simultaneously investigating the drivers of
290 growth, recruitment, and mortality is important for gaining a more comprehensive
291 understanding of local plant community responses and mechanisms underlying species range
292 dynamics (Normand et al., 2014).

293 Dendroecological methods have the unique ability to provide post-hoc time-series that enable
294 the assessment of drivers of demographic responses across time, for example, linking
295 previous growth and recruitment pulses to climate (Büntgen et al., 2018, 2015). Further
296 investigation of the drivers of growth and recruitment in long time-series is important for
297 several reasons. First, what drives growth and recruitment is not always the same. For
298 example, *Salix* spp. growth in Yukon, Canada was best explained by variation in summer
299 temperatures while recruitment pulses were best explained by variation in winter
300 temperatures (Myers-Smith and Hik, 2018). Second, drivers of recruitment vary among sites
301 and species. For example, while winter temperature drives variation in recruitment in the
302 Yukon (Myers-Smith and Hik, 2018), increasing summer temperatures have been found to
303 drive recruitment in eastern Greenland (Büntgen et al., 2015).

304 *Steps forward: Adjust sampling strategies to obtain demographic data from shrub*
305 *chronologies*

306 Obtaining data on age dynamics requires a different sampling approach than traditionally
307 used for growth data. To establish retrospective demographic time series, it is necessary to
308 sample a large and unbiased subset of the population to get an accurate estimate of the age
309 structure. This also requires sampling of the oldest part of the shrub to obtain the maximum
310 age estimation for each individual. Additionally, shrub mortality across life stages could be
311 quantified, by, e.g., long-term monitoring of individuals in permanent plots (such as with a
312 long-lived herb in (Edelfeldt et al., 2019)) or systematically collecting dead shrubs and dating
313 their year of death by using dendroecology or radio-carbon dating of their outermost ring (cf.
314 (Pizano et al., 2014). High-resolution landscape photomosaics, captured from overlooks with
315 phenocams or drones can also provide valuable baselines for identifying periods before/after
316 recruitment/mortality pulses. Taken in sequence over many years at long-term monitoring
317 sites, these data can provide landscape-scale insights into such processes.

318 **2.2.3 Gap 3: Other potential drivers of shrub change aside from warming have not been**
319 **adequately explored. Increased focus on multiple drivers, their spatial and temporal**
320 **variation, and events could reveal insights into complex relationships.**

321 *Finding: Summer warming was the most common driver investigated*

322 Most studies focused on climate as the main driving factor, with 76% and 53% of the studies
323 investigating summer temperature and precipitation, respectively (Fig. 2e, Table 1). In total,
324 we identified 39 different drivers, but most of these have been included in only 1 study each
325 (mean: 7.8, median: 2, range: 1-62) (Table 1). Relating to the limited temporal availability for
326 the different groups of drivers (Appendix C), time-series were only used for a subset of the
327 drivers (e.g., climate, sea ice) while many drivers were assessed by average site

328 characteristics. Furthermore, most studies focus on responses to mean conditions and trends
329 with only 21% considering extreme abiotic or biotic events (Fig. 2h).

330 The focus on few climatic drivers leaves a gap in our ecological understanding of how the
331 climate sensitivity of shrubs is influenced by other factors including micro-environmental
332 conditions, plant-plant and trophic interactions, and mostly overlooks the effects of extreme
333 events, such as icing or the loss of snow cover in winter (Phoenix and Bjerke, 2016).

334 *Research priority: Investigate a variety of potential drivers*

335 Closing this gap requires more studies to focus on the interaction between shrub responses
336 and variation in abiotic and biotic drivers across time and space, including both trends and
337 events. Consideration of time-series of non-climatic drivers is important for several reasons.
338 First, although climate has long been known to affect shrub growth, considerable variation
339 between sites has been observed (Myers-Smith et al., 2015a). With a decoupling observed
340 between tree growth and climate in recent decades (Martin et al., 2017; Wilmking et al., 2020), it is clear
341 there are other driving and limiting factors to be considered, e.g., soil moisture has been
342 identified as a key limiting factor for shrub growth and driver of variation among Arctic sites
343 (Ackerman et al., 2017; Myers-Smith et al., 2015a).

344 Second, the climatic data frequently used is often not ecologically meaningful in terms of
345 spatial and temporal relevance to the actual conditions impacting shrub growth (Körner and
346 Hiltbrunner, 2018). For instance, the growing conditions experienced by low-lying shrubs,
347 where much of the growth occurs near the soil surface, likely differs from the annual mean
348 temperatures recorded 2m above ground kilometers away from the sampled shrub. In
349 addition, the same conditions may have very different consequences on seedlings in early
350 spring compared to dormant stems in the winter. Thus, recording of the actual growing
351 conditions at the relevant scale for shrub growth as well as the relevant timing in relation to

352 phenology and life stage is essential to gain a holistic understanding of the drivers behind
353 shrub growth.

354 Third, responses of Arctic shrubs to abiotic and biotic disturbances are understudied even
355 though both have been identified as important co-drivers for tundra vegetation growth and
356 expansion (Myers-Smith et al., 2011). For example, wildfires may have greater effect on
357 alder recruitment than temperature in Canada (Lantz et al., 2010); defoliation events due to
358 insect outbreaks influence the climate sensitivity of *Salix glauca* and *Alnus viridis* growth in
359 western Greenland (Prendin et al., 2020; Wilmking et al., 2018); and the presence of large
360 herbivores alters shrub response to warming in western Greenland (Post et al., 2021; Post and
361 Pedersen, 2008).

362 *Steps forward: Utilize in-situ and remote sensing-derived data*

363 Although growth time series can span many decades, we typically lack such temporal data for
364 potential ecological covariates (e.g., temperature, precipitation, length of growing season,
365 etc.) (Fig. 3, Appendix C). This is due to either the lack of long-time meteorological station
366 data or local environmental variation (e.g., soil moisture) only being measured at the time of
367 sampling. Remote sensing may be used to overcome this gap in some cases (further discussed
368 in Section 3).

369 Various sources of remote sensing data can be used to measure or estimate most, but not all,
370 of the drivers identified in our review at least to some degree (Table 1). Contemporary
371 satellite-derived measures (e.g., such as indices of soil moisture or topography variables
372 based on satellite-derived digital elevation models) can provide information at grain sizes
373 suitable for cross-site and increasingly also intra-site studies. These measures, including
374 gridded datasets, provide opportunities to incorporate the potential covariates and scales of
375 inquiry needed to better identify and evaluate ecological drivers. For example, they can

376 provide information about spatial and temporal patterns of disturbance events or maps of
377 historical influences on contemporary ecological processes (Babst et al., 2010; Prendin et al.,
378 2020).

379 While gridded datasets and weather stations can provide valuable information on trends
380 within sites and relative differences among sites, the conditions experienced by low-lying
381 shrubs can be drastically different than those quantified using such methods. Therefore, to
382 truly understand the mechanisms behind shrub responses, quantification of the growing
383 conditions at the relevant spatial and temporal scales for the shrub is necessary (Körner and
384 Hiltbrunner, 2018). Many downscaling approaches are emerging in the literature (e.g.
385 Microclima (Maclean et al., 2018)) but they remain unvalidated and little used in published
386 empirical dendroecological studies to date. Recording *in-situ* microsite growing conditions is
387 now more feasible and affordable through recent sensor developments, e.g., measuring
388 synchronous soil moisture, as well as soil, surface, and air temperature using specialized
389 loggers (e.g., Tomst TMS-4 ([Wild et al., 2019](#))), and can reveal patterns not discernable with
390 the more common macro-site data (Lembrechts et al., 2020).

391 **2.2.4 Gap 4: Cross-site and -study comparisons are limited by varying sampling**
392 **methods. Systematic and unbiased sampling of plant parts and individuals could reveal**
393 **more cross-scale responses**

394 *Finding: Sampling methods varied between studies and sites*

395 Sampling efforts vary across climatic (Fig. 4a) and geographic (Fig. 4b) space for all drivers
396 and responses. Less than half (43%) of the included studies systematically selected sampling
397 sites and even fewer (33%) sampled shrub individuals using an unbiased approach (Fig. 2f).
398 When considering site and shrub selection combined, 23% of studies sampled shrubs using
399 unbiased criteria in systematically chosen sites, and 2% selectively sampled shrubs in non-

400 systematically chosen sites (Fig. 2f). The majority of studies (82%) analyzed either root
401 collars or stems, while only 12% included below ground parts (Fig. 2g). Most studies (44%)
402 analyzed one shrub part, while fewer analyzed two or three shrub parts (26% and 11%) (Fig.
403 2g).

404 While sites were distributed across the Arctic, responses and drivers were not all assessed
405 evenly across regions and bioclimatic zones (Fig. 4), thus potentially biasing our
406 understanding of ongoing changes across the Arctic (Metcalf et al., 2018). Additionally,
407 differences among sampling techniques (shrub and site selection and shrub part sampled)
408 make it challenging to compare data and draw generalizable conclusions from different
409 studies.

410 *Research priority: Implement sampling approaches that allow for comparison of shrub*
411 *responses across individuals and sites*

412 Closing this gap requires sampling that captures the variation across the tundra biome, both in
413 shrub responses and in environmental conditions. Unbiased sampling is important in relation
414 to several research topics. First, while it is common in dendroecological studies to sample the
415 largest individuals in the attempt to maximize chronology length, this approach does not give
416 insight into the range of responses in a population. Trees selected to maximize climate signal
417 have been found to overestimate climate sensitivity (Klesse et al., 2018), which can
418 potentially skew our understanding of ongoing and future responses to climate change. In
419 generally harsh environments such as the Arctic, we might further expect the tallest/oldest
420 individuals to grow and survive at favorable and protected microsites. To our knowledge such
421 biases have not yet been directly tested for Arctic shrubs.

422 Second, different shrub parts are sampled across different sites and studies (Fig. 2g) (Ropars
423 et al., 2017). This non-uniform sampling is problematic when attempting to generalize

424 findings and compare datasets because shrub climate sensitivity is non-uniform within
425 individuals (Ropars et al., 2017; Shetti et al., 2018) thus hindering accurate cross- individual
426 and site comparisons. While serial sectioning of above and below parts is well established
427 and recommended to aid with cross-dating (Bär et al., 2006; Buchwal et al., 2013;
428 Kolishchuk, 1990; Myers-Smith et al., 2015b; Wilmking et al., 2012), it is less common to
429 use these sections to compare growth sensitivities along the length of the stem and roots to
430 assess intra-individual variation. Notably, below ground relationships remain understudied,
431 despite most of Arctic plant biomass occurring below ground (Iversen et al., 2015; Mokany et
432 al., 2006) and being especially important for soil dynamics and carbon storage.

433 *Steps forward: Representative sampling of sites and individuals*

434 Representative sampling of sites and individuals can be reached through standardized
435 protocols, data sharing, and the use of remote sensing derived stratifications as the basis for
436 dendroecological sampling designs. This will ultimately facilitate the selection of comparable
437 sections, individuals and sampling sites that cover the ecological variation of interest for the
438 specific study questions (discussed further in Section 3).

439 **2.2.5 Gap 5: Insight on the structural-functional relationships behind growth trends is**
440 **limited. Incorporation of dendroanatomy could enhance this understanding**

441 *Finding: Few studies assessed functional mechanisms behind growth trends*

442 In the included studies, 10% (8) of the studies assessed xylem anatomy, with seven assessing
443 quantitative (i.e., cell lumen area, cell grouping and cell wall thickness) and one assessing
444 qualitative (reaction wood and scars) anatomical features (Fig. 2b). Out of these eight studies,
445 one assessed three responses (growth, recruitment, and qualitative wood anatomy); five
446 assessed growth and quantitative wood anatomy; and two focused solely on quantitative

447 wood anatomy (Fig. 2c). In summary, few studies combine assessment of xylem traits
448 together with ring widths, leaving a gap in our understanding in the mechanisms behind the
449 variation of intra- and inter- individual responses to ecological drivers.

450 *Research priority: Investigate cellular mechanisms behind growth responses*

451 Closing this gap requires analyses of shrub responses at a cellular level to fully understand
452 growth trends. Xylem functional traits (e.g., the ones related to hydraulic transport and
453 mechanical support) have been shown to provide valuable information on functional and
454 structural adaptation (Fonti et al., 2010). First, while ring widths quantify annual growth,
455 anatomical traits reveal insights of plant functioning on an intra-annual scale (von Arx et al.,
456 2016) (further discussed in Section 3.2). Second, anatomical traits may reveal shrub
457 responses to growing conditions that are not detectable with just annual ring widths. For
458 example, Buras et al. (2017) found that cell wall thickness, but not ring width, in *Juniperus*
459 *communis ssp. nana* in Kobbefjord (southwest Greenland) was significantly correlated with
460 Greenland Ice Sheet melt. It is possible that no significant relationship was found with ring
461 widths due to a higher sensitivity of radial growth to episodically occurring moth outbreaks
462 (Buras et al., 2017). Third, anatomical features can impact the climate-growth relationship.
463 For example, Nielsen et al. (2017) found that *Betula nana* in western Greenland optimized
464 growth by increasing its hydraulic conductivity. These examples demonstrate how
465 dendroanatomy can be incorporated into dendroecological studies to lead to a better
466 understanding of shrub responses and the mechanisms behind them.

467 *Steps forward: Adjust dendroecological protocols to incorporate anatomical analyses*

468 Although still uncommon, recent studies have shown the potential for quantifying xylem
469 anatomical traits in dendroecological shrub studies. The inclusion of such analyses can give a
470 more comprehensive picture of changes that occur within shrubs and can allow for

471 estimations of future changes in the community, for example, why some species or
472 individuals thrive more than others at a given site (Huang et al., 2021). Notably, linking
473 anatomical features and their relative position within growth rings to variables of interest can
474 provide further mechanistic understanding behind the cellular responses at a higher (intra-
475 annual) temporal resolution (Fonti et al., 2010; Lenz et al., 2013). Incorporating such
476 microscopic features into dendroecological studies requires careful sample preparation
477 (Gärtner et al., 2015) and techniques (von Arx et al., 2016) that can be included in protocols
478 that have already been implemented for the small size of shrub rings (Schweingruber et al.,
479 2013). In addition to dendroanatomy, the field is further moving towards fine-scale analyses
480 with advancements in technology and techniques, for example, using high-precision
481 dendrometers to capture variations in shrub stems at unprecedented resolutions (Dobbert et
482 al., 2021b, 2021c).

483

484 **3. SYNTHESIS: CROSS-SCALE INTEGRATION FROM CELLS TO SATELLITES** 485 **TO IMPROVE OUR UNDERSTANDING OF ARCTIC SHRUB COMMUNITY** 486 **DYNAMICS**

487 Our targeted review summarizes the methodologies used in recent Arctic dendroecological
488 studies. Despite the numerous recent advances in the field, we found that in our selected
489 literature: a) most studies focused on how the growth of a limited number of species is
490 affected by trends in specific climatic drivers, mostly summer temperature and precipitation
491 (Gaps 1-3), b) sampling is not uniform across geographic or climatic gradients (Gap 4), and
492 c) only few studies investigate several plant parts and the mechanisms behind the observed
493 shrub responses (Gap 4, 5). More generally, this reflects a field that is producing diverse and
494 informative research, but few studies that are specifically designed to produce cross-scale
495 insights. To address this, we propose that research priorities should more often encompass the
496 quantification of additional responses (e.g., growth, recruitment, and mortality), and their

497 cellular mechanistic underpinnings, of a range of plant parts, individuals and species sampled
498 unbiasedly within and across sites. Doing so will require comprehensive sampling; cellular
499 analyses; and data on trends and events of multiple drivers derived from *in-situ* and satellite-
500 based measurements. This would ultimately allow for linking fine-scale individual shrub
501 responses to large-scale trends and events occurring across the tundra landscapes and biome.
502 However, conducting such multidisciplinary studies is challenging and requires a synergetic
503 approach.

504 One possible approach to upscale *in-situ* information to larger scales is to combine emerging
505 approaches in dendroanatomical and remote sensing analyses with a plant community-based
506 perspective on dendroecological sampling. By presenting this approach we hope to inspire
507 more dendro and Arctic ecologists to adopt spatially unbiased, systematic, and cross-scale
508 sampling methodologies, which we think are needed to further our ecological understanding
509 of shrub dynamics across the rapidly changing Arctic.

510 In the following subsections we briefly outline the three main components of the proposed
511 synergetic approach: dendroanatomical analyses (section 3.1); remote sensing and geospatial
512 analyses (section 3.2), and community-based dendroecological (CBDE) sampling (section
513 3.3) (Fig. 5). We then conclude with a synthesis of the approach (section 3.4) and provide
514 examples of ecological questions that such an approach can address (section 3.5).

515 **3.1. Dendroanatomy as a key to understand the cellular mechanistic underpinnings** 516 **behind growth responses at inter-annual level**

517 Dendroanatomical analyses, based on quantitative wood anatomy, will provide novel and high-
518 resolution information from dated woody plant annual rings (Fig. 5f) (Fonti et al., 2010) and
519 insight in the cellular mechanistic underpinnings behind inter-annual growth responses (Gap
520 5). Adjustments in the xylem structure remain permanently fixed and chronologically archived

521 in the annual growth-rings (Fonti et al., 2010). Thus, they provide an explicit time record of
522 structural and functional responses, linking biological processes and plant functioning (Fonti
523 and Jansen, 2012). The analysis of xylem structure, therefore, allows to reconstruct growth
524 dynamics under different climatic and environmental conditions over decades and even
525 centuries (von Arx et al., 2016).

526 While growing, plants continuously adjust their structure to achieve an optimal balance of
527 carbon costs for the competing biomechanical and hydraulic requirements that sustain
528 transpiration and thus carbon assimilation (Prendin et al., 2018a). The axial organization of
529 cells follows a pattern common to all vascular plants (Anfodillo et al., 2013; Olson et al., 2014).
530 Conduits progressively increase in diameter from the apex towards the stem base, which
531 minimizes the adverse effect of height growth on pathway length resistance (Carrer et al.,
532 2014). This pattern is also reflected radially with the conduits increasing in dimension from the
533 pith outwards (Carrer et al., 2014), but largely stable during ontogeny (Prendin et al., 2018b).
534 To disentangle how climatic and environmental factors are influencing the structural plant
535 development it is fundamental to use an appropriate standardization procedure to remove the
536 hydraulic-path length effect (Lechthaler et al., 2019) on the different anatomical traits (Prendin
537 et al., 2018b).

538 The recent advancements in methods and standard protocols (Gärtner and Schweingruber,
539 2013; von Arx et al., 2016), automated image-analysis systems (von Arx et al., 2013; von Arx
540 and Carrer, 2014; von Arx and Dietz, 2005) and processing using artificial intelligence
541 (Resente et al., 2021) allow to significantly increase the number of measured anatomical
542 features, while also reducing the time required for the analysis. Nevertheless, the establishment
543 of the new quantitative wood anatomy network (Q-NET) reflects the need of a growing
544 community to share knowledge and experience, but also to advance methodologies (Resente et

545 al., 2021) and to profit from interdisciplinary collaborations to tackle broader ecological
546 questions at multiple temporal and spatial scales (von Arx et al., 2021).

547 **3.2 Incorporating remote sensing and geospatial analyses to promote cross-scale** 548 **understanding**

549 Remote sensing and geospatial analyses are essential tools for linking site-based drivers and
550 responses to ecological patterns that span across spatiotemporal scales that are infeasible to
551 study with *in-situ* measurements. There are a variety of available remote sensing products and
552 approaches, spanning from direct measures of spectral or structural properties of a landscape
553 to more derived products. In the approach proposed here, remote sensing contributes to
554 bridge Gap 3 (lack of drivers studied) and 4 (lack of multi-site analyses) by allowing for: 1)
555 the selection of representative sampling of shrub individuals, within study sites, based on
556 landscape stratification; and 2) the collection of additional data on vegetation responses and
557 potential drivers at relevant scales to assess shrub responses to a wide range of variables.
558 These two aspects are discussed in further detail below.

559 The landscape of an Arctic study site might include large environmental variation at different
560 scales due to topography and small-scale heterogeneity (Fig. 5a,b). A useful method for the
561 selection of representative study sites is to stratify the landscape to select locations that
562 represent the variation of the relevant co-variates based on the ecological questions and
563 drivers of interest. Methodologies that classify and stratify vegetation as well as
564 environmental parameters such as climate and soil moisture are well established at global,
565 regional (Lara et al., 2018; Raynolds et al., 2019), and local scales (Bartsch et al., 2020;
566 Boelman et al., 2016). As an example, landscapes can be stratified based on elevation (e.g.,
567 using satellite-derived digital elevation models), proxies for plant available water (e.g., using

568 normalized difference water index) and vegetation productivity (e.g., using normalised
569 difference vegetation index, NDVI) (Fig. 5a,c).

570 Additionally, remote sensing can be used to collect data on a range of potential drivers as
571 well as on vegetation changes. Satellite-derived indices have provided evidence of
572 widespread Arctic vegetation change, for example, increasing NDVI values show a greening
573 of the Arctic tundra in many regions (Berner et al., 2020). Drones (Assmann et al., 2020;
574 Cunliffe et al., 2020; Fraser et al., 2016; Siewert and Olofsson, 2020), phenocams
575 (Richardson et al., 2018), and repeat photography (Myers-Smith et al., 2019a; Tape et al.,
576 2006) can be utilized to capture fine-scale patterns and dynamics of Arctic tundra
577 communities to bridge scale gaps between site- and satellite-based data. For example, drone
578 flights over a sampling area can obtain high-resolution (centimeter-scale) visual,
579 multispectral and thermal imagery, and detailed digital elevation models using drone-based
580 light detection and ranging (LiDAR) or structure from motion (SfM) photogrammetry.

581 Challenges and opportunities using satellite imagery and its interpretation in the Arctic
582 (Beamish et al., 2020; Berner et al., 2020; Myers-Smith et al., 2020) along with the
583 challenges and opportunities for combining dendroecological studies of trees with remote
584 sensing (Babst et al., 2018) have been recently discussed in more detail. These studies also
585 highlight the importance of ensuring the limitations and intricacies of remote sensing
586 measures are accounted for in the analyses and interpretation of the acquired data,
587 particularly with reference to scale and proxies. In addition, there is a need for continued
588 updating of dendrochronologies (Babst et al., 2017; Larson et al., 2013) as they can quickly
589 lag behind the recent rapid environmental changes and most current remote sensing datasets
590 (Fig. 3).

591 3.3 A plant community-based perspective for dendroecological studies

592 Community-based dendroecological (CBDE) sampling can be used to increase understanding
593 of shrub growth, expansion, and demographics (Gap 2) and how the responses of shrub
594 individuals, species, and communities relate to micro-environmental conditions, plant-plant
595 interactions, and larger scale climatic factors (Gap 1, Gap 3). Yet, it has not been widely
596 implemented (but see, for example, (Büntgen et al., 2015)). This CBDE approach focuses on
597 collecting a subset of shrubs that is representative of the study site (Gap 4). This implies,
598 however, sampling a larger number of individuals at both plot- and site-level than what is
599 typical in dendro-based protocols. Using systematic unbiased sampling strategies of a large
600 number of individuals (Fig. 5b,d) will allow for assessing within-population variation and age
601 dynamics, thus allowing for the study of retrospective growth and establishment patterns
602 (Fig. 5f). Compared to a sampling strategy that focuses on the largest individuals, unbiased
603 sampling might result in a shorter chronology with more variation in climate-growth patterns.
604 It will, however, be better suited for answering ecological questions on shrub communities
605 (e.g., see section 3.5). To assess demographic responses, it is essential to get the best estimate
606 possible of shrub ages by extracting complete shrubs to collect the root collar. Furthermore,
607 sampling of the whole individuals allows for assessment of growth responses of different
608 shrub parts (Ropars et al., 2017; Shetti et al., 2018) and quantification of biomass (Berner et
609 al., 2015). In addition, we suggest collecting a range of *in-situ* data both on the sampled
610 individuals and the surrounding growing conditions. Depending on research questions, such
611 data could include information on functional traits, e.g., plant height and specific leaf area
612 (Fig. 5e), the exact location of the sampled individual, vegetation cover, soil condition, and
613 micro topography.

614 3.4 Synthesis of synergetic approach

615 In summary, the core of our proposed synergetic approach is to collect a representative
616 sample of shrubs from across and within environmental gradients by stratifying the landscape
617 using remote sensing tools and geospatial analyses. This will allow for cross-scale
618 investigations and demographic assessments of shrub communities and individuals (i.e.,
619 contributing to bridging Gap 1 and 4). While excavating entire shrubs and collecting both
620 above and below ground parts allows for the collection of the oldest part of the shrub (the
621 root collar) which is necessary for age estimation, this also provides the opportunity to assess
622 intra- individual variation in responses to various drivers obtained from remote sensing and
623 *in-situ* data (Gap 2-4). Combined with detailed analyses of cellular responses and functional
624 traits (e.g., stem, root, leaf), this would allow for multidisciplinary studies linking different
625 responses and ecosystem functions across time and space (Gap 5).

626 Our approach is not without limitations, mainly due to the inherent difficulties when working
627 with tiny and irregular growth rings of Arctic shrubs and the costly workload involved with
628 collecting and analyzing large numbers of individuals from remote locations. However,
629 making data comparable across sites and studies could help alleviate some of the practical
630 and logistical constraints of data collection by facilitating further collaboration and data
631 sharing (e.g. through open access repositories). Additionally, the approach can and should be
632 adjusted to better suit particular research questions, needs, and constraints. For example, if
633 long and clear time-series are required, typical in dendroclimatic studies, CBDE sampling can
634 be supplemented by additionally collecting the largest individuals at a site - as long as such
635 selectively sampled individuals are clearly marked and are not taken into account in
636 demographic assessments requiring representative sampling.

637 To facilitate synergetic integration of dendroanatomy, remote sensing, and CBDE-sampling
638 across sites, there is a need for standardized analytical methods and detailed metadata. This is
639 especially important since shrub-rings and anatomical features are not uniform within
640 individuals. Standardized sampling, sample preparation, and analyses will help with
641 disentangling environmental effects from axial trends of xylem anatomical traits (Anfodillo et
642 al., 2013; Carrer et al., 2014; Lechthaler et al., 2019; Olson and Rosell, 2013) and shrub-
643 rings. The goal of this being to make data comparable between sampled individuals, sites
644 (Lechthaler et al., 2019), and growth forms (e.g., climbing, prostrate etc.). Extensive
645 metadata with information on how the plant was selected (Klesse et al., 2018), where on the
646 plant the wood was sampled, and distance of the sample to the shoot apex (Fig. 5e)
647 (Lechthaler et al., 2019; Olson et al., 2021) should be included in all studies. Recording
648 coordinates of the location of each collected individual will facilitate the integration of
649 satellite derived environmental data. Finally, recommended methods of microsection
650 preparation and measurements of shrub rings (Myers-Smith et al., 2015b) should be followed,
651 giving attention to prepare high quality microsections which will allow quantification of
652 anatomical features such as lumen area, cell-wall thickness, and the number and distribution
653 of the main cell types (von Arx et al., 2016).

654

655 We hope the approach presented here can be adopted in future studies to further our
656 ecological understanding of shrub dynamics across the rapidly changing Arctic. The wealth
657 of data that can be attainable with such synergic approaches will allow addressing a larger
658 variety of questions and will foster collaboration both within the field of dendroecology and
659 between dendroecology and other scientific fields.

660 **3.5 Examples of key ecological questions that could be addressed**

661 The synergetic approach can be used to answer a wide range of interdisciplinary questions
662 and can be adjusted to fit different needs, here we highlight three examples:

663 *Example 1: Linking ring widths with biomass*

664 With the collection of whole shrubs, ring widths can be linked with biomass-allometry
665 functions and plant traits. Knowledge on species-specific allometric relationships (Berner et
666 al., 2015) and tools to more easily include allometry in dendroecological studies (Kašpar et
667 al., 2019) can be incorporated to study biomass-allometry patterns across a range of species
668 and site characteristics. This area of research is gaining attention because of its important role
669 in carbon storage and how it provides the possibility for ring width data to contribute to
670 vegetation-climate and carbon models.

671 *Example 2: Linking microscopic responses to remotely-sensed patterns*

672 With the incorporation of data from a micro (e.g., cells, rings) to a macro scale (e.g.,
673 satellites), local field-based biotic and abiotic measurements can be linked with remotely
674 sensed information (e.g., satellite derived indices) across time and space. For example, it
675 could be used to assess the links between NDVI and ring widths (Macias-Fauria et al., 2012;
676 Ropars et al., 2015; Weijers et al., 2018b), as well as xylem anatomical traits (Prendin et al.,
677 2020). Systematically sampling across the landscape, quantifying shrub growth and
678 anatomical traits, and using near surface remote sensing devices (e.g., drones, phenocams)
679 can be used to help build the bridges between microscopic responses, field-based
680 observations, and patterns observed by satellites that will allow us to gain a more complete
681 picture of Arctic shrub dynamics.

682 *Example 3: Linking above and below ground traits*

683 The collection of information on a wide range of plant traits from a variety of systematically
684 selected individuals, e.g., wood (density, anatomy), leaf (specific leaf area, leaf phosphorus

685 concentration, leaf dry matter content), below ground (root growth), and structural
686 characteristics (maximum height) (Iturrate-Garcia et al., 2020; Kunstler et al., 2016), enables
687 insight into links between different traits among individuals, species, and locations. As an
688 example, collecting data on full shrubs can facilitate more below-ground trait research, which
689 is particularly important, but still understudied in tundra ecosystems (Bjorkman et al., 2018;
690 Iversen et al., 2015; Myers-Smith et al., 2019b) where most biomass allocation is below
691 ground (Iversen et al., 2015; Mokany et al., 2006). This multidisciplinary approach can
692 provide a more comprehensive understanding of the impacts of vegetation change and
693 ecosystem processes (Díaz et al., 2007, 2004).

694 **4. OUTLOOK**

695 Dendroecology provides a unique opportunity to study current and past shrub dynamics in the
696 drastically changing Arctic. Here we propose a synergetic approach that combines a
697 comprehensive and community-based sampling design; quantitative wood anatomical
698 analyses that can reveal underlying functional mechanisms; and satellite-based and *in-situ*
699 data describing trends and events of multiple drivers. Such a multidisciplinary and cross-scale
700 community-based perspective in dendroecological studies could improve insight in relation to
701 understanding: (i) the cellular mechanistic underpinnings of inter- and intraspecific variation
702 in shrub growth (e.g., adjustments in xylem functional traits); (ii) the links between
703 demographic responses (e.g., growth, recruitment, mortality) and drivers (e.g., abiotic and
704 biotic trends and events), within and among shrubs communities; (iii) links and allocation
705 between above and below ground plant responses; and (iv) how plant responses translate into
706 ecosystem functions. These insights, achieved through the implementation of standardized
707 protocols, data sharing, dendroanatomical analyses, and remote sensing, can ultimately be

708 used to improve quantification of the magnitude and extent of current shrub community
709 responses.

710 Development of efficient management and conservation strategies requires information at
711 local to landscape/regional scales. As there is remaining uncertainty regarding how Arctic
712 vegetation is changing and will continue to change on local scales in the current period of
713 rapid warming, this improved understanding is necessary for predicting the impacts on Arctic
714 biodiversity and ecosystems across time and space. These comprehensive predictions can
715 ultimately better inform the development of local and regionally relevant climate change
716 conservation and mitigation strategies.

717

718 **APPENDIX A: LITERATURE REVIEW DATA**

719 **APPENDIX B: ARTICLES INCLUDED IN REVIEW**

720 **APPENDIX C: EXAMPLES OF GRIDDED GLOBAL DATASETS**

721 **REFERENCES**

722 Abdalati, W., 2007. Greenland Ice Sheet Melt Characteristics Derived from Passive
723 Microwave Data: 1979-2007. Digit. Media.

724 Ackerman, D., Griffin, D., Hobbie, S.E., Finlay, J.C., 2017. Arctic shrub growth trajectories
725 differ across soil moisture levels. *Glob. Chang. Biol.* 23, 4294–4302.

726 <https://doi.org/10.1111/gcb.13677>

727 AMAP, 2021. Arctic Climate Change Update 2021: Key Trends and Impacts. Oslo.

728 AMAP, 2017. Snow, Water, Ice and Permafrost in the Arctic (SWIPA). Oslo.

729 Anderson, K., Gaston, K.J., 2013. Lightweight unmanned aerial vehicles will revolutionize

730 spatial ecology. *Front. Ecol. Environ.* 11, 138–146. <https://doi.org/10.1890/120150>

731 Anfodillo, T., Petit, G., Crivellaro, A., 2013. Axial conduit widening in woody species: A
732 still neglected anatomical pattern, in: *IAWA Journal*. pp. 352–364.
733 <https://doi.org/10.1163/22941932-00000030>

734 Assmann, J.J., Myers-Smith, I.H., Kerby, J.T., Cunliffe, A.M., Daskalova, G.N., 2020. Drone
735 data reveal heterogeneity in tundra greenness and phenology not captured by satellites.
736 *Environ. Res. Lett.* 15, 125002. <https://doi.org/10.1088/1748-9326/abbf7d>

737 Babst, F., Bodesheim, P., Charney, N., Friend, A.D., Girardin, M.P., Klesse, S., Moore,
738 D.J.P., Seftigen, K., Björklund, J., Bouriaud, O., Dawson, A., DeRose, R.J., Dietze,
739 M.C., Eckes, A.H., Enquist, B., Frank, D.C., Mahecha, M.D., Poulter, B., Record, S.,
740 Trouet, V., Turton, R.H., Zhang, Z., Evans, M.E.K., 2018. When tree rings go global:
741 Challenges and opportunities for retro- and prospective insight. *Quat. Sci. Rev.* 197, 1–
742 20. <https://doi.org/10.1016/j.quascirev.2018.07.009>

743 Babst, F., Esper, J., Parlow, E., 2010. Landsat TM/ETM+ and tree-ring based assessment of
744 spatiotemporal patterns of the autumnal moth (*Epirrita autumnata*) in northernmost
745 Fennoscandia. *Remote Sens. Environ.* 114, 637–646.
746 <https://doi.org/10.1016/j.rse.2009.11.005>

747 Babst, F., Poulter, B., Bodesheim, P., Mahecha, M.D., Frank, D.C., 2017. Improved tree-ring
748 archives will support earth-system science. *Nat. Ecol. Evol.* 1, 1–2.
749 <https://doi.org/10.1038/s41559-016-0008>

750 Bär, A., Bräuning, A., Löffler, J., 2006. Dendroecology of dwarf shrubs in the high
751 mountains of Norway - A methodological approach. *Dendrochronologia* 24, 17–27.
752 <https://doi.org/10.1016/j.dendro.2006.05.001>

753 Bär, A., Pape, R., Bräuning, A., Löffler, J., 2008. Growth-ring variations of dwarf shrubs
754 reflect regional climate signals in alpine environments rather than topoclimatic
755 differences. *J. Biogeogr.* 35, 625–636. <https://doi.org/10.1111/j.1365-2699.2007.01804.x>

756 Bartsch, A., Widhalm, B., Leibman, M., Ermokhina, K., Kumpula, T., Skarin, A., Wilcox,
757 E.J., Jones, B.M., Frost, G. V., Höfler, A., Pointner, G., 2020. Feasibility of tundra
758 vegetation height retrieval from Sentinel-1 and Sentinel-2 data. *Remote Sens. Environ.*
759 237, 111515. <https://doi.org/10.1016/j.rse.2019.111515>

760 Beamish, A., Reynolds, M.K., Epstein, H., Frost, G. V., Macander, M.J., Bergstedt, H.,
761 Bartsch, A., Kruse, S., Miles, V., Tanis, C.M., Heim, B., Fuchs, M., Chabrillat, S.,
762 Shevtsova, I., Verdonen, M., Wagner, J., 2020. Recent trends and remaining challenges
763 for optical remote sensing of Arctic tundra vegetation: A review and outlook. *Remote*
764 *Sens. Environ.* 246, 111872. <https://doi.org/10.1016/j.rse.2020.111872>

765 Belke-Brea, M., Domine, F., Barrere, M., Picard, G., Arnaud, L., 2020. Impact of shrubs on
766 winter surface albedo and snow specific surface area at a low arctic site: In situ
767 measurements and simulations. *J. Clim.* 33, 597–609. [https://doi.org/10.1175/JCLI-D-](https://doi.org/10.1175/JCLI-D-19-0318.1)
768 [19-0318.1](https://doi.org/10.1175/JCLI-D-19-0318.1)

769 Belshe, E.F., Schuur, E.A.G., Grosse, G., 2013. Quantification of upland thermokarst features
770 with high resolution remote sensing. *Environ. Res. Lett.* 8. [https://doi.org/10.1088/1748-](https://doi.org/10.1088/1748-9326/8/3/035016)
771 [9326/8/3/035016](https://doi.org/10.1088/1748-9326/8/3/035016)

772 Berner, L.T., Alexander, H.D., Loranty, M.M., Ganzlin, P., Mack, M.C., Davydov, S.P.,
773 Goetz, S.J., 2015. Biomass allometry for alder, dwarf birch, and willow in boreal forest
774 and tundra ecosystems of far northeastern Siberia and north-central Alaska. *For. Ecol.*
775 *Manage.* 337, 110–118. <https://doi.org/10.1016/j.foreco.2014.10.027>

776 Berner, L.T., Massey, R., Jantz, P., Forbes, B.C., Macias-Fauria, M., Myers-Smith, I.,

777 Kumpula, T., Gauthier, G., Andreu-Hayles, L., Gaglioti, B. V., Burns, P., Zetterberg, P.,
778 D'Arrigo, R., Goetz, S.J., 2020. Summer warming explains widespread but not uniform
779 greening in the Arctic tundra biome. *Nat. Commun.* 11, 1–12.
780 <https://doi.org/10.1038/s41467-020-18479-5>

781 Bhatt, U.S., Walker, D.A., Raynolds, M.K., Bieniek, P.A., Epstein, H.E., Comiso, J.C.,
782 Pinzon, J.E., Tucker, C.J., Polyakov, I. V., 2013. Recent declines in warming and
783 vegetation greening trends over pan-arctic tundra. *Remote Sens.* 5, 4229–4254.
784 <https://doi.org/10.3390/rs5094229>

785 Bjorkman, A.D., García Criado, M., Myers-Smith, I.H., Ravolainen, V., Jónsdóttir, I.S.,
786 Westergaard, K.B., Lawler, J.P., Aronsson, M., Bennett, B., Gardfjell, H., Heiðmarsson,
787 S., Stewart, L., Normand, S., 2020. Status and trends in Arctic vegetation: Evidence
788 from experimental warming and long-term monitoring. *Ambio* 49, 678–692.
789 <https://doi.org/10.1007/s13280-019-01161-6>

790 Bjorkman, A.D., Myers-Smith, I.H., Elmendorf, S.C., Normand, S., Rüger, N., Beck, P.S.A.,
791 Blach-Overgaard, A., Blok, D., Cornelissen, J.H.C., Forbes, B.C., Georges, D., Goetz,
792 S.J., Guay, K.C., Henry, G.H.R., HilleRisLambers, J., Hollister, R.D., Karger, D.N.,
793 Kattge, J., Manning, P., Prevéy, J.S., Rixen, C., Schaepman-Strub, G., Thomas, H.J.D.,
794 Vellend, M., Wilmking, M., Wipf, S., Carbognani, M., Hermanutz, L., Lévesque, E.,
795 Molau, U., Petraglia, A., Soudzilovskaia, N.A., Spasojevic, M.J., Tomaselli, M.,
796 Vowles, T., Alatalo, J.M., Alexander, H.D., Anadon-Rosell, A., Angers-Blondin, S.,
797 Beest, M. te, Berner, L., Björk, R.G., Buchwal, A., Buras, A., Christie, K., Cooper, E.J.,
798 Dullinger, S., Elberling, B., Eskelinen, A., Frei, E.R., Grau, O., Grogan, P., Hallinger,
799 M., Harper, K.A., Heijmans, M.M.P.D., Hudson, J., Hülber, K., Iturrate-Garcia, M.,
800 Iversen, C.M., Jaroszynska, F., Johnstone, J.F., Jørgensen, R.H., Kaarlejärvi, E., Klady,

801 R., Kuleza, S., Kulonen, A., Lamarque, L.J., Lantz, T., Little, C.J., Speed, J.D.M.,
802 Michelsen, A., Milbau, A., Nabe-Nielsen, J., Nielsen, S.S., Ninot, J.M., Oberbauer, S.F.,
803 Olofsson, J., Onipchenko, V.G., Rumpf, S.B., Semenchuk, P., Shetti, R., Collier, L.S.,
804 Street, L.E., Suding, K.N., Tape, K.D., Trant, A., Treier, U.A., Tremblay, J.P.,
805 Tremblay, M., Venn, S., Weijers, S., Zamin, T., Boulanger-Lapointe, N., Gould, W.A.,
806 Hik, D.S., Hofgaard, A., Jónsdóttir, I.S., Jorgenson, J., Klein, J., Magnusson, B.,
807 Tweedie, C., Wookey, P.A., Bahn, M., Blonder, B., van Bodegom, P.M., Bond-
808 Lamberty, B., Campetella, G., Cerabolini, B.E.L., Chapin, F.S., Cornwell, W.K., Craine,
809 J., Dainese, M., de Vries, F.T., Díaz, S., Enquist, B.J., Green, W., Milla, R., Niinemets,
810 Ü., Onoda, Y., Ordoñez, J.C., Ozinga, W.A., Penuelas, J., Poorter, H., Poschlod, P.,
811 Reich, P.B., Sandel, B., Schamp, B., Sheremetev, S., Weiher, E., 2018. Plant functional
812 trait change across a warming tundra biome. *Nature* 562, 57–62.
813 <https://doi.org/10.1038/s41586-018-0563-7>

814 Blok, D., Schaepman-Strub, G., Bartholomeus, H., Heijmans, M.M.P.D., Maximov, T.C.,
815 Berendse, F., 2011. The response of Arctic vegetation to the summer climate: Relation
816 between shrub cover, NDVI, surface albedo and temperature. *Environ. Res. Lett.* 6,
817 035502. <https://doi.org/10.1088/1748-9326/6/3/035502>

818 Boelman, N.T., Holbrook, J.D., Greaves, H.E., Krause, J.S., Chmura, H.E., Magney, T.S.,
819 Perez, J.H., Eitel, J.U.H., Gough, L., Vierling, K.T., Wingfield, J.C., Vierling, L.A.,
820 2016. Airborne laser scanning and spectral remote sensing give a bird's eye perspective
821 on arctic tundra breeding habitat at multiple spatial scales. *Remote Sens. Environ.* 184,
822 337–349. <https://doi.org/10.1016/j.rse.2016.07.012>

823 Bonfils, C.J.W., Phillips, T.J., Lawrence, D.M., Cameron-Smith, P., Riley, W.J., Subin, Z.M.,
824 2012. On the influence of shrub height and expansion on northern high latitude climate.

825 Environ. Res. Lett. 7, 015503. <https://doi.org/10.1088/1748-9326/7/1/015503>

826 Boulanger-Lapointe, N., Lévesque, E., Boudreau, S., Henry, G.H.R., Schmidt, N.M., 2014.
827 Population structure and dynamics of Arctic willow (*Salix arctica*) in the High Arctic. *J.*
828 *Biogeogr.* 41, 1967–1978. <https://doi.org/10.1111/jbi.12350>

829 Brodzik, M., Armstrong, R., 2013. Northern Hemisphere EASE-Grid 2.0 Weekly Snow
830 Cover and Sea Ice Extent. Version 4. Boulder, Color. USA NASA DAAC Natl. Snow
831 Ice Data Center.

832 Buchwal, A., Rachlewicz, G., Fonti, P., Cherubini, P., Gärtner, H., 2013. Temperature
833 modulates intra-plant growth of *Salix polaris* from a high Arctic site (Svalbard). *Polar*
834 *Biol.* 36, 1305–1318. <https://doi.org/10.1007/s00300-013-1349-x>

835 Büntgen, U., Bolze, N., Hellmann, L., Sittler, B., Frauenberger, B., Piermattei, A.,
836 Kirilyanov, A., Schweingruber, F.H., Ludemann, T., Krusic, P.J., 2018. Long-term
837 recruitment dynamics of arctic dwarf shrub communities in coastal east Greenland.
838 *Dendrochronologia* 50, 70–80. <https://doi.org/10.1016/j.dendro.2018.05.005>

839 Büntgen, U., Hellmann, L., Tegel, W., Normand, S., Myers-Smith, I., Kirilyanov, A. V.,
840 Nievergelt, D., Schweingruber, F.H., 2015. Temperature-induced recruitment pulses of
841 Arctic dwarf shrub communities. *J. Ecol.* 103, 489–501. [https://doi.org/10.1111/1365-](https://doi.org/10.1111/1365-2745.12361)
842 [2745.12361](https://doi.org/10.1111/1365-2745.12361)

843 Buras, A., Lehejček, J., Michalová, Z., Morrissey, R.C., Svoboda, M., Wilmking, M., 2017.
844 Shrubs shed light on 20th century Greenland Ice Sheet melting. *Boreas* 46, 667–677.
845 <https://doi.org/10.1111/bor.12244>

846 Cahoon, S.M.P., Sullivan, P.F., Shaver, G.R., Welker, J.M., Post, E., 2012. Interactions
847 among shrub cover and the soil microclimate may determine future Arctic carbon

848 budgets. *Ecol. Lett.* 15, 1415–1422. <https://doi.org/10.1111/j.1461-0248.2012.01865.x>

849 Carrer, M., Brunetti, M., Castagneri, D., 2016. The imprint of extreme climate events in
850 century-long time series of wood anatomical traits in high-elevation conifers. *Front.*
851 *Plant Sci.* 7, 1–12. <https://doi.org/10.3389/fpls.2016.00683>

852 Carrer, M., Von Arx, G., Castagneri, D., Petit, G., 2014. Distilling allometric and
853 environmental information from time series of conduit size: The standardization issue
854 and its relationship to tree hydraulic architecture. *Tree Physiol.* 35, 27–33.
855 <https://doi.org/10.1093/treephys/tpu108>

856 Cooley, S.W., Ryan, J.C., Smith, L.C., Horvat, C., Pearson, B., Dale, B., Lynch, A.H., 2020.
857 Coldest Canadian Arctic communities face greatest reductions in shorefast sea ice. *Nat.*
858 *Clim. Chang.* 10, 533–538. <https://doi.org/10.1038/s41558-020-0757-5>

859 Cunliffe, A.M., J Assmann, J., N Daskalova, G., Kerby, J.T., Myers-Smith, I.H., 2020.
860 Aboveground biomass corresponds strongly with drone-derived canopy height but
861 weakly with greenness (NDVI) in a shrub tundra landscape. *Environ. Res. Lett.* 15, 0–
862 31. <https://doi.org/10.1088/1748-9326/aba470>

863 Demarco, J., MacK, M.C., Bret-Harte, M.S., Burton, M., Shaver, G.R., 2014. Long-term
864 experimental warming and nutrient additions increase productivity in tall deciduous
865 shrub tundra. *Ecosphere* 5, art72. <https://doi.org/10.1890/ES13-00281.1>

866 Díaz, S., Hodgson, J.G., Thompson, K., Cabido, M., Cornelissen, J.H.C., Jalili, A.,
867 Montserrat-Martí, G., Grime, J.P., Zarrinkamar, F., Asri, Y., Band, S.R., Basconcelo, S.,
868 Castro-Díez, P., Funes, G., Hamzehee, B., Khoshnevi, M., Pérez-Harguindeguy, N.,
869 Pérez-Rontomé, M.C., Shirvany, F.A., Vendramini, F., Yazdani, S., Abbas-Azimi, R.,
870 Bogaard, A., Boustani, S., Charles, M., Dehghan, M., De Torres-Espuny, L., Falczuk,
871 V., Guerrero-Campo, J., Hynd, A., Jones, G., Kowsary, E., Kazemi-Saeed, F., Maestro-

872 Martínez, M., Romo-Díez, A., Shaw, S., Siavash, B., Villar-Salvador, P., Zak, M.R.,
873 2004. The plant traits that drive ecosystems: Evidence from three continents. *J. Veg. Sci.*
874 15, 295–304. <https://doi.org/10.1111/j.1654-1103.2004.tb02266.x>

875 Díaz, S., Lavorel, S., De Bello, F., Quétier, F., Grigulis, K., Robson, T.M., 2007.
876 Incorporating plant functional diversity effects in ecosystem service assessments. *Proc.*
877 *Natl. Acad. Sci. U. S. A.* 104, 20684–20689. <https://doi.org/10.1073/pnas.0704716104>

878 Dobbert, S., Pape, R., Löffler, J., 2021a. Contrasting growth response of evergreen and
879 deciduous arctic-alpine shrub species to climate variability. *Ecosphere* 12.
880 <https://doi.org/10.1002/ecs2.3688>

881 Dobbert, S., Pape, R., Löffler, J., 2021b. Contrasting growth response of evergreen and
882 deciduous arctic-alpine shrub species to climate variability. *Ecosphere* 12.
883 <https://doi.org/10.1002/ecs2.3688>

884 Dobbert, S., Pape, R., Löffler, J., 2021c. How does spatial heterogeneity affect inter- and
885 intraspecific growth patterns in tundra shrubs? *J. Ecol.* 0–2.
886 <https://doi.org/10.1111/1365-2745.13784>

887 Edelfeldt, S., Lindell, T., Dahlgren, J.P., 2019. Age-Independent Adult Mortality in a Long-
888 Lived Herb. *Diversity* 11, 1–14.

889 Elmendorf, S.C., Henry, G.H.R., Hollister, R.D., Björk, R.G., Bjorkman, A.D., Callaghan, T.
890 V., Collier, L.S., Cooper, E.J., Cornelissen, J.H.C., Day, T.A., Fosaa, A.M., Gould,
891 W.A., Grétarsdóttir, J., Harte, J., Hermanutz, L., Hik, D.S., Hofgaard, A., Jarrad, F.,
892 Jónsdóttir, I.S., Keuper, F., Klanderud, K., Klein, J.A., Koh, S., Kudo, G., Lang, S.I.,
893 Loewen, V., May, J.L., Mercado, J., Michelsen, A., Molau, U., Myers-Smith, I.H.,
894 Oberbauer, S.F., Pieper, S., Post, E., Rixen, C., Robinson, C.H., Schmidt, N.M., Shaver,
895 G.R., Stenström, A., Tolvanen, A., Totland, Ø., Troxler, T., Wahren, C.H., Webber, P.J.,

896 Welker, J.M., Wookey, P.A., 2012. Global assessment of experimental climate warming
897 on tundra vegetation: Heterogeneity over space and time. *Ecol. Lett.* 15, 164–175.
898 <https://doi.org/10.1111/j.1461-0248.2011.01716.x>

899 Fonti, P., Jansen, S., 2012. Xylem plasticity in response to climate. *New Phytol.* 195, 734–
900 736. <https://doi.org/10.1111/j.1469-8137.2012.04252.x>

901 Fonti, P., von Arx, G., García-González, I., Eilmann, B., Sass-Klaassen, U., Gärtner, H.,
902 Eckstein, D., 2010. Studying global change through investigation of the plastic
903 responses of xylem anatomy in tree rings. *New Phytol.* 185, 42–53.
904 <https://doi.org/10.1111/j.1469-8137.2009.03030.x>

905 Forbes, B.C., Kumpula, T., Meschyb, N., Laptander, R., MacIas-Fauria, M., Zetterberg, P.,
906 Verdonen, M., Skarin, A., Kim, K.Y., Boisvert, L.N., Stroeve, J.C., Bartsch, A., 2016.
907 Sea ice, rain-on-snow and tundra reindeer nomadism in Arctic Russia. *Biol. Lett.* 12, 4–
908 8. <https://doi.org/10.1098/rsbl.2016.0466>

909 Fraser, R.H., Olthof, I., Lantz, T.C., Schmitt, C., 2016. UAV photogrammetry for mapping
910 vegetation in the low-Arctic. *Arct. Sci.* 2, 79–102. <https://doi.org/10.1139/as-2016-0008>

911 Fritts, H.C., 1976. *Tree Rings and Climate*.

912 Fritts, H.C., 1971. *Dendroclimatology and dendroecology*, Quaternary Research. Cambridge
913 University Press. [https://doi.org/10.1016/0033-5894\(71\)90057-3](https://doi.org/10.1016/0033-5894(71)90057-3)

914 Fritts, H.C., Swetnam, T.W., 1989. *Dendroecology: A Tool for Evaluating Variations in Past*
915 *and Present Forest Environments*, in: M. Begon, A.H. Fitter, E.D. Ford, A.M. (Ed.),
916 *Advances in Ecological Research*. pp. 111–188. [https://doi.org/10.1016/S0065-](https://doi.org/10.1016/S0065-2504(08)60158-0)
917 [2504\(08\)60158-0](https://doi.org/10.1016/S0065-2504(08)60158-0)

918 Gamm, C.M., Sullivan, P.F., Buchwal, A., Dial, R.J., Young, A.B., Watts, D.A., Cahoon,

919 S.M.P., Welker, J.M., Post, E., 2018. Declining growth of deciduous shrubs in the
920 warming climate of continental western Greenland. *J. Ecol.* 106, 640–654.
921 <https://doi.org/10.1111/1365-2745.12882>

922 Gärtner, H., Cherubini, P., Fonti, P., von Arx, G., Schneider, L., Nievergelt, D., Verstege, A.,
923 Bast, A., Schweingruber, F.H., Büntgen, U., 2015. A technical perspective in modern
924 tree-ring research - How to overcome dendroecological and wood anatomical
925 challenges. *J. Vis. Exp.* 2015, 1–10. <https://doi.org/10.3791/52337>

926 Gärtner, H., Schweingruber, F., 2013. Microscopic preparation techniques for plant stem
927 analysis. Verlag Dr. Kessel, Remagen-Oberwinter. Kessel 95, 78.

928 Gough, L., Moore, J.C., Shaver, G.R., Simpson, R.T., Johnson, D.R., 2014. Erratum: Above-
929 and belowground responses of arctic tundra ecosystems to altered soil nutrients and
930 mammalian herbivory (*Ecology* (2012) 93 (1683-1694)). *Ecology* 95, 2028.
931 <https://doi.org/10.1890/0012-9658-95.7.2028>

932 Grenfell, T.C., Putkonen, J., 2008. A method for the detection of the severe rain-on-snow
933 event on Banks Island, October 2003, using passive microwave remote sensing. *Water*
934 *Resour. Res.* 44, 1–9. <https://doi.org/10.1029/2007WR005929>

935 Guay, K.C., Beck, P.S.A., Berner, L.T., Goetz, S.J., Baccini, A., Buermann, W., 2014.
936 Vegetation productivity patterns at high northern latitudes: A multi-sensor satellite data
937 assessment. *Glob. Chang. Biol.* 20, 3147–3158. <https://doi.org/10.1111/gcb.12647>

938 Hampe, A., Petit, R.J., 2005. Conserving biodiversity under climate change: The rear edge
939 matters. *Ecol. Lett.* 8, 461–467. <https://doi.org/10.1111/j.1461-0248.2005.00739.x>

940 Howat, I.M., Negrete, A., Smith, B.E., 2014. The Greenland Ice Mapping Project (GIMP)
941 land classification and surface elevation data sets. *Cryosphere* 8, 1509–1518.

942 <https://doi.org/10.5194/tc-8-1509-2014>

943 Huang, W., Fonti, P., Lundqvist, S.O., Larsen, J.B., Hansen, J.K., Thygesen, L.G., 2021.
944 Differences in xylem response to drought provide hints to future species selection. *New*
945 *For.* <https://doi.org/10.1007/s11056-021-09885-8>

946 IPCC Working Group II, 2022. *Climate Change 2022: Impacts, Adaptation, and*
947 *Vulnerability. Contribution of Working Group II to the Sixth Assessment Report of the*
948 *Intergovernmental Panel on Climate Change.*

949 Iturrate-Garcia, M., Heijmans, M., Cornelissen, J.H.C., Schweingruber, F., Niklaus, P.,
950 Schaepman-Strub, G., 2020. Plant trait response of tundra shrubs to permafrost thaw and
951 nutrient addition. *Biogeosciences Discuss.* 1–29. <https://doi.org/10.5194/bg-2019-498>

952 Iversen, C.M., Sloan, V.L., Sullivan, P.F., Euskirchen, E.S., McGuire, A.D., Norby, R.J.,
953 Walker, A.P., Warren, J.M., Wullschleger, S.D., 2015. The unseen iceberg: Plant roots
954 in arctic tundra. *New Phytol.* 205, 34–58. <https://doi.org/10.1111/nph.13003>

955 Johnstone, J.F., Henry, G.H.R., 1997. Retrospective analysis of growth and reproduction in
956 *Cassiope tetragona* and relations to climate in the Canadian high arctic. *Arct. Alp. Res.*
957 29, 459–469. <https://doi.org/10.2307/1551993>

958 Karger, D.N., Conrad, O., Böhner, J., Kawohl, T., Kreft, H., Soria-Auza, R.W.,
959 Zimmermann, N.E., Linder, H.P., Kessler, M., 2017. Climatologies at high resolution
960 for the earth’s land surface areas. *Sci. Data* 4, 1–20.
961 <https://doi.org/10.1038/sdata.2017.122>

962 Kašpar, J., Tumajer, J., Treml, V., 2019. IncrementR: Analysing height growth of trees and
963 shrubs in R. *Dendrochronologia* 53, 48–54. <https://doi.org/10.1016/j.dendro.2018.11.001>

964 Kempainen, J., Niittynen, P., Virkkala, A.M., Happonen, K., Riihimäki, H., Aalto, J., Luoto,

965 M., 2021. Dwarf Shrubs Impact Tundra Soils: Drier, Colder, and Less Organic Carbon.
966 *Ecosystems* 24, 1378–1392. <https://doi.org/10.1007/s10021-020-00589-2>

967 Klesse, S., DeRose, R.J., Guiterman, C.H., Lynch, A.M., O'Connor, C.D., Shaw, J.D., Evans,
968 M.E.K., 2018. Sampling bias overestimates climate change impacts on forest growth in
969 the southwestern United States. *Nat. Commun.* 9, 1–9. [https://doi.org/10.1038/s41467-](https://doi.org/10.1038/s41467-018-07800-y)
970 [018-07800-y](https://doi.org/10.1038/s41467-018-07800-y)

971 Kolishchuk, V., 1990. Dendroclimatological study of prostrate woody plants, in: Cook, E.R.,
972 Kairiukstis, L.A. (Eds.), *Methods of Dendrochronology: Applications in the*
973 *Environmental Sciences*. Kluwer, Dordrecht, Boston, London, pp. 51–55.

974 Körner, C., Hiltbrunner, E., 2018. The 90 ways to describe plant temperature. *Perspect. Plant*
975 *Ecol. Evol. Syst.* 30, 16–21. <https://doi.org/10.1016/j.ppees.2017.04.004>

976 Kunstler, G., Falster, D., Coomes, D.A., Hui, F., Kooyman, R.M., Laughlin, D.C., Poorter,
977 L., Vanderwel, M., Vieilledent, G., Wright, S.J., Aiba, M., Baraloto, C., Caspersen, J.,
978 Cornelissen, J.H.C., Gourlet-Fleury, S., Hanewinkel, M., Herault, B., Kattge, J.,
979 Kurokawa, H., Onoda, Y., Peñuelas, J., Poorter, H., Uriarte, M., Richardson, S., Ruiz-
980 Benito, P., Sun, I.F., Ståhl, G., Swenson, N.G., Thompson, J., Westerlund, B., Wirth, C.,
981 Zavala, M.A., Zeng, H., Zimmerman, J.K., Zimmermann, N.E., Westoby, M., 2016.
982 Plant functional traits have globally consistent effects on competition. *Nature* 529, 204–
983 207. <https://doi.org/10.1038/nature16476>

984 Kwok, R., 2018. Ecology's remote-sensing revolution. *Nature* 556, 137–138.
985 <https://doi.org/10.1038/d41586-018-03924-9>

986 Lantz, T.C., Gergel, S.E., Henry, G.H.R., 2010. Response of green alder (*Alnus viridis* subsp.
987 *fruticosa*) patch dynamics and plant community composition to fire and regional
988 temperature in north-western Canada. *J. Biogeogr.* 37, 1597–1610.

989 <https://doi.org/10.1111/j.1365-2699.2010.02317.x>

990 Lara, M.J., Nitze, I., Grosse, G., Martin, P., David McGuire, A., 2018. Reduced arctic tundra
991 productivity linked with landform and climate change interactions. *Sci. Rep.* 8, 1–10.
992 <https://doi.org/10.1038/s41598-018-20692-8>

993 Larson, E.R., Allen, S., Flinner, N.L., Labarge, S.G., Wilding, T.C., 2013. The need and
994 means to update chronologies in a dynamic environment. *Tree-Ring Res.* 69, 21–27.
995 <https://doi.org/10.3959/1536-1098-69.1.21>

996 Lawrence, D.M., Swenson, S.C., 2011. Permafrost response to increasing Arctic shrub
997 abundance depends on the relative influence of shrubs on local soil cooling versus large-
998 scale climate warming. *Environ. Res. Lett.* 6, 045504. [https://doi.org/10.1088/1748-
999 9326/6/4/045504](https://doi.org/10.1088/1748-9326/6/4/045504)

1000 Le Moullec, M., Sandal, L., Grøtan, V., Buchwal, A., Hansen, B.B., 2020. Climate
1001 synchronises shrub growth across a high-arctic archipelago: contrasting implications of
1002 summer and winter warming. *Oikos* 129, 1012–1027. <https://doi.org/10.1111/oik.07059>

1003 Lechthaler, S., Turnbull, T.L., Gelmini, Y., Pirotti, F., Anfodillo, T., Adams, M.A., Petit, G.,
1004 2019. A standardization method to disentangle environmental information from axial
1005 trends of xylem anatomical traits. *Tree Physiol.* 39, 495–502.
1006 <https://doi.org/10.1093/treephys/tpy110>

1007 Lembrechts, J.J., Aalto, J., Ashcroft, M.B., Frenne, P. De, Kopecký, M., Lenoir, J., Luoto,
1008 M., Maclean, I.M.D., 2020. SoilTemp : A global database of near-surface temperature.
1009 *Glob. Chang. Biol.* 6616–6629. <https://doi.org/10.1111/gcb.15123>

1010 Lenz, A., Hoch, G., Körner, C., 2013. Early season temperature controls cambial activity and
1011 total tree ring width at the alpine treeline. *Plant Ecol. Divers.* 6, 365–375.

1012 <https://doi.org/10.1080/17550874.2012.711864>

1013 Lesica, P., Crone, E.E., 2017. Arctic and boreal plant species decline at their southern range
1014 limits in the Rocky Mountains. *Ecol. Lett.* 20, 166–174.
1015 <https://doi.org/10.1111/ele.12718>

1016 Loranty, M.M., Goetz, S.J., Beck, P.S.A., 2011. Erratum: Tundra vegetation effects on pan-
1017 Arctic albedo (*Environmental Research Letters* (2011) 6 (024014)). *Environ. Res. Lett.*
1018 6, 024014. <https://doi.org/10.1088/1748-9326/6/2/024014>

1019 Macias-Fauria, M., Forbes, B.C., Zetterberg, P., Kumpula, T., 2012. Eurasian Arctic greening
1020 reveals teleconnections and the potential for structurally novel ecosystems. *Nat. Clim.*
1021 *Chang.* 2, 613–618. <https://doi.org/10.1038/nclimate1558>

1022 Maclean, I.M.D., Mosedale, J.R., Bennie, J.J., 2018. Microclima: An R package for
1023 modelling meso- and microclimate. *Methods Ecol. Evol.* 10, 280–290.
1024 <https://doi.org/10.1111/2041-210X.13093>

1025 Manzanedo, R.D., Pederson, N., 2019. Towards a More Ecological Dendroecology. *Tree-*
1026 *Ring Res.* 75, 152–159. <https://doi.org/10.3959/1536-1098-75.2.152>

1027 Martin, A.C., Jeffers, E.S., Petrokofsky, G., Myers-Smith, I., Macias-Fauria, M., 2017. Shrub
1028 growth and expansion in the Arctic tundra: An assessment of controlling factors using
1029 an evidence-based approach. *Environ. Res. Lett.* 12, 085007.
1030 <https://doi.org/10.1088/1748-9326/aa7989>

1031 McMahon, S.M., Harrison, S.P., Armbruster, W.S., Bartlein, P.J., Beale, C.M., Edwards,
1032 M.E., Kattge, J., Midgley, G., Morin, X., Prentice, I.C., 2011. Improving assessment and
1033 modelling of climate change impacts on global terrestrial biodiversity. *Trends Ecol.*
1034 *Evol.* 26, 249–259. <https://doi.org/10.1016/j.tree.2011.02.012>

1035 Mekonnen, Z.A., Riley, W.J., Berner, L.T., Bouskill, N.J., Torn, M.S., Iwahana, G., Breen,
1036 A.L., Myers-Smith, I.H., Criado, M.G., Liu, Y., Euskirchen, E.S., Goetz, S.J., Mack,
1037 M.C., Grant, R.F., 2021. Arctic tundra shrubification: a review of mechanisms and
1038 impacts on ecosystem carbon balance. *Environ. Res. Lett.* 16, 053001.
1039 <https://doi.org/10.1088/1748-9326/abf28b>

1040 Metcalfe, D.B., Hermans, T.D.G., Ahlstrand, J., Becker, M., Berggren, M., Björk, R.G.,
1041 Björkman, M.P., Blok, D., Chaudhary, N., Chisholm, C., Classen, A.T., Hasselquist,
1042 N.J., Jonsson, M., Kristensen, J.A., Kumordzi, B.B., Lee, H., Mayor, J.R., Prevéy, J.,
1043 Pantazatou, K., Rousk, J., Sponseller, R.A., Sundqvist, M.K., Tang, J., Uddling, J.,
1044 Wallin, G., Zhang, W., Ahlström, A., Tenenbaum, D.E., Abdi, A.M., 2018. Patchy field
1045 sampling biases understanding of climate change impacts across the Arctic. *Nat. Ecol.*
1046 *Evol.* 2, 1443–1448. <https://doi.org/10.1038/s41559-018-0612-5>

1047 Mokany, K., Raison, R.J., Prokushkin, A.S., 2006. Critical analysis of root: Shoot ratios in
1048 terrestrial biomes. *Glob. Chang. Biol.* 12, 84–96. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2486.2005.001043.x)
1049 [2486.2005.001043.x](https://doi.org/10.1111/j.1365-2486.2005.001043.x)

1050 Myburg, A.A., Lev-Yadun, S., Sederoff, R.R., 2013. Xylem Structure and Function, in: *ELS*.
1051 Wiley. <https://doi.org/10.1002/9780470015902.a0001302.pub2>

1052 Myers-Smith, I.H., Elmendorf, S.C., Beck, P.S.A., Wilmking, M., Hallinger, M., Blok, D.,
1053 Tape, K.D., Rayback, S.A., Macias-Fauria, M., Forbes, B.C., Speed, J.D.M., Boulanger-
1054 Lapointe, N., Rixen, C., Lévesque, E., Schmidt, N.M., Baittinger, C., Trant, A.J.,
1055 Hermanutz, L., Collier, L.S., Dawes, M.A., Lantz, T.C., Weijers, S., Jørgensen, R.H.,
1056 Buchwal, A., Buras, A., Naito, A.T., Ravolainen, V., Schaepman-Strub, G., Wheeler,
1057 J.A., Wipf, S., Guay, K.C., Hik, D.S., Vellend, M., 2015a. Climate sensitivity of shrub
1058 growth across the tundra biome. *Nat. Clim. Chang.* 5, 887–891.

1059 <https://doi.org/10.1038/nclimate2697>

1060 Myers-Smith, I.H., Forbes, B.C., Wilmking, M., Hallinger, M., Lantz, T., Blok, D., Tape,
1061 K.D., Macias-Fauria, M., Sass-Klaassen, U., Lévesque, E., Boudreau, S., Ropars, P.,
1062 Hermanutz, L., Trant, A., Collier, L.S., Weijers, S., Rozema, J., Rayback, S.A., Schmidt,
1063 N.M., Schaepman-Strub, G., Wipf, S., Rixen, C., Ménard, C.B., Venn, S., Goetz, S.,
1064 Andreu-Hayles, L., Elmendorf, S., Ravolainen, V., Welker, J., Grogan, P., Epstein, H.E.,
1065 Hik, D.S., 2011. Shrub expansion in tundra ecosystems: Dynamics, impacts and research
1066 priorities. *Environ. Res. Lett.* 6, 045509. <https://doi.org/10.1088/1748-9326/6/4/045509>

1067 Myers-Smith, I.H., Grabowski, M.M., Thomas, H.J.D., Angers-Blondin, S., Daskalova, G.N.,
1068 Bjorkman, A.D., Cunliffe, A.M., Assmann, J.J., Boyle, J.S., McLeod, E., McLeod, S.,
1069 Joe, R., Lennie, P., Arey, D., Gordon, R.R., Eckert, C.D., 2019a. Eighteen years of
1070 ecological monitoring reveals multiple lines of evidence for tundra vegetation change.
1071 *Ecol. Monogr.* 89. <https://doi.org/10.1002/ecm.1351>

1072 Myers-Smith, I.H., Hallinger, M., Blok, D., Sass-Klaassen, U., Rayback, S.A., Weijers, S.,
1073 Trant, A.J., Tape, K.D., Naito, A.T., Wipf, S., Rixen, C., Dawes, M.A., Wheeler, J.A.,
1074 Buchwal, A., Baittinger, C., Macias-Fauria, M., Forbes, B.C., Lévesque, E., Boulanger-
1075 Lapointe, N., Beil, I., Ravolainen, V., Wilmking, M., 2015b. Methods for measuring
1076 arctic and alpine shrub growth: A review. *Earth-Science Rev.*
1077 <https://doi.org/10.1016/j.earscirev.2014.10.004>

1078 Myers-Smith, I.H., Hik, D.S., 2018. Climate warming as a driver of tundra shrubline
1079 advance. *J. Ecol.* 106, 547–560. <https://doi.org/10.1111/1365-2745.12817>

1080 Myers-Smith, I.H., Kerby, J.T., Phoenix, G.K., Bjerke, J.W., Epstein, H.E., Assmann, J.J.,
1081 John, C., Andreu-Hayles, L., Angers-Blondin, S., Beck, P.S.A., Berner, L.T., Bhatt,
1082 U.S., Bjorkman, A.D., Blok, D., Bryn, A., Christiansen, C.T., Cornelissen, J.H.C.,

1083 Cunliffe, A.M., Elmendorf, S.C., Forbes, B.C., Goetz, S.J., Hollister, R.D., de Jong, R.,
1084 Loranty, M.M., Macias-Fauria, M., Maseyk, K., Normand, S., Olofsson, J., Parker, T.C.,
1085 Parmentier, F.J.W., Post, E., Schaepman-Strub, G., Stordal, F., Sullivan, P.F., Thomas,
1086 H.J.D., Tømmervik, H., Treharne, R., Tweedie, C.E., Walker, D.A., Wilmking, M.,
1087 Wipf, S., 2020. Complexity revealed in the greening of the Arctic. *Nat. Clim. Chang.* 10,
1088 106–117. <https://doi.org/10.1038/s41558-019-0688-1>

1089 Myers-Smith, I.H., Thomas, H.J.D., Bjorkman, A.D., 2019b. Plant traits inform predictions
1090 of tundra responses to global change. *New Phytol.* 221, 1742–1748.
1091 <https://doi.org/10.1111/nph.15592>

1092 Nedbal, V., Láska, K., Brom, J., 2020. Mitigation of arctic tundra surface warming by plant
1093 evapotranspiration: Complete energy balance component estimation using LANDSAT
1094 satellite data. *Remote Sens.* 12, 1–16. <https://doi.org/10.3390/rs12203395>

1095 Nielsen, S.S., Arx, G. Von, Damgaard, C.F., Abermann, J., Buchwal, A., Büntgen, U., Treier,
1096 U.A., Barfod, A.S., Normand, S., 2017. Xylem Anatomical Trait Variability Provides
1097 Insight on the Climate-Growth Relationship of *Betula nana* in Western Greenland.
1098 *Arctic, Antarct. Alp. Res.* 49, 359–371. <https://doi.org/10.1657/AAAR0016-041>

1099 Normand, S., Zimmermann, N.E., Schurr, F.M., Lischke, H., 2014. Demography as the basis
1100 for understanding and predicting range dynamics. *Ecography (Cop.)*. 37, 1149–1154.
1101 <https://doi.org/10.1111/ecog.01490>

1102 Obu, J., Westermann, S., Käab, A., Bartsch, A., 2018. Ground Temperature Map, 2000-2016,
1103 Northern Hemisphere Permafrost. Alfred Wegener Institute, Helmholtz Cent. Polar Mar.
1104 Res. Bremerhaven. <https://doi.org/https://doi.org/10.1594/PANGAEA.888600>

1105 Olson, M.E., Anfodillo, T., Gleason, S.M., McCulloh, K.A., 2021. Tip-to-base xylem conduit
1106 widening as an adaptation: causes, consequences, and empirical priorities. *New Phytol.*

1107 229, 1877–1893. <https://doi.org/10.1111/nph.16961>

1108 Olson, M.E., Anfodillo, T., Rosell, J.A., Petit, G., Crivellaro, A., Isnard, S., León-Gómez, C.,
1109 Alvarado-Cárdenas, L.O., Castorena, M., 2014. Universal hydraulics of the flowering
1110 plants: Vessel diameter scales with stem length across angiosperm lineages, habits and
1111 climates. *Ecol. Lett.* 17, 988–997. <https://doi.org/10.1111/ele.12302>

1112 Olson, M.E., Rosell, J.A., 2013. Vessel diameter-stem diameter scaling across woody
1113 angiosperms and the ecological causes of xylem vessel diameter variation. *New Phytol.*
1114 197, 1204–1213. <https://doi.org/10.1111/nph.12097>

1115 Parmentier, F.-J.W., Nilsen, L., Tømmervik, H., Cooper, E.J., 2021. A distributed time-lapse
1116 camera network to track vegetation phenology with high temporal detail and at varying
1117 scales. *Earth Syst. Sci. Data* 13, 3593–3606. <https://doi.org/10.5194/essd-13-3593-2021>

1118 Pearl, J.K., Keck, J.R., Tintor, W., Siekacz, L., Herrick, H.M., Meko, M.D., Pearson, C.L.,
1119 2020. New frontiers in tree-ring research. *Holocene* 30, 923–941.
1120 <https://doi.org/10.1177/0959683620902230>

1121 Pearson, R.G., Phillips, S.J., Loranty, M.M., Beck, P.S.A., Damoulas, T., Knight, S.J., Goetz,
1122 S.J., 2013. Shifts in Arctic vegetation and associated feedbacks under climate change.
1123 *Nat. Clim. Chang.* 3, 673–677. <https://doi.org/10.1038/nclimate1858>

1124 Phoenix, G.K., Bjerke, J.W., 2016. Arctic browning: extreme events and trends reversing
1125 arctic greening. *Glob. Chang. Biol.* 22, 2960–2962. <https://doi.org/10.1111/gcb.13261>

1126 Pizano, C., Barón, A.F., Schuur, E.A.G., Crummer, K.G., Mack, M.C., 2014. Effects of
1127 thermo-erosional disturbance on surface soil carbon and nitrogen dynamics in upland
1128 arctic tundra. *Environ. Res. Lett.* 9, 075006. [https://doi.org/10.1088/1748-](https://doi.org/10.1088/1748-9326/9/7/075006)
1129 [9326/9/7/075006](https://doi.org/10.1088/1748-9326/9/7/075006)

1130 Porter, C., Morin, P., Howat, I., Noh, M.-J., Bates, B., Peterman, K., Keeseey, S., Schlenk, M.,
1131 Gardiner, J., Tomko, K., Willis, M., Kelleher, C., Cloutier, M., Husby, E., Foga, S.,
1132 Nakamura, H., Pl, M., 2018. ArcticDEM. Harvard Dataverse, V1.

1133 Post, E., Cahoon, S.M.P., Kerby, J.T., Pedersen, C., Sullivan, P.F., 2021. Herbivory and
1134 warming interact in opposing patterns of covariation between arctic shrub species at
1135 large and local scales. *Proc. Natl. Acad. Sci. U. S. A.* 118, 1–9.
1136 <https://doi.org/10.1073/pnas.2015158118>

1137 Post, E., Pedersen, C., 2008. Opposing plant community responses to warming with and
1138 without herbivores. *Proc. Natl. Acad. Sci. U. S. A.* 105, 12353–12358.
1139 <https://doi.org/10.1073/pnas.0802421105>

1140 Prendin, A.L., Carrer, M., Karami, M., Hollesen, J., Bjerregaard Pedersen, N., Pividori, M.,
1141 Treier, U.A., Westergaard-Nielsen, A., Elberling, B., Normand, S., 2020. Immediate and
1142 carry-over effects of insect outbreaks on vegetation growth in West Greenland assessed
1143 from cells to satellite. *J. Biogeogr.* 47, 87–100. <https://doi.org/10.1111/jbi.13644>

1144 Prendin, A.L., Mayr, S., Beikircher, B., von Arx, G., Petit, G., 2018a. Xylem anatomical
1145 adjustments prioritize hydraulic efficiency over safety as Norway spruce trees grow
1146 taller. *Tree Physiol.* 38, 1088–1097. <https://doi.org/10.1093/treephys/tpy065>

1147 Prendin, A.L., Petit, G., Fonti, P., Rixen, C., Dawes, M.A., von Arx, G., 2018b. Axial xylem
1148 architecture of *Larix decidua* exposed to CO₂ enrichment and soil warming at the tree
1149 line. *Funct. Ecol.* 32, 273–287. <https://doi.org/10.1111/1365-2435.12986>

1150 Rayback, S.A., Henry, G.H.R., 2005. Dendrochronological potential of the arctic dwarf-shrub
1151 *Cassiope tetragona*. *Tree-Ring Res.* 61, 43–53. [https://doi.org/10.3959/1536-1098-](https://doi.org/10.3959/1536-1098-61.1.43)
1152 [61.1.43](https://doi.org/10.3959/1536-1098-61.1.43)

1153 Reynolds, M.K., Walker, D.A., Balser, A., Bay, C., Campbell, M., Cherosov, M.M., Daniëls,
1154 F.J.A., Eidesen, P.B., Ermokhina, K.A., Frost, G. V., Jedrzejek, B., Jorgenson, M.T.,
1155 Kennedy, B.E., Kholod, S.S., Lavrinenko, I.A., Lavrinenko, O. V., Magnússon, B.,
1156 Matveyeva, N. V., Metúsalemsson, S., Nilsen, L., Olthof, I., Pospelov, I.N., Pospelova,
1157 E.B., Pouliot, D., Razzhivin, V., Schaepman-Strub, G., Šibík, J., Telyatnikov, M.Y.,
1158 Troeva, E., 2019. A raster version of the Circumpolar Arctic Vegetation Map (CAVM).
1159 Remote Sens. Environ. 232, 111297. <https://doi.org/10.1016/j.rse.2019.111297>

1160 Reichle, L.M., Epstein, H.E., Bhatt, U.S., Reynolds, M.K., Walker, D.A., 2018. Spatial
1161 Heterogeneity of the Temporal Dynamics of Arctic Tundra Vegetation. Geophys. Res.
1162 Lett. 45, 9206–9215. <https://doi.org/10.1029/2018GL078820>

1163 Resente, G., Gillert, A., Trouillier, M., Anadon-Rosell, A., Peters, R.L., von Arx, G., von
1164 Lukas, U., Wilmking, M., 2021. Mask, Train, Repeat! Artificial Intelligence for
1165 Quantitative Wood Anatomy. Front. Plant Sci. 12, 1–14.
1166 <https://doi.org/10.3389/fpls.2021.767400>

1167 Richardson, A.D., Hufkens, K., Milliman, T., Aubrecht, D.M., Chen, M., Gray, J.M.,
1168 Johnston, M.R., Keenan, T.F., Klosterman, S.T., Kosmala, M., Melaas, E.K., Friedl,
1169 M.A., Frohling, S., 2018. Tracking vegetation phenology across diverse North American
1170 biomes using PhenoCam imagery. Sci. Data 5, 1–24.
1171 <https://doi.org/10.1038/sdata.2018.28>

1172 Robinson, D.A., Estilow, T.W., Program, N.C., 2012. NOAA Climate Data Record (CDF) or
1173 Northern Hemisphere (NH) Snow Cover Extent (SCE), Version 1 (2000-01-01 to 2019-
1174 12-31). NOAA Natl. Centers Environ. Information. <https://doi.org/10.7289/V5N014G9>

1175 Ropars, P., Angers-Blondin, S., Gagnon, M., Myers-Smith, I.H., Lévesque, E., Boudreau, S.,
1176 2017. Different parts, different stories: climate sensitivity of growth is stronger in root

- 1177 collars vs. stems in tundra shrubs. *Glob. Chang. Biol.* 23, 3281–3291.
- 1178 <https://doi.org/10.1111/gcb.13631>
- 1179 Ropars, P., Lévesque, E., Boudreau, S., 2015. How do climate and topography influence the
1180 greening of the forest-tundra ecotone in northern Québec? A dendrochronological
1181 analysis of *Betula glandulosa*. *J. Ecol.* 103, 679–690. [https://doi.org/10.1111/1365-](https://doi.org/10.1111/1365-2745.12394)
1182 [2745.12394](https://doi.org/10.1111/1365-2745.12394)
- 1183 Rozema, J., Weijers, S., Broekman, R., Blokker, P., Buizer, B., Werleman, C., El Yaqine, H.,
1184 Hoogedoorn, H., Fuertes, M.M., Cooper, E., 2009. Annual growth of *Cassiope tetragona*
1185 as a proxy for Arctic climate: Developing correlative and experimental transfer
1186 functions to reconstruct past summer temperature on a millennial time scale. *Glob.*
1187 *Chang. Biol.* 15, 1703–1715. <https://doi.org/10.1111/j.1365-2486.2009.01858.x>
- 1188 Schmidt, N.M., Baittinger, C., Kollmann, J., Forchhammer, M.C., 2010. Consistent
1189 dendrochronological response of the dioecious *Salix arctica* to variation in local snow
1190 precipitation across gender and vegetation types. *Arctic, Antarct. Alp. Res.* 42, 471–475.
1191 <https://doi.org/10.1657/1938-4246-42.4.471>
- 1192 Schweingruber, F.H., 1996. *Tree Rings and Environment: Dendroecology*, P. Haupt. P.
1193 Haupt. <https://doi.org/10.2307/1224418>
- 1194 Schweingruber, F.H., Hellmann, L., Tegel, W., Braun, S., Nievergelt, D., Büntgen, U., 2013.
1195 Evaluating the wood anatomical and dendroecological potential of arctic dwarf shrub
1196 communities. *IAWA J.* 34, 485–497. <https://doi.org/10.1163/22941932-00000039>
- 1197 Schweingruber, F.H., Poschlod, P., 2005. *Growth Rings in Herbs and Shrubs: Life Span, Age*
1198 *Determination and Stem Anatomy*. *Forest Snow and Landscape Research*.
- 1199 Seddon, A.W.R., Macias-Fauria, M., Long, P.R., Benz, D., Willis, K.J., 2016. Sensitivity of

1200 global terrestrial ecosystems to climate variability. *Nature* 531, 229–232.
1201 <https://doi.org/10.1038/nature16986>

1202 Shetti, R., Buras, A., Smiljanic, M., Wilmking, M., 2018. Climate sensitivity is affected by
1203 growth differentiation along the length of *Juniperus communis* L. shrub stems in the
1204 Ural Mountains. *Dendrochronologia* 49, 29–35.
1205 <https://doi.org/10.1016/j.dendro.2018.02.006>

1206 Siewert, M.B., Olofsson, J., 2020. Scale-dependency of Arctic ecosystem properties revealed
1207 by UAV. *Environ. Res. Lett.* 15, 094030. <https://doi.org/10.1088/1748-9326/aba20b>

1208 Skarin, A., Verdonen, M., Kumpula, T., Maclas-Fauria, M., Alam, M., Kerby, J., Forbes,
1209 B.C., 2020. Reindeer use of low Arctic tundra correlates with landscape structure.
1210 *Environ. Res. Lett.* 15. <https://doi.org/10.1088/1748-9326/abf15>

1211 Sun, Q., Miao, C., Duan, Q., Ashouri, H., Sorooshian, S., Hsu, K.L., 2018. A Review of
1212 Global Precipitation Data Sets: Data Sources, Estimation, and Intercomparisons. *Rev.*
1213 *Geophys.* 56, 79–107. <https://doi.org/10.1002/2017RG000574>

1214 Tape, K., Sturm, M., Racine, C., 2006. The evidence for shrub expansion in Northern Alaska
1215 and the Pan-Arctic. *Glob. Chang. Biol.* 12, 686–702. <https://doi.org/10.1111/j.1365-2486.2006.01128.x>

1217 Virtanen, R., Oksanen, L., Oksanen, T., Cohen, J., Forbes, B.C., Johansen, B., Käyhkö, J.,
1218 Olofsson, J., Pulliainen, J., Tømmervik, H., 2016. Where do the treeless tundra areas of
1219 northern highlands fit in the global biome system: Toward an ecologically natural
1220 subdivision of the tundra biome. *Ecol. Evol.* 6, 143–158.
1221 <https://doi.org/10.1002/ece3.1837>

1222 von Arx, G., Carrer, M., 2014. Roxas -A new tool to build centuries-long tracheid-lumen

1223 chronologies in conifers. *Dendrochronologia* 32, 290–293.
1224 <https://doi.org/10.1016/j.dendro.2013.12.001>

1225 von Arx, G., Carrer, M., Crivellaro, A., De Micco, V., Fonti, P., Lens, F., Prendin, A.L.,
1226 Rosner, S., Sass-Klaassen, U., 2021. Q-NET – a new scholarly network on quantitative
1227 wood anatomy. *Dendrochronologia*. <https://doi.org/10.1016/j.dendro.2021.125890>

1228 von Arx, G., Crivellaro, A., Prendin, A.L., Čufar, K., Carrer, M., 2016. Quantitative wood
1229 anatomy—practical guidelines. *Front. Plant Sci.* 7, 1–13.
1230 <https://doi.org/10.3389/fpls.2016.00781>

1231 von Arx, G., Dietz, H., 2005. Automated image analysis of annual rings in the roots of
1232 perennial forbs. *Int. J. Plant Sci.* 166, 723–732. <https://doi.org/10.1086/431230>

1233 von Arx, G., Kueffer, C., Fonti, P., 2013. Quantifying plasticity in vessel grouping - Added
1234 value from the image analysis tool ROXAS. *IAWA J.* 34, 433–445.
1235 <https://doi.org/10.1163/22941932-00000035>

1236 Vowles, T., Björk, R.G., 2019. Implications of evergreen shrub expansion in the Arctic. *J.*
1237 *Ecol.* 107, 650–655. <https://doi.org/10.1111/1365-2745.13081>

1238 Walker, D.A., Reynolds, M.K., Daniëls, F.J.A., Einarsson, E., Elvebakk, A., Gould, W.A.,
1239 Katenin, A.E., Kholod, S.S., Markon, C.J., Melnikov, Evgeny S., Moskalenko, Natalia
1240 G., Talbot, S.S., Yurtsev, B.A., Bliss, L.C., Edlund, S.A., Zoltai, S.C., Wilhelm, M.,
1241 Bay, C., Gudjónsson, G., Moskalenko, N. G., Ananjeva, G. V., Drozdov, D.S.,
1242 Konchenko, L.A., Korostelev, Y. V., Melnikov, E. S., Ponomareva, O.E., Matveyeva, N.
1243 V., Safranova, I.N., Shelkunova, R., Polezhaev, A.N., Johansen, B.E., Maier, H.A.,
1244 Murray, D.F., Fleming, M.D., Trahan, N.G., Charron, T.M., Lauritzen, S.M., Vairin,
1245 B.A., 2005. The Circumpolar Arctic vegetation map. *J. Veg. Sci.* 16, 267–282.
1246 <https://doi.org/10.1111/j.1654-1103.2005.tb02365.x>

- 1247 Weijers, S., Myers-Smith, I.H., Löffler, J., 2018a. A Warmer and greener cold world:
1248 Summer warming increases shrub growth in the alpine and high arctic tundra. *Erdkunde*
1249 72, 63–85. <https://doi.org/10.3112/ERDKUNDE.2018.01.04>
- 1250 Weijers, S., Pape, R., Löffler, J., Myers-Smith, I.H., 2018b. Contrasting shrub species
1251 respond to early summer temperatures leading to correspondence of shrub growth
1252 patterns. *Environ. Res. Lett.* 13, 034005. <https://doi.org/10.1088/1748-9326/aaa5b8>
- 1253 Wild, J., Kopecký, M., Macek, M., Šanda, M., Jankovec, J., Haase, T., 2019. Climate at
1254 ecologically relevant scales: A new temperature and soil moisture logger for long-term
1255 microclimate measurement. *Agric. For. Meteorol.* 268, 40–47.
1256 <https://doi.org/10.1016/j.agrformet.2018.12.018>
- 1257 Wilmking, M., Buras, A., Lehejček, J., Lange, J., Shetti, R., van der Maaten, E., 2018.
1258 Influence of larval outbreaks on the climate reconstruction potential of an Arctic shrub.
1259 *Dendrochronologia* 49, 36–43. <https://doi.org/10.1016/j.dendro.2018.02.010>
- 1260 Wilmking, M., Hallinger, M., Van Bogaert, R., Kyncl, T., Babst, F., Hahne, W., Juday, G.P.,
1261 De Luis, M., Novak, K., Völlm, C., 2012. Continuously missing outer rings in woody
1262 plants at their distributional margins. *Dendrochronologia* 30, 213–222.
1263 <https://doi.org/10.1016/j.dendro.2011.10.001>
- 1264 Wilmking, M., van der Maaten-Theunissen, M., van der Maaten, E., Scharnweber, T., Buras,
1265 A., Biermann, C., Gurskaya, M., Hallinger, M., Lange, J., Shetti, R., Smiljanic, M.,
1266 Trouillier, M., 2020. Global assessment of relationships between climate and tree
1267 growth. *Glob. Chang. Biol.* 26, 3212–3220. <https://doi.org/10.1111/gcb.15057>
- 1268 Wilson, J.W., 1964. Annual growth of *Salix arctica* in the high-arctic. *Ann. Bot.* 28, 71–76.
1269 <https://doi.org/10.1093/oxfordjournals.aob.a083896>

1270 Woodcock, H., Bradley, R.S., 1994. *Salix Arctica* (PALL.): Its potential for
1271 dendroclimatological studies in the high Arctic. *Dendrochronologia* 11–22.

1272 Zeng, H., Jia, G., Epstein, H., 2011. Recent changes in phenology over the northern high
1273 latitudes detected from multi-satellite data. *Environ. Res. Lett.* 6.
1274 <https://doi.org/10.1088/1748-9326/6/4/045508>

1275

1276

Table 1: Drivers and driver categories identified in the literature search, potential remote sensing sources that could be used to quantify the drivers, and example datasets or studies that include such drivers. Left column: Overview of drivers we identified in the 82 publications included in our review of recent Arctic dendroecological studies. We placed each driver in one of nine categories (the categories are written in bold). Numbers inside parentheses represent how many studies included the driver or category at least once. Middle column: Remote sensing sources available as direct measures or proxies for the driver variables. Right column: Selected datasets or studies related to the driver and remote sensing source. The sources and examples list are not exhaustive, but a relevant selection based on our knowledge at the time of writing.

Driver (number of studies)	Remote Sensing Source	Example Dataset or Study
Air temperature (62)		
Summer Temperature (62)	Gridded climate reanalysis datasets including satellite data	ERA5 reanalysis of climate variables at 0.1° resolution
Winter Temperature (42)		CHELSA downscaled ERA5 data at 33 arcsec resolution (Karger et al., 2017)
Frost day frequency (2)		Also, see (Sun et al., 2018) for a recent review on gridded precipitation datasets.
Precipitation (45)		
Summer Precipitation (43)		
Winter Precipitation (37)		
Wet Day Frequency (2)		
Rain on Snow Events ^b (4)	Microwave remote sensing	(Forbes et al., 2016) (Grenfell and Putkonen, 2008)
Geophysical factors, processes, and disturbances (39)		
Habitat characteristics (18)	Multispectral Satellite imagery Airborne Laser Scanning	Global vegetation mapping (Raynolds et al., 2019) Fine-scale habitat mapping with airborne laser scanning (Boelman et al., 2016)
Soil nutrients (5)	-	-
Soil moisture (4)	Microwave remote sensing Optical imagery	Accurate models at coarse grain sizes (e.g. Copernicus Soil Moisture Product) Proxies at medium grain sizes: (Bartsch et al., 2020)
Standardized Precipitation-Evapotranspiration Index (SPEI) (5)	-	-
Ground temperature (2)	Infrared reflectance from multispectral imagery	Landsat (https://www.usgs.gov/core-science-systems/nli/landsat/landsat-collection-2-surface-temperature) or Copernicus Imagery (https://land.copernicus.eu/global/products/lst)
Soil temperature (1)	Models based on satellite data	ESA GlobPermafrost (Obu et al., 2018)
Potential Evapotranspiration (PET) (2)	Infrared reflectance from multispectral imagery	(Nedbal et al., 2020)
Road ^a (1)	-	-
Slope (1)	DEMs derived from satellite and airborne remote sensing, including laser scanning, stereographic imagery and radar data	Panarctic datasets like the Arctic DEM (Porter et al., 2018)
Solar radiation index (2)		National / Regional datasets like GIMP (Howat et al., 2014)

Thaw depth ^a (1)	-	-
Thaw ponds/ slumps ^b (2)	Fine grain optical imagery	IKONOS (Belshe et al., 2013) or Worldview imagery.
Tsunami ^b / Storm surge ^b (2)	-	-
Dew point (1)	-	-
Humidity (1)	-	-
Sea ice and ice caps (8)		
Sea ice (5)	Microwave (coarse grain) Optical imagery (fine grain)	NSIDC CDR Sea-ice concentration (https://nsidc.org/data/g02202) (Cooley et al., 2020)
Glacial retreat (1)	Various sources	See special issue in Remote Sensing (https://www.mdpi.com/journal/remotesensing/special_issues/remotesensing_glaciers)
Greenland Ice Sheet (1)	Passive microwave and various other sources	NSIDC Greenland Ice Sheet Melt (Abdalati, 2007)
Years since glaciation (1)	-	-
Biotic factors, interactions, and disturbances (20)		
Biotic site characteristics (14)	-	-
Caribou & Sheep herbivory ^{a b} (4)	-	-
Insect outbreak ^b (3)	Multispectral imagery	(Prendin et al., 2020)
Fire (2)	Multispectral imagery	MODIS Fire Product (https://modis.gsfc.nasa.gov/data/dataproduct/mod14.php)
Climate system (6)		
Arctic Oscillation (5)	Gridded climate reanalysis datasets including satellite data	ERA5 reanalysis at 0.1° resolution
North Atlantic Oscillation (3)		
Greenland Blocking Index (1)		
Scandinavian Pattern (1)		
Growing season (6)		
Growing season length and timing (6)	Multispectral imagery	MODIS and AVHRR imagery (Zeng et al., 2011)
Snow cover and depth (9)		
Snow cover / Snow depth ^a (9)	Microwave remote sensing Multispectral imagery	NSIDC CDR Snow Cover (Brodzik and Armstrong, 2013)(Robinson et al., 2012)
Other (4)		
Cloud cover (2)	Optical imagery	MODIS imagery (Seddon et al., 2016)
Sunlight duration (1)	DEMs derived from satellite and airborne remote sensing, including laser scanning, stereographic imagery and radar data	Global datasets like the Arctic DEM National / Regional datasets

^a variables we believe could be theoretically measured using remote sensing data, but further developments of the methods are likely required. ^b variables we identify as events with the remainder being trends.

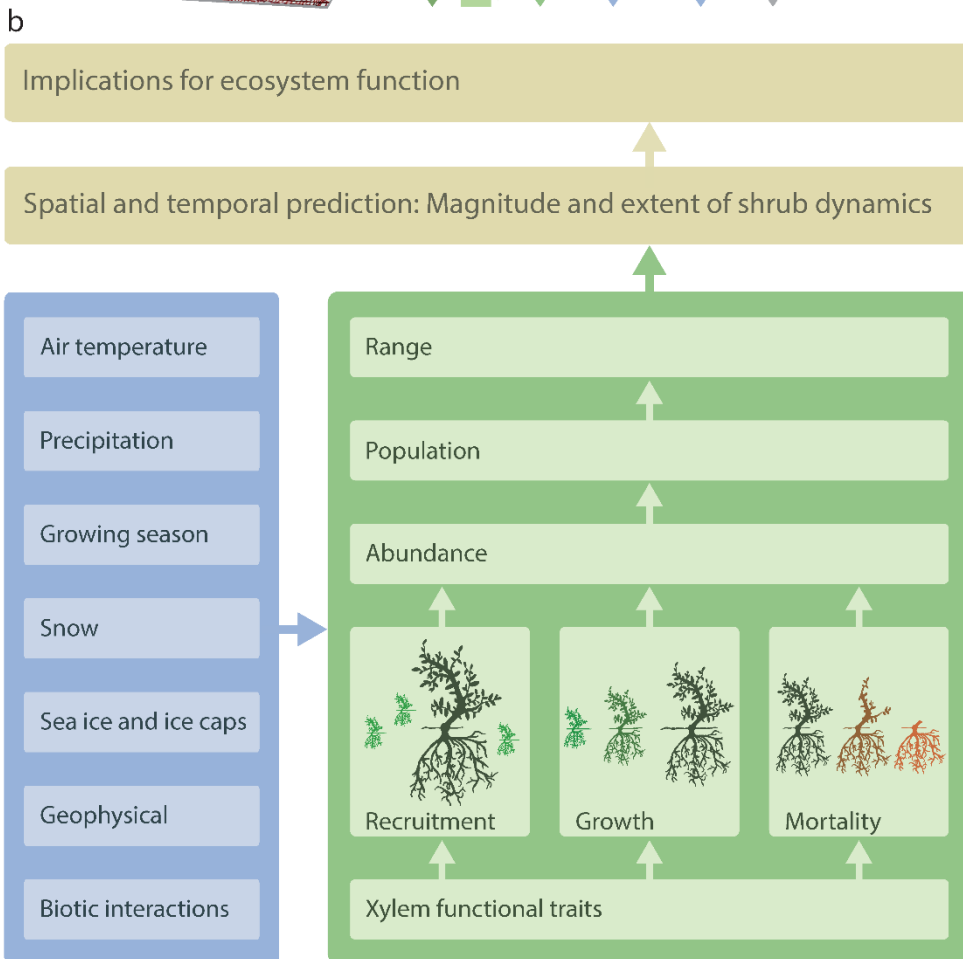
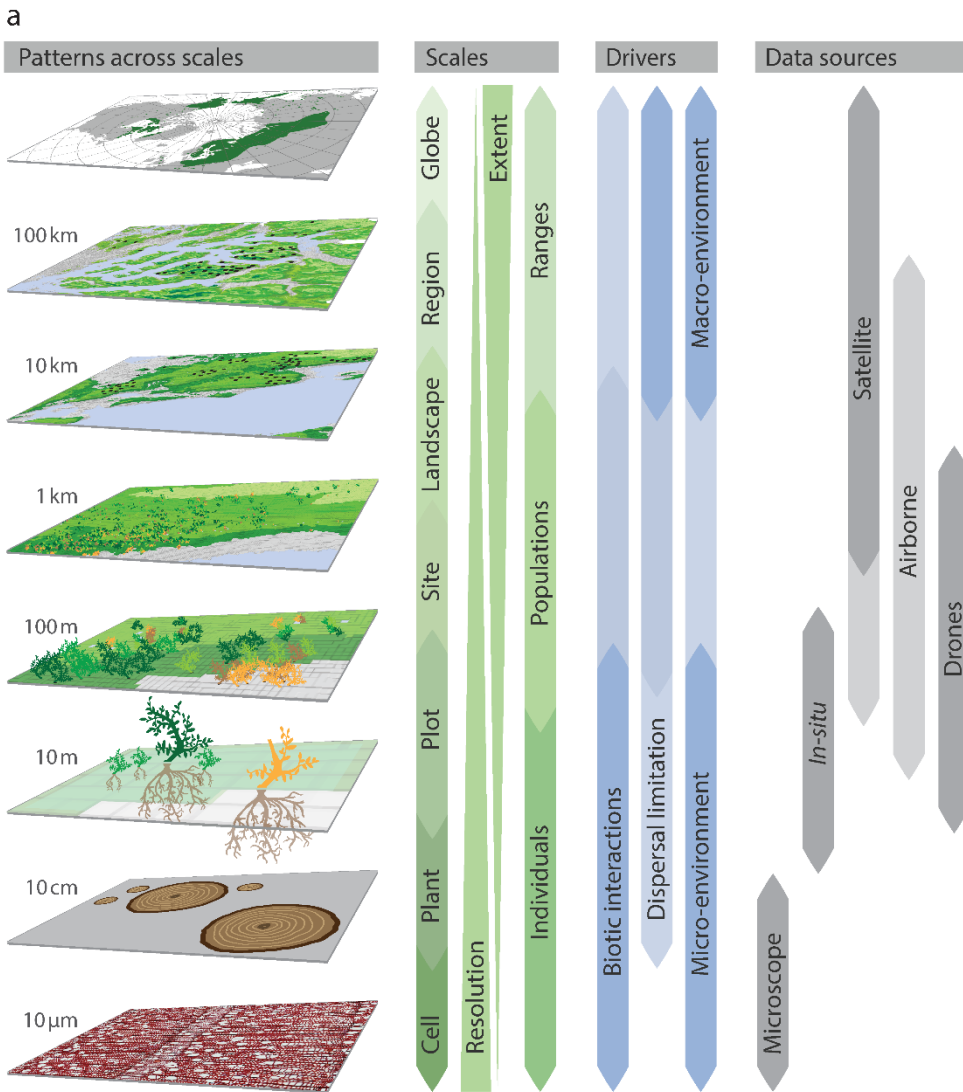


Figure 1: Linking shrub dynamics and their drivers across spatial and temporal scales. **(a)** Effects of environmental drivers are initiated at the level of individuals through their cellular and demographic responses (e.g., recruitment, growth, mortality). They propagate across populations to determine plant community and species range dynamics (green arrows). Key drivers (blue arrows) of local to macro-scale shrub dynamics (i.e., biotic interactions, dispersal limitation, and macro- to micro-environmental conditions) are expected to vary in importance across scales (different shadings). Combining cellular analyses; *in-situ* measurements; and satellite, airborne, and drone-based remote sensing contributes to linking these dynamics and drivers across spatial and temporal scales (grey arrows). **(b)** Interlinking information on multiple drivers (blue) and multiple responses (green) from local- to macroecological scales provides the basis for understanding why shrub responses differ across individuals, communities and species geographic ranges. Such understanding can allow for better spatial and temporal predictions of the magnitude and extent of the ongoing shrub dynamics, ultimately increasing our ability to estimate the implications for ecosystem functioning (yellow).

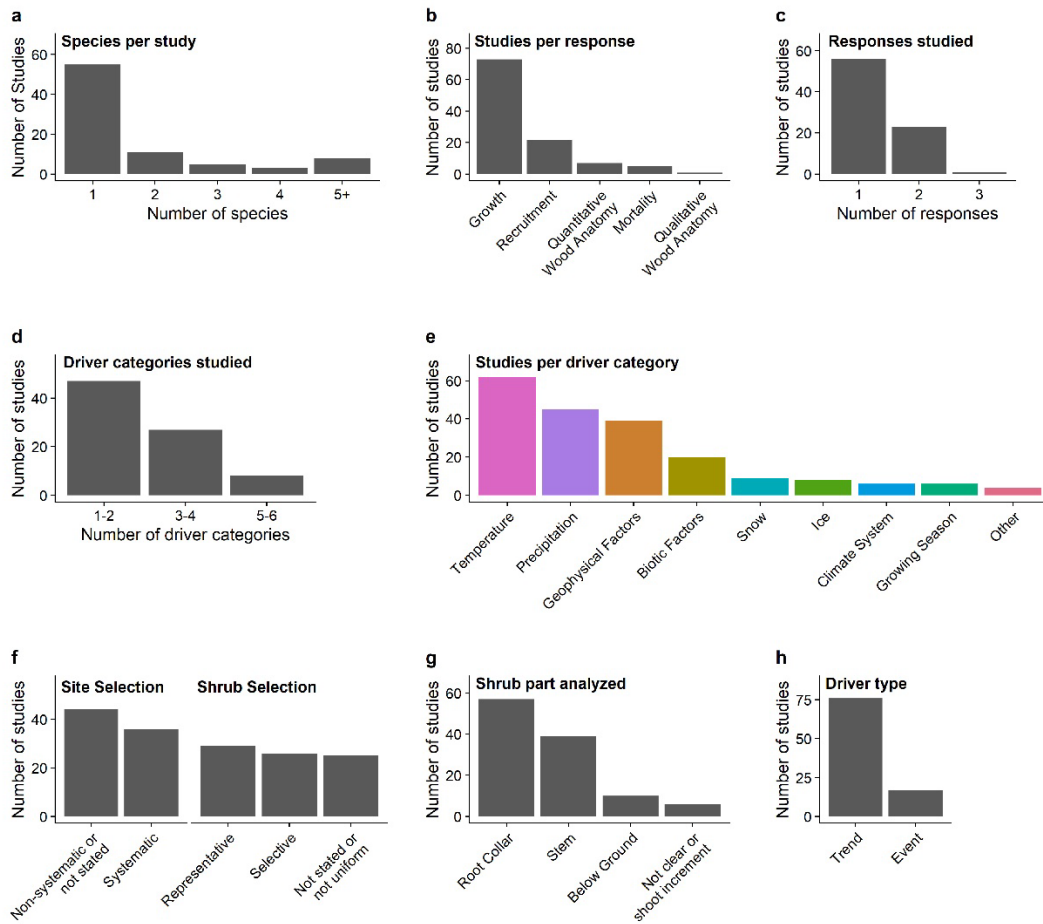


Figure 2: Overview of the main drivers, responses and methods used based on 82 Arctic dendrochronological studies from 2005-2022, showing: (a) the number of species, (b) the response type, (c) the number of responses, (d) the number of driver categories, (e) the driver categories, (f) the site and shrub selection method, (g) the shrub part analyzed and (h) the inclusion of drivers categorized as events or trends.

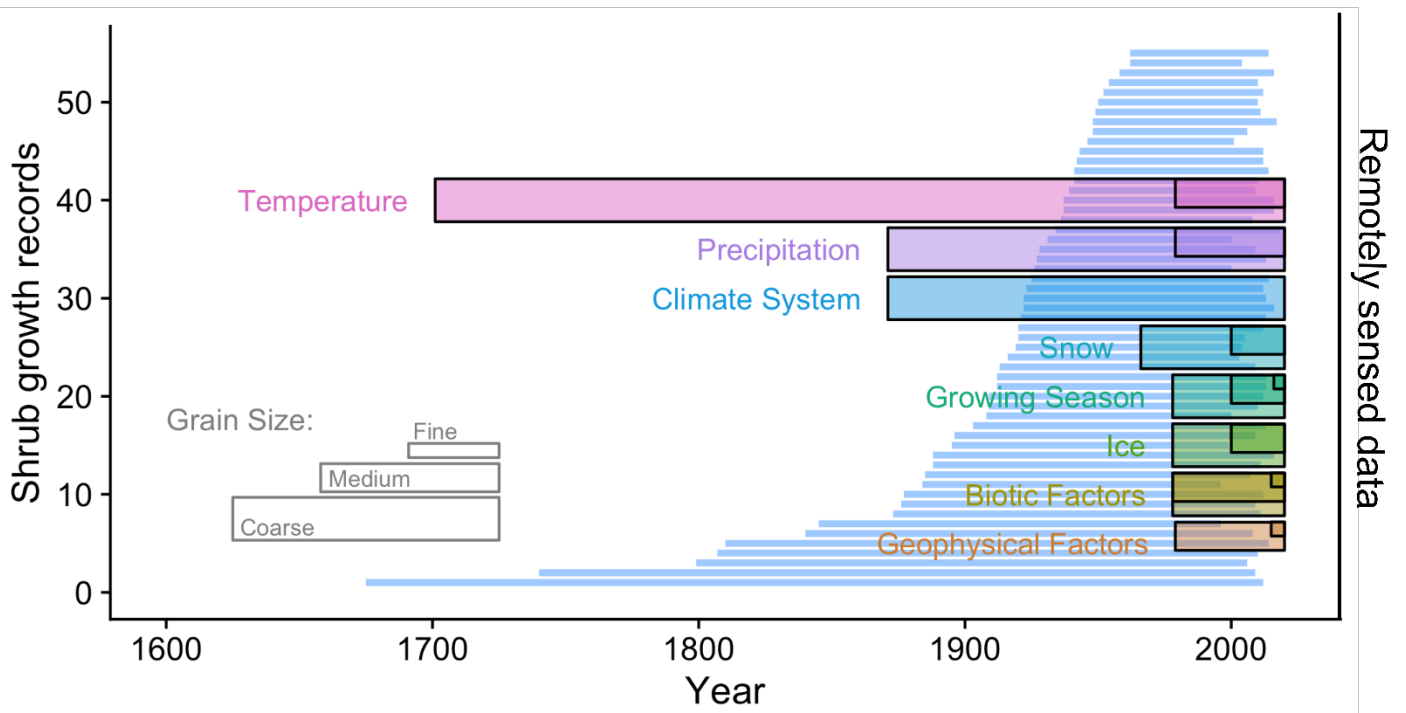


Figure 3: Extent of dendrochronological growth records and remotely sensed drivers. Left y-axis: Maximum extent of dendrochronological growth records from the reviewed studies (range: 1675-2018). We present the longest period covered by shrub growth in each study site, for all species combined. Right y-axis: The temporal availability of fully or partially remote sensing-based spatial datasets with global coverage that could provide direct or indirect measures for some of the drivers in the major categories identified in our review (see Table 1). The width of the bars indicates the grain size available: coarse (> 1 km), (10 m – 1 km) and fine (≤ 10 m). Table S1 provides a more detailed description of the remotely sensed data included.

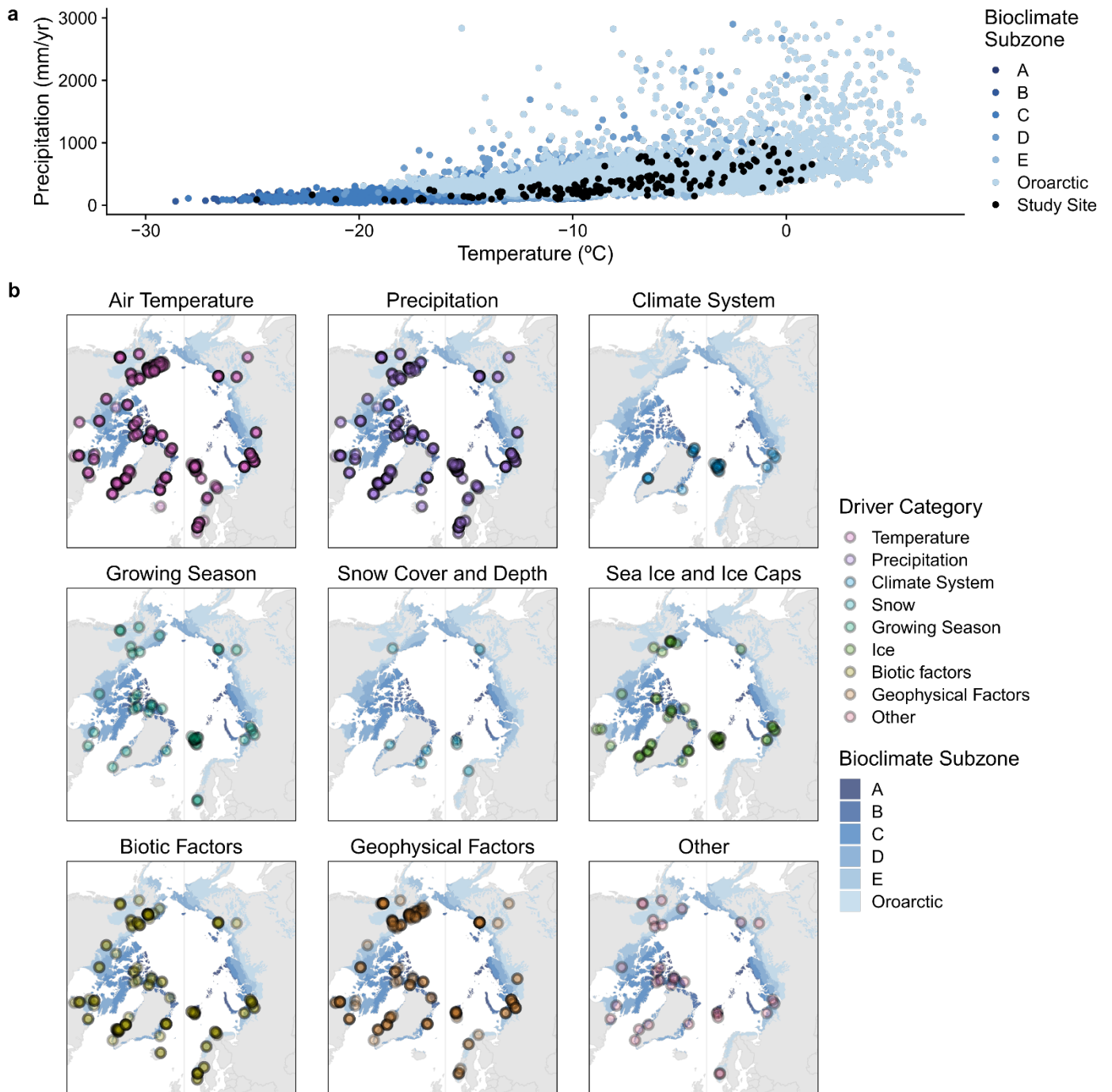


Figure 4: Location of study sites in the 82 reviewed studies in bioclimatic and geographic space. (a) The mean annual temperature and sum of annual precipitation at the study sites from 1979-2013, extracted from the CHELSA dataset v1.2.1 (Karger et al., 2017), plotted on top of 14k random locations across six Arctic bioclimate subzones to illustrate the climate space of the Arctic tundra. (b) Spatial distribution of study site locations by major driver category in the context of the tundra subzones.

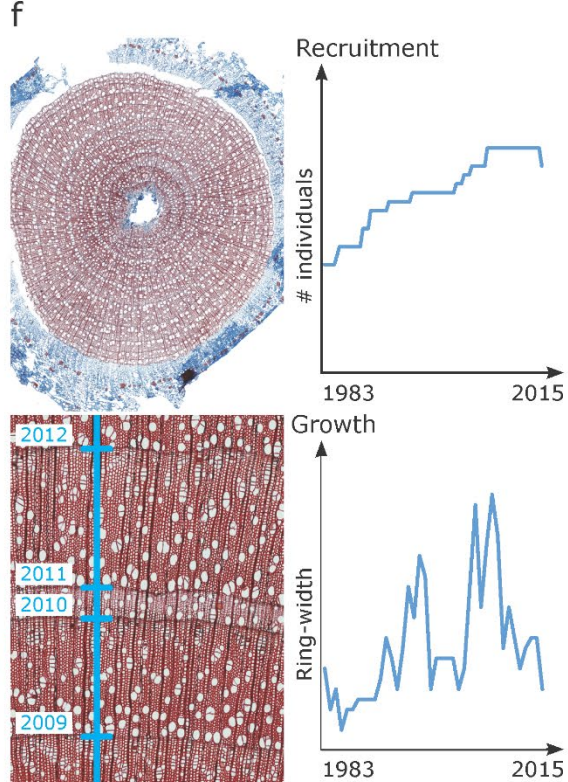
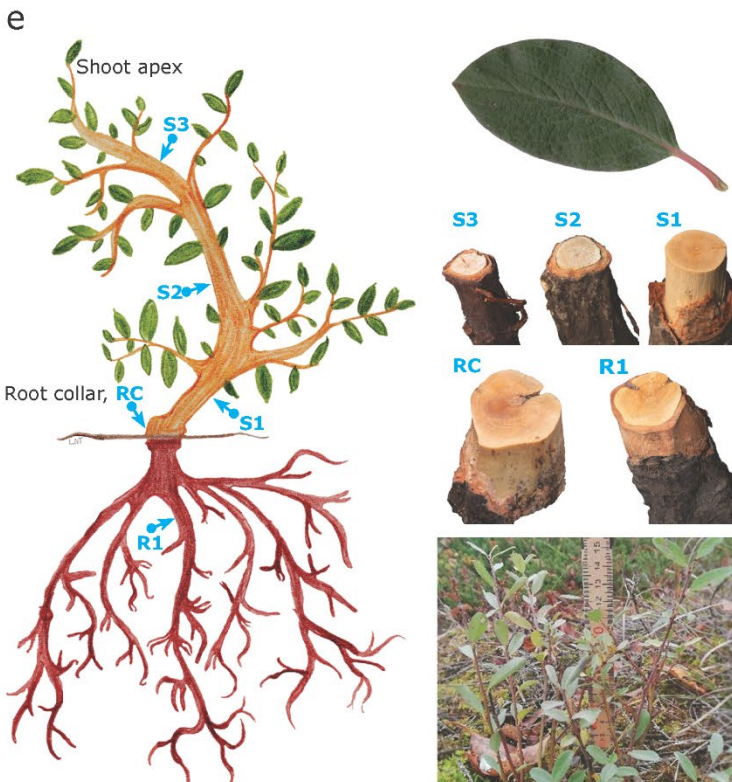
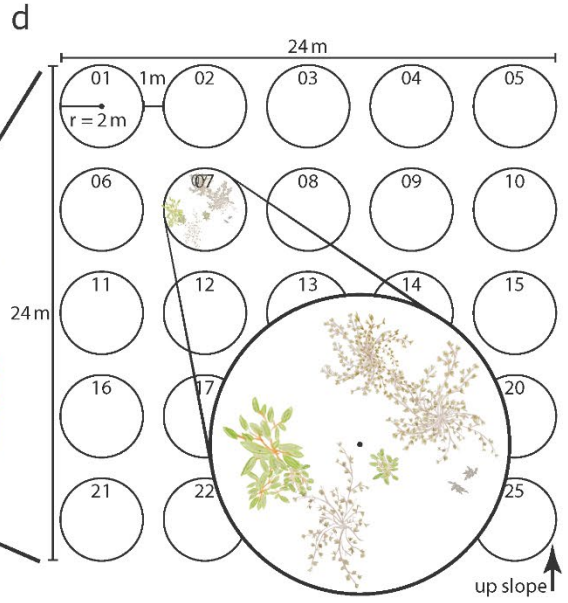
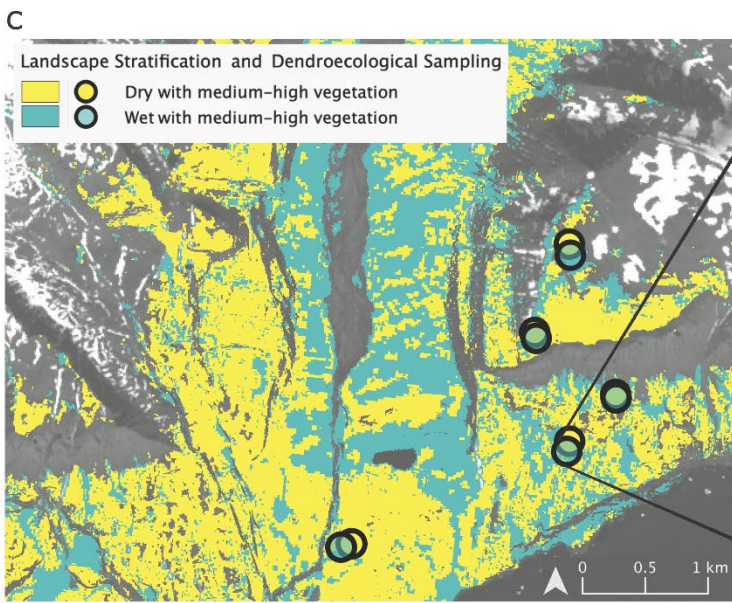


Figure 5: Illustration of the main components of our synergetic approach for cross-scale understanding of Arctic shrub community dynamics: (c) remote sensing-based environmental stratification; (b, d, e) community-based dendroecological (CBDE) sampling, and (f) dendroanatomy. Arctic environments can be heterogeneous at both (a) landscape and (b) local scales, as exemplified with this mountainous site located on Qeqertarsuaq (Disko Island), in western Greenland (69°16 N 53°27 W). (c) The landscape was stratified based on information from remotely-sensed data sources, i.e., a moisture proxy and vegetation greenness, and was the basis for selection of sampling plots (b,c). At each plot, a systematic sampling design was applied (d) by setting up a grid of 25 circles. In each circle, the shrub closest to the center point (d) was excavated, as completely as possible, and abiotic and biotic information was recorded (e). To aid with cross-dating and to facilitate within-individual analyses, sections of wood samples were taken along the main stem (S1, S2, S3) and root (R1), including the oldest part of the plant (the root collar (RC)), with the distance between each sample and the shoot apex recorded. Additional data and samples were collected per shrub, e.g., leaves for genetic and functional trait analyses (plant height, specific leaf area). The wood samples allow for detailed inference of growth and recruitment of the collected plants (f). In addition, anatomically detected growth anomalies (e.g., year 2010 (f)) can help to understand extreme events (here defoliation caused by an insect outbreak, e.g., Prendin et al 2020).

References (Figures and Tables)

- Abdalati, W., 2007. Greenland Ice Sheet Melt Characteristics Derived from Passive Microwave Data: 1979-2007. Digit. Media.
- Bartsch, A., Widhalm, B., Leibman, M., Ermokhina, K., Kumpula, T., Skarin, A., Wilcox, E.J., Jones, B.M., Frost, G. V., Höfler, A., Pointner, G., 2020. Feasibility of tundra vegetation height retrieval from Sentinel-1 and Sentinel-2 data. *Remote Sens. Environ.* 237, 111515. <https://doi.org/10.1016/j.rse.2019.111515>
- Belshe, E.F., Schuur, E.A.G., Grosse, G., 2013. Quantification of upland thermokarst features with high resolution remote sensing. *Environ. Res. Lett.* 8. <https://doi.org/10.1088/1748-9326/8/3/035016>
- Boelman, N.T., Holbrook, J.D., Greaves, H.E., Krause, J.S., Chmura, H.E., Magney, T.S., Perez, J.H., Eitel, J.U.H., Gough, L., Vierling, K.T., Wingfield, J.C., Vierling, L.A., 2016. Airborne laser scanning and spectral remote sensing give a bird's eye perspective on arctic tundra breeding habitat at multiple spatial scales. *Remote Sens. Environ.* 184, 337–349. <https://doi.org/10.1016/j.rse.2016.07.012>
- Brodzik, M., Armstrong, R., 2013. Northern Hemisphere EASE-Grid 2.0 Weekly Snow Cover and Sea Ice Extent. Version 4. Boulder, Color. USA NASA DAAC Natl. Snow Ice Data Center.
- Cooley, S.W., Ryan, J.C., Smith, L.C., Horvat, C., Pearson, B., Dale, B., Lynch, A.H., 2020. Coldest Canadian Arctic communities face greatest reductions in shorefast sea ice. *Nat. Clim. Chang.* 10, 533–538. <https://doi.org/10.1038/s41558-020-0757-5>
- Forbes, B.C., Kumpula, T., Meschytyb, N., Laptander, R., MacIas-Fauria, M., Zetterberg, P., Verdonen, M., Skarin, A., Kim, K.Y., Boisvert, L.N., Stroeve, J.C., Bartsch, A., 2016. Sea ice, rain-on-snow and tundra reindeer nomadism in Arctic Russia. *Biol. Lett.* 12, 4–8. <https://doi.org/10.1098/rsbl.2016.0466>
- Grenfell, T.C., Putkonen, J., 2008. A method for the detection of the severe rain-on-snow event on Banks Island, October 2003, using passive microwave remote sensing. *Water Resour. Res.* 44, 1–9. <https://doi.org/10.1029/2007WR005929>
- Howat, I.M., Negrete, A., Smith, B.E., 2014. The Greenland Ice Mapping Project (GIMP) land classification and surface elevation data sets. *Cryosphere* 8, 1509–1518. <https://doi.org/10.5194/tc-8-1509-2014>
- Karger, D.N., Conrad, O., Böhner, J., Kawohl, T., Kreft, H., Soria-Auza, R.W., Zimmermann, N.E., Linder, H.P., Kessler, M., 2017. Climatologies at high resolution for the earth's land surface areas. *Sci. Data* 4, 1–20. <https://doi.org/10.1038/sdata.2017.122>
- Nedbal, V., Láška, K., Brom, J., 2020. Mitigation of arctic tundra surface warming by plant evapotranspiration: Complete energy balance component estimation using LANDSAT satellite data. *Remote Sens.* 12, 1–16. <https://doi.org/10.3390/rs12203395>
- Obu, J., Westermann, S., Kääh, A., Bartsch, A., 2018. Ground Temperature Map, 2000-2016, Northern Hemisphere

Permafrost. Alfred Wegener Institute, Helmholtz Cent. Polar Mar. Res. Bremerhaven.
<https://doi.org/https://doi.org/10.1594/PANGAEA.888600>

- Porter, C., Morin, P., Howat, I., Noh, M.-J., Bates, B., Peterman, K., Keeseey, S., Schlenk, M., Gardiner, J., Tomko, K., Willis, M., Kelleher, C., Cloutier, M., Husby, E., Foga, S., Nakamura, H., Pl, M., 2018. ArcticDEM. Harvard Dataverse, V1.
- Prendin, A.L., Carrer, M., Karami, M., Hollesen, J., Bjerregaard Pedersen, N., Pividori, M., Treier, U.A., Westergaard-Nielsen, A., Elberling, B., Normand, S., 2020. Immediate and carry-over effects of insect outbreaks on vegetation growth in West Greenland assessed from cells to satellite. *J. Biogeogr.* 47, 87–100.
<https://doi.org/10.1111/jbi.13644>
- Raynolds, M.K., Walker, D.A., Balsler, A., Bay, C., Campbell, M., Cherosov, M.M., Daniëls, F.J.A., Eidesen, P.B., Ermokhina, K.A., Frost, G. V., Jedrzejek, B., Jorgenson, M.T., Kennedy, B.E., Kholod, S.S., Lavrinenko, I.A., Lavrinenko, O. V., Magnússon, B., Matveyeva, N. V., Metúsalemsson, S., Nilsen, L., Olthof, I., Pospelov, I.N., Pospelova, E.B., Pouliot, D., Razzhivin, V., Schaepman-Strub, G., Šibík, J., Telyatnikov, M.Y., Troeva, E., 2019. A raster version of the Circumpolar Arctic Vegetation Map (CAVM). *Remote Sens. Environ.* 232, 111297.
<https://doi.org/10.1016/j.rse.2019.111297>
- Robinson, D.A., Estilow, T.W., Program, N.C., 2012. NOAA Climate Data Record (CDF) or Northern Hemisphere (NH) Snow Cover Extent (SCE), Version 1 (2000-01-01 to 2019-12-31). NOAA Natl. Centers Environ. Information. <https://doi.org/10.7289/V5N014G9>
- Seddon, A.W.R., Macias-Fauria, M., Long, P.R., Benz, D., Willis, K.J., 2016. Sensitivity of global terrestrial ecosystems to climate variability. *Nature* 531, 229–232. <https://doi.org/10.1038/nature16986>
- Sun, Q., Miao, C., Duan, Q., Ashouri, H., Sorooshian, S., Hsu, K.L., 2018. A Review of Global Precipitation Data Sets: Data Sources, Estimation, and Intercomparisons. *Rev. Geophys.* 56, 79–107.
<https://doi.org/10.1002/2017RG000574>
- Zeng, H., Jia, G., Epstein, H., 2011. Recent changes in phenology over the northern high latitudes detected from multi-satellite data. *Environ. Res. Lett.* 6. <https://doi.org/10.1088/1748-9326/6/4/045508>