

1 **Modeling the transport of aggregating nanoparticles**
2 **in porous media**

3 by

4
5 **Vasileios E. Katzourakis* and Constantinos V. Chrysikopoulos**

6
7 *School of Environmental Engineering, Technical University of Crete, Chania*
8 *73100, Greece*

9
10
11
12
13
14 Manuscript to be submitted to
15 *Water Resources Research*

16
17
18
19
20 Key words: nanoparticles, transport, aggregation, reversible attachment,
21 irreversible attachment, porous media, mathematical modeling.

22
23 November 8, 2020

24
25
26
27 *Corresponding author (biliskatz@yahoo.gr).

Abstract

A novel mathematical model was developed to describe the transport of nanoparticles in water saturated, homogeneous porous media with uniform flow. The model accounts for the simultaneous migration and aggregation of nanoparticles. The nanoparticles are assumed to be found suspended in the aqueous phase or attached reversibly or irreversibly onto the solid matrix. The Derjaguin-Landau-Verwey-Overbeek theory was used to account for possible repulsive interactions between aggregates. Nanoparticle aggregation was represented by the Smoluchowski population balance equation (PBE). Both reaction-limited aggregation and diffusion-limited aggregation were considered. Particle-size dependent dispersivity was accounted for. In order to overcome the substantial difficulties introduced by the PBE, the governing coupled partial differential equations were solved by employing adaptive operator splitting methods, which decoupled the reactive transport and aggregation into distinct physical processes. The results from various model simulations showed that the transport of nanoparticles in porous media is substantially different than the transport of conventional biocolloids. In particular, aggregation was shown to either decrease or increase nanoparticle attachment onto the solid matrix, depending on particle size, and to yield early or late breakthrough, respectively. Finally, useful conclusions were drawn regarding possible erroneous results generated when aggregation, particle-size dependent dispersivity or nanoparticle surface charges are neglected.

1. Introduction

In recent years, nanotechnology has become one of the most promising industry sector with many applications in healthcare, medicine, molecular biology, semiconductor physics, and agriculture. However, despite their significant benefits some nanomaterials, such as metal oxide nanoparticles are considered toxic (IARC, 2010). Nanoparticles enter the environment from wastewaters originating from industrial or house-hold sources, which do not undergo proper treatment (Benn & Westerhoff, 2008; Brar et al., 2010; Gottschalk et al., 2009; Mueller & Nowack, 2008), and from accidental release or inappropriate disposal of nanomaterials (Brar et al., 2010; Nowack &

68 Bucheli, 2007; Wiesner et al., 2006). These nanomaterials often contribute to
69 the pollution of aquatic and other terrestrial environments.

70 Nanoparticle transport differs significantly from conventional biocolloid
71 transport, because particles may aggregate and form larger particles with
72 different physical characteristics (Solovitch et al., 2010). Consequently, the
73 classical filtration theory may fail to capture the attachment dynamics of
74 nanoparticles (Chen et al., 2011; Chowdhury et al., 2011; Fang et al., 2009;
75 Godinez & Darnault, 2011; Heidmann, 2013; Zhang et al., 2014). The
76 aggregation process can be classified into two distinct categories: (i) diffusion-
77 limited aggregation (DLA), and (ii) reaction-limited aggregation (RLA) (Wijnen
78 et al., 1991; Gaudreault et al., 2015). When no repulsive forces are present
79 between particles, then every collision leads to attachment. This is essentially
80 a DLA process, which is usually referred to as “fast aggregation” and yields
81 aggregates with plenty of void spaces. If repulsive forces exist between
82 particles, then aggregation is slowed down because multiple collisions may be
83 needed before a successful particle attachment. This is a RLA process, which
84 is usually referred to as “slow aggregation” and produces dense aggregates
85 (Gaudreault et al., 2015; Lin et al., 1990; Weitz et al., 1991; Weitz & Lin,
86 1986).

87 Nanoparticle aggregation is an important process for particle attachment
88 during transport in porous media. However, the available mathematical
89 models for particle transport, which are based on colloid filtration theory
90 (CFT), depth-dependent retention and blocking, despite their success in fitting
91 relatively well experimental data, frequently do not capture the
92 physicochemical processes that nanoparticles undergo during transport in
93 porous media (Goldberg et al., 2014). Also, the available mathematical
94 models that try to couple the transport equation with an expression for
95 aggregation (Chatterjee & Gupta, 2009; Raychoudhury et al., 2012; Taghavy
96 et al., 2015; Quik et al., 2015; Babakhani et al., 2018) may provide improved
97 results, but either they do not take into account for appropriate particle
98 dispersion or they fail to account for the existence of repulsive forces between
99 charged particles. Other models use simplifying or empirical reaction rates
100 (Babakhani et al., 2019), and general attachment equations (Wang et al,
101 2018) to account for transport and aggregation of particles. The mathematical

102 model developed by Babakhani (2019) takes into account transport and
103 aggregation of nanoparticles and evaluates their size exclusion. It was shown
104 that accounting for particle aggregation improved substantially the predictive
105 ability of the model. However, the Babakhani (2019) mathematical model
106 does not contain explicit transport and aggregation terms and does not
107 account for thorough particle attachment onto the solid matrix (e.g. with a two
108 site reversible/irreversible kinetic model).

109 The aim of this work is to develop a novel mathematical model for the
110 description of the transport of aggregating nanoparticles, in water saturated,
111 homogeneous porous media, with fully developed uniform flow, in which there
112 is a clear formulation of how transport and aggregation terms are coupled.
113 The model accounts for changes in particle attachment onto the solid matrix
114 due to evolving size of aggregated particles and for potential repulsive
115 interactions between particles. To the best of our knowledge such unique
116 model for the transport of suspended nanoparticles undergoing two-site
117 attachment and aggregation in porous media is not available in literature.

118

119 **2. Mathematical developments**

120 **2.1 Transport of nanoparticles**

121 The proposed nanoparticle transport model assumes that particles can
122 aggregate and partition between the aqueous phase and the solid matrix. The
123 forming aggregates can be classified based on their average diameter into κ
124 clusters, where $\kappa=1,2,3,\dots$ is the cluster incremental number (i.e. cluster $\kappa = 1$
125 consists of monomers, while cluster $\kappa = 2$ consists of dimers). Nanoparticles
126 can be found suspended in the aqueous phase with number concentration n_κ
127 $[n_{\kappa}/L^3]$ (where n_{κ} is the number of aggregates of cluster κ), or attached onto
128 the solid matrix $n_\kappa^* [n_{\kappa}/M_s]$ (where M_s is the mass of the solid matrix).
129 Consequently, the governing partial differential equation describing the
130 transport of nanoparticles that belong to cluster κ , in one-dimensional,
131 homogeneous, water saturated porous media with developed one-directional
132 uniform flow, accounting for non-equilibrium attachment onto the solid matrix
133 is essentially the well-established transport equation for colloids (Sim &
134 Chrysikopoulos, 1998; Katzourakis & Chrysikopoulos, 2014) written in terms

135 of particle number density (number concentration instead of mass
 136 concentration) with an additional sink/source term which accounts for
 137 nanoparticle aggregation (Lee et al., 2000; Sabelfeld & Kolodko, 2002):

$$138 \quad \frac{\partial n_k(t,x)}{\partial t} + \frac{\rho_b}{\theta} \frac{\partial n_k^*(t,x)}{\partial t} - (D_x)_k \frac{\partial^2 n_k(t,x)}{\partial x^2} + U \frac{\partial n_k(t,x)}{\partial x} = (F_n)_k(t,x) + (A_n)_k(t,x) \quad (1)$$

139 where U [L/t] is the average interstitial velocity; $(D_x)_k$ [L²/t] is the longitudinal
 140 hydrodynamic dispersion coefficient of the suspended nanoparticles that
 141 belong to cluster k ; ρ_b [Ms/L³] is the bulk density of the solid matrix; θ [-] is the
 142 porosity of the porous medium; x [L] is the spatial coordinate in the
 143 longitudinal direction; t [t] is time; $(F_n)_k(t,x)$ [npk/L³t] is a general source
 144 configuration form of the nanoparticles that belong to cluster k ; and $(A_n)_k(t,x)$
 145 [npk/L³t] is the aggregation source/sink term for nanoparticles that belong to
 146 cluster k .

147 The nanoparticle aggregation source/sink term is assumed to be
 148 accurately represented by the Smoluchowski population balance equation
 149 (PBE), which describes the evolution of the mass spectrum of a collection of
 150 particles due to successive merges (Smoluchowski, 1916):

$$151 \quad (A_n)_k = \frac{dn_k}{dt} = \frac{1}{2} \sum_{i=1}^{k-1} b_{i,k-i} n_i n_{k-i} - n_k \sum_{i=1}^{\infty} b_{k,i} n_i \quad (2)$$

152 where $b_{i,k}$ is an aggregation kernel, referring to the collision frequency of the
 153 nanoparticles. The 1/2 multiplier in front of the first summation term corrects
 154 for the double counting of particle collisions.

155 The attachment of nanoparticles onto the solid matrix is assumed to be
 156 reversible or irreversible. Consequently, the number density of nanoparticles
 157 attached onto the solid matrix, n_k^* [npk/Ms], is the sum of the reversibly, $n_k^{*(r)}$
 158 [npk/Ms], and irreversibly, $n_k^{*(i)}$ [npk/Ms], attached particle concentrations:

$$159 \quad n_k^* = n_k^{*(r)} + n_k^{*(i)} \quad (3)$$

160 Therefore, the corresponding nanoparticles accumulation term in equation (1)
 161 is expressed as:

$$162 \quad \frac{\partial n_k^*}{\partial t} = \frac{\partial n_k^{*(r)}}{\partial t} + \frac{\partial n_k^{*(i)}}{\partial t} \quad (4)$$

163 The reversible nanoparticle accumulation term is described by the following
 164 nonequilibrium equation (Sim & Chrysikopoulos, 1998; Sim & Chrysikopoulos,
 165 1999):

$$166 \quad \frac{\rho_b}{\theta} \frac{\partial n_k^{*(r)}}{\partial t} = r_{n_k - n_k^{*(r)}} n_k - r_{n_k^{*(r)} - n_k} \frac{\rho_b}{\theta} n_k^{*(r)} \quad (5)$$

167 where $r_{n_k - n_k^{*(r)}} [1/t]$ is the rate coefficient of reversible nanoparticle attachment
 168 onto the solid matrix, and $r_{n_k^{*(r)} - n_k} [1/t]$ is the rate coefficient of reversible
 169 nanoparticle detachment from the solid matrix. The irreversible accumulation
 170 term is described by the following nonequilibrium equation (Compère et al.,
 171 2001; Katzourakis & Chrysikopoulos, 2014):

$$172 \quad \frac{\rho_b}{\theta} \frac{\partial n_k^{*(i)}}{\partial t} = r_{n_k - n_k^{*(i)}} n_k \quad (6)$$

173 where $r_{n_k - n_k^{*(i)}} [1/t]$ is the rate coefficient of irreversible nanoparticle attachment
 174 onto the solid matrix. It should be noted that under steep velocity changes or
 175 time varying salinity and pH fluctuations alternative expressions to equation
 176 (5) exist in the literature (Bedrikovetsky et al., 2011, 2012; Russell &
 177 Bedrikovetsky, 2018).

178 The general form of the source configuration of the nanoparticles of
 179 cluster k, can be written as (Sim & Chrysikopoulos, 1999):

$$180 \quad (F_n)_k(t, x) = G_k(t) W(x) \quad (7)$$

181 where $G_k(t) [np_k/L^2t]$ is the particle release function, and $W(x) [1/L]$ describes a
 182 point source geometry:

$$183 \quad W(x) = \delta(x - x_0) \quad (8)$$

184 where $\delta(x - x_0) [1/L]$ is the Dirac delta function, and $x_0 [L]$ is the Cartesian x-
 185 coordinate of the source centre. For a broad pulse, the function $G_k(t)$ is given
 186 by:

$$187 \quad G_k(t) = \frac{Nr_k}{\theta} H(t_p - t) \quad (9)$$

188 where $Nr_k [np_k/L^2t]$ is the point source release rate of particles that belong to
 189 cluster k; $t_p [t]$ is the source release period over which nanoparticles enter the

190 porous medium; and $H(t)$ [-] is the unit step or Heaviside function ($H(t<0)=0$,
 191 $H(t\geq 0)=1$). For an instantaneous source, $G_k(t)$ is given by:

$$192 \quad G_k(t) = \frac{(N_{inj})_k}{A_c \theta} \delta(t) \quad (10)$$

193 where $(N_{inj})_k$ [np_k] is the injected number of particles that belong to cluster k ,
 194 A_c [L^2] is the cross-sectional area of the porous medium, and $\delta(t)$ [$1/t$] is the
 195 Dirac delta function. Note that using equations (7)-(10), it is possible to define
 196 a broad pulse or instantaneous nanoparticle point source, located anywhere
 197 within the aquifer with x -coordinate $x=x_0$.

198
 199

200 **2.2 Initial and boundary equations**

201 The initial condition and the appropriate boundary conditions for a one-
 202 dimensional confined aquifer with finite dimensions are as follows:

$$203 \quad n_k(0,x) = 0 \quad (11)$$

$$204 \quad n_k(t,0) = \begin{cases} n_k^0, & t \leq t_p \\ 0, & t > t_p \end{cases} \quad (12)$$

$$205 \quad n_k(t,0) = 0 \quad (13)$$

$$206 \quad \frac{\partial n_k^2(t, L_x)}{\partial x^2} = 0 \quad (14)$$

207 where L_x [L] is the length of the porous medium and n_k^0 [np_k] is the initial
 208 constant aqueous phase concentration of cluster k . Condition (11) establishes
 209 that initially there are no nanoparticles within the porous medium. Condition
 210 (12) represents a broad pulse injection with constant nanoparticle
 211 concentration at the inlet. Condition (13) indicates that nanoparticles are not
 212 entering the aquifer through the inlet, but they are injected at a specific
 213 location within the aquifer according to equations (7)-(10). The downstream
 214 boundary condition (14) preserves concentration slope continuity for the finite
 215 length aquifer (Shamir & Harleman, 1967). It should be noted that the initial
 216 and boundary conditions (11)-(14) are applied k times, once for each cluster.

217

218 **2.3 Aggregation kernel**

219 The term b_{ij} in equation (2) represents the collision rate between particles that
 220 belong to clusters i and j . A variety of collision frequency kernels are available
 221 in the literature that account for different physicochemical conditions. One of
 222 the most commonly used kernels for DLA processes (Axford, 1997;
 223 Smoluchowski, 1917), which accounts for collisions resulting from Brownian
 224 diffusion while ignoring negligible contributions from fluid shear and
 225 sedimentation (Petosa et al., 2010; Taghavy et al., 2015) is:

$$226 \quad b_{ij}^{DLA} = \frac{2k_B T (r_i + r_j)^2}{3\mu_w r_i r_j} \quad (15)$$

227 where k_B [$M \cdot L^2 / (t^2 \cdot T)$] is the Boltzmann constant; T [K] is temperature; r_k [L]
 228 is the radius of a nanoparticle that belongs to cluster k ; and μ_w [$M / (t \cdot L)$] is the
 229 dynamic viscosity of water. The ratio $k_B T / \mu_w$ characterizes the diffusion of
 230 suspended particles due to Brownian movement. Larger values of this ratio
 231 (caused by temperature increase) result to increased collision frequency.
 232 Also, the parabolic ratio $(r_i + r_j)^2 / r_i r_j$ indicates that the collision frequency is
 233 higher between particles of different sizes than for particles of the same size.
 234 For RLA processes, the collision frequency kernel, b_{ij}^{RLA} , must account for
 235 repulsive forces produced when similarly charged particles interact. This can
 236 be achieved by using the Fuchs stability ratio $w_{ij} > 1$ [-], which is defined as
 237 the ratio of aggregation rate of a particle in the absence of repulsive
 238 interactions to the aggregation rate when the repulsive interactions are
 239 present (Fuchs, 1934; Lattuada et al., 2003). Values of w_{ij} close to unity
 240 indicate fast aggregation and refer to an "unstable" particle suspension, while
 241 larger values of $w_{ij} \gg 1$ indicate slow aggregation and refer to a "stable"
 242 particle suspension. The b_{ij}^{RLA} is related to b_{ij}^{DLA} as follows (Amal et al., 1990;
 243 Arosio et al., 2012):

$$244 \quad b_{ij}^{RLA} = \frac{b_{ij}^{DLA}}{w_{ij}} \quad (16)$$

245 where w_{ij} [-] can be expressed as (Axford, 1997; Liu et al., 2011; Reerink &
 246 Overbeek, 1954):

247
$$w_{ij} = 2 \int_2^{\infty} \exp \left[\frac{(\Phi_{\text{tot}})_{ij}}{k_B T} \right] \frac{1}{s^2} ds \quad (17)$$

248 where the dimensionless parameter s [-] is given by:

249
$$s = \frac{2R}{r_i + r_j} \quad (18)$$

250 where R [L] is the distance between the centers of two colliding particles; r_i [L]
 251 and r_j [L] are the radii of particles i and j , respectively; $(\Phi_{\text{tot}})_{ij}$ [M·L²/t²] is the
 252 total interaction energy between particles i and j , which is a function of s and
 253 can be calculated from the DLVO theory. It is evident from equation (17) that
 254 the ratio of the interaction energy to the thermal energy $(\Phi_{\text{tot}})_{ij}/k_B T$ dictates
 255 the value of stability ratio w_{ij} . If the available energy $k_B T$ is consistently greater
 256 than the energy barrier $k_B T > (\Phi_{\text{tot}})_{ij}$, regardless of distance s , then the Fuchs
 257 ratio will obtain values close to unity $w_{ij} \approx 1$ and fast aggregation DLA will
 258 occur. Otherwise, the existing thermal energy $k_B T$ will not be able to
 259 overcome easily the energy barrier and slow aggregation RLA will take place.
 260 Furthermore, the dimensionless distance of the two particles, s , indicates that
 261 the effects of the energy barrier $(\Phi_{\text{tot}})_{ij}$ decay fast with distance. Therefore,
 262 increased interaction potential over shorter distances leads to higher Fuchs
 263 stability ratio w_{ij} .

264

265 **2.4 Aggregate structure**

266 According to the coalesced sphere assumption, two spherical particles
 267 collide and form a new spherical aggregate. The mass of the produced
 268 aggregate is the sum of the masses of the two initial particles, while the same
 269 is true for their volumes. Therefore, the aggregate density is maintained
 270 constant. However, in reality, the resulting aggregates contain void spaces.
 271 The relation between the diameter of the final aggregate, $(d_p)_k$, and the initial
 272 monomer, $(d_p)_1$, is (Feder, 1988; Lee et al., 2000):

273
$$(N_p)_k = \zeta \left[\frac{(d_p)_k}{(d_p)_1} \right]^{D_F} \quad (19)$$

274 where $(N_p)_k$ [npk] is the number of particles present in an aggregate that
 275 belongs to cluster k; $(d_p)_k$ [L] is the diameter of the produced aggregate that
 276 belongs to cluster k; ζ [-] is the packing factor, which accounts for the void
 277 pore space within the spherical aggregate and depends on the shape of both
 278 monomers and aggregates; D_f [-] is the fractal dimension of an aggregate and
 279 depends on the type of aggregation. For spherical monomers in close packing
 280 $\zeta=0.7405$, whereas, in random packing $\zeta=0.637$ (Feder, 1988). The slow RLA
 281 process usually yields aggregates with $D_f=2.1$, while the fast DLA yields
 282 aggregates with $D_f=1.75$ (Gaudreault et al., 2015; Lin et al., 1989). Finally,
 283 the mean particle diameter of aggregates suspended in the solution, \bar{d}_p [L],
 284 can be written as a function of the individual aggregate diameters:

$$285 \quad \bar{d}_p = \frac{\sum_{i=1}^k (N_p)_i (d_p)_i}{\sum_{i=1}^k (N_p)_i} \quad (20)$$

286 Equations (16) and (17) for the description of aggregation kernel b_{ij}^{RLA}
 287 are not practical, because the exact way that the total interaction potential
 288 $(\Phi_{tot})_{ij}$ scales with the aggregate size, frequently is unknown. Therefore, in
 289 the absence of experimental information relating the aggregate structure, a
 290 scaling factor P_{ij} is used (Arosio et al., 2012; Nicoud et al., 2014; Sandkühler
 291 et al., 2004) and equation (16) takes the form:

$$292 \quad b_{ij}^{RLA} = \frac{b_{ij}^{DLA}}{w_{11}} P_{ij} \quad (21)$$

293 where w_{11} [-] is the Fuchs ratio for aggregation of two monomers; P_{ij} [-] is
 294 often represented by the product kernel: $P_{ij}=(ij)^\lambda$ (Arosio et al., 2012; Family et
 295 al., 1985), which has been proven to perform well (Lattuada et al., 2003;
 296 Nicoud et al., 2014). The value of the exponent λ [-] is typically within the
 297 range 0.25-0.5 (Lin et al., 1990; Sandkühler et al., 2004). Assuming that the
 298 interactions between two aggregates are governed mainly by the monomers
 299 on the surface of the aggregates, the coefficient λ can be expressed

300 analytically as $\lambda = 1 - 1/D_F$ (Arosio et al., 2012; Nicoud et al., 2014), and b_{ij}^{RLA}
 301 becomes:

$$302 \quad b_{ij}^{RLA} = \frac{b_{ij}^{DLA}}{w_{11}} (ij)^{1 - \frac{1}{D_F}} \quad (22)$$

303 Note that b_{ij}^{RLA} should never be greater than b_{ij}^{DLA} , because the latter one is
 304 the maximum aggregation rate, where every collision results in aggregation.
 305 Consequently, if the ratio $w_{11} / P_{ij} < 1$, it must be set equal to unity (Sandkühler
 306 et al., 2004). Please note that DLA occurs in the absence of repulsive
 307 interactions, making aggregates with lower fractal dimensions, while RLA
 308 occurs in the presence of repulsive interactions, making aggregates with
 309 higher fractal dimensions.

310

311 **2.5 Interaction between particles**

312 According to the DLVO theory the total interaction energy $\Phi_{DLVO}(h)$
 313 between two smooth and homogeneous surfaces can be estimated as the
 314 sum of the electrostatic repulsion energy arising from the interaction of
 315 electrical double layers, the attractive van der Waals forces, and the Born
 316 repulsion energy (Loveland et al., 1996):

$$317 \quad \Phi_{DLVO}(h) = \Phi_{vdW}(h) + \Phi_{dl}(h) + \Phi_{Born}(h) \quad (23)$$

318 where Φ_{vdW} [J] is the van der Waals energy estimated by the relationship
 319 reported by Gregory (1981), Φ_{dl} [J] is the electrostatic interaction energy
 320 estimated by the relationship reported by Hogg et al. (1966), Φ_{Born} [J], is the
 321 Born interaction energy estimated by the relationship provided by Ruckenstein
 322 & Prieve (1976), and h [L] is the separation distance between two
 323 approaching particle surfaces.

324

325 **2.6 Filtration theory**

326 The forward rate coefficient found on the right-hand side of equation (5)
 327 can be defined as (Sim & Chrysikopoulos, 1995):

$$328 \quad r_{n_k - n_k^*(t)} = U\Phi F(n_k^*) \quad (24)$$

329 where Φ [1/L] is the filter coefficient; $F(n_k^*)$ [-] is the dynamic blocking function
 330 that accounts for porosity variations when particle attachment increases. For

331 submicron particles, such as nanoparticles, it can be assumed that the porous
 332 medium is “clean,” and $F(n_k^*) = 1$. The filter coefficient Φ can be calculated as
 333 (Rajagopalan & Tien, 1976):

$$334 \quad \Phi = \frac{3(1-\theta)}{2d_c} \eta \quad (25)$$

335 where d_c [L] is the average diameter of the collector; and η [-] is the single
 336 collector removal efficiency (Yao et al., 1971):

$$337 \quad \eta = \alpha \eta_o \quad (26)$$

338 where α [-] is the collision efficiency; and η_o [-] is the single collector contact
 339 efficiency, which can be estimated by the correlation developed by Tufenkji
 340 and Elimelech (2004). Note that using equations (24)-(26) it is possible to
 341 calculate the forward rate coefficient of nanoparticle attachment onto the solid
 342 matrix, $r_{n_k - n_k^*(r)}$, as a function of the aggregated particle size.

343

344

345 **3. Numerical methods**

346 **3.1. General solution procedure**

347 The solution of the governing nanoparticle transport equation (1) is quite
 348 difficult because multiple physical processes (dispersion, advection,
 349 attachment, aggregation) are accounted as a “family” of coupled partial
 350 differential equations and in conjunction with equations (2)-(10) a closed
 351 system of equations is formed consisting of $3 \times k$ unknowns ($n_k, n_k^{s(r)}, n_k^{s(i)}$).
 352 Every time the total number of classes k_{\max} increases by one, three more
 353 unknown variables are added along with a new set of equations (1)-(10),
 354 making sure the new system is well defined. A direct solution approach for
 355 equations (1)-(10) is not possible because the nonlinear PBE equation (2) is
 356 coupled to the governing equation (1). Also, conventional numerical
 357 approaches would require enormous memory. One efficient alternative
 358 method of solution is to decouple the physical processes through operator
 359 splitting schemes and solve them one at a time (Barry et al., 2000; Kanney et
 360 al., 2003; Steefel & MacQuarrie, 1996; Wood & Baptista, 1993).

361 The solution approach employed here was to decouple the reactive
 362 transport from the aggregation process by using an adaptive double step in
 363 conjunction with the symmetrically weighted sequential (SWS) splitting
 364 operator method (Botchev et al., 2004). The SWS is a second-order accurate
 365 in time scheme. The double adaptive time step allows estimation of the local
 366 error by either executing one time-step of size Δt or two sequential steps of
 367 size $\Delta t/2$. Therefore, depending on the resulting relative error of these two
 368 steps, Δt was adjusted to meet specific criteria. The decoupled processes
 369 were solved separately. First, the transport equation (1), without the
 370 aggregation source/sink term $(A_n)_k$ and the attachment term $(\rho_b/\theta)(\partial n_k^*/\partial t)$,
 371 was solved using the implicit second-order Crank-Nicolson scheme. Next, the
 372 resulting concentration values were updated by an iterative process, which
 373 involved the solution of equations (4)-(6) for the attachment process
 374 (Kinzelbach et al., 1991). Finally, the aggregation process described by
 375 equation (2) was solved with subroutine Dodesol (Intel® Ordinary Differential
 376 Equations Solver Library), which in conjunction with the SWS scheme, is
 377 capable of solving systems of ordinary differential equations with a variable or
 378 a priori unknown stiffness.

379

380 3.2. Number of clusters

381 The Smoluchowski equation (2) describes the particle aggregation
 382 process, but it does not set explicitly an upper limit on the number of clusters
 383 that may occur. As the aggregation process progresses, larger nanoparticles
 384 are created. However, the solution of the Smoluchowski equation with a
 385 differential equation solver requires a finite number of clusters. There is no
 386 limitation how big the max number of clusters, k_{max} , can be, because
 387 everytime the number of unknowns $(n_k, n_k^{*(r)}, n_k^{*(i)})$ increases so does the
 388 number of available equations and the system remains closed. Because there
 389 is an exponential relation between the number of clusters and the number of
 390 calculations needed, k_{max} should be as small as possible. In this work k_{max} was
 391 selected by repeating the same simulation multiple times, while each time the
 392 k_{max} value was progressively increased until a subsequent increase in the

393 k_{\max} value did not alter significantly the resulting breakthrough curves. The
 394 accepted maximum relative error on the non-negligible concentrations
 395 between different simulations for the selected k_{\max} was lower than <2%.

396

397 **4. Model simulations and discussion**

398 **4.1 Numerical model verification**

399 The present nanoparticle transport model was compared against: (i) a
 400 simple aggregation process under batch conditions (without transport), to
 401 validate the accuracy of the numerical methods used for the solution of the
 402 aggregation process; and (ii) a simple transport simulation (without
 403 aggregation) carried out with the commercial software Comsol™, to ensure
 404 that the transport was accurately solved. For the first comparison the
 405 aggregation equation (2) with kernel $b_{ij} = 1$ was compared to the following
 406 analytical solution (Smoluchowski, 1916):

$$407 \quad n_k(k, t) = \left(1 + \frac{t}{2}\right)^{-2} \left(\frac{t}{2+t}\right)^{k-1} \quad (27)$$

408 The resulting dimensionless concentrations (n_k / n_1^0) are shown in Figure 1a
 409 for two different clusters ($k=10, 20$). Clearly, there is a perfect match between
 410 the analytical and numerical solution. For the second comparison, a
 411 hypothetical one-dimensional aquifer with length $L_x=0.6$ m and cross-section
 412 $A_c=4.91 \times 10^{-4}$ m², consisting of sand grains (collectors) with diameter
 413 $d_c = 6 \times 10^{-4}$ m was considered. Subsequently, this hypothetical aquifer will
 414 be referred to as “1-D aquifer”. A constant number concentration $n_1^0=1 \times 10^3$
 415 np_1/m^3 entered the 1-D aquifer at $x=0$ m, for a time period of $t_p=15$ hr. The
 416 model simulations were conducted with $(D_x)_{n_1} = 0.09$ [m/hr²], $r_{n_k - n_k^*(r)} = 0.25$
 417 [1/hr²], $r_{n_k^*(r) - n_k} = 0.01$ [1/hr²], $U=0.3$ [m/hr²] and $G(t)=0$ [np_k/t] and other required
 418 parameter values listed in Table 1. All aggregate clusters were assigned the
 419 same dispersion coefficient and forward attachment rate, in order to have a
 420 direct comparison with the Comsol™ transport model. Note that the Comsol™
 421 model employed the same equations used in the numerical model developed

422 here, but the term $(A_n)_k(t,x)$ in equation (1), which describes nanoparticle
 423 aggregation, was removed. The resulting dimensionless breakthrough
 424 concentrations (n_1^T / n_1^0) , shown in Figure 1b, are in perfect agreement with the
 425 results from the present nanoparticle transport model. Note that n_1^T [np_1/L^3] is
 426 the total number concentration of suspended nanoparticles (sum of
 427 nanoparticles initially present in cluster $k=1$, which at subsequent times
 428 contribute to formation of aggregates in various clusters), and n_1^0 [np_1/L^3] is
 429 the initially injected number concentration of particles that belong to cluster
 430 $k=1$.

431

432 4.2 Attachment rate

433 Assuming that the attachment of nanoparticles onto collector grains is
 434 controlled mainly by the collision efficiency, the forward rate coefficient of
 435 reversible nanoparticle attachment onto the solid matrix, $r_{n_k-n_k^*(t)}$, as described
 436 by the filtration theory (FT) equations (24)-(26), can be calculated for any
 437 cluster k and $(d_p)_k$. For illustration purposes, the coefficient $r_{n_k-n_k^*(t)}$ was
 438 calculated as a function of $(d_p)_k$ for a collision efficiency $\alpha=0.0048$ [-], a
 439 collector grain diameter $d_c = 6 \times 10^{-4}$ m, two interstitial velocities ($U=0.2, 0.3$
 440 m/hr). Furthermore, the collision efficiency, $\alpha=0.0048$ [-], represents the
 441 average of multiple experimental values reported by Syngouna &
 442 Chrysikopoulos (2012). All other required parameter values are listed in Table
 443 1. Note that velocity effects are beyond the scope of this work, and only a
 444 narrow range of velocities are used in the simulations of this study ($U=0.2, 0.3$
 445 m/hr). The results are presented in Figure 2 and indicate that $r_{n_k-n_k^*(t)}$
 446 decreases to a minimum value at $(d_p)_1=850$ nm. Beyond this minimum the
 447 coefficient $r_{n_k-n_k^*(t)}$ increases monotonically with increasing $(d_p)_k$. Therefore,
 448 particles with $(d_p)_1 < 850$ nm are expected to exhibit reduction in the
 449 attachment rate with increasing particle diameter, whereas particles with

450 $(d_p)_1 > 850$ nm are expected to exhibit an increase in the attachment rate with
451 increasing particle diameter. Note that Figure 2 resembles the single-collector
452 efficiency plot reported by Yao et al. (1971), because the forward attachment
453 rate coefficient is linearly correlated with the single-collector efficiency (see
454 equations (24)-(26)).

455

456 4.3 Broadpulse source

457 The present nanoparticle model (equations (1)-(9), (11), (12), (14), (19))
458 accounting for combined reversible and irreversible attachment, assuming
459 diffusion-limited aggregation (DLA or fast aggregation) with successful
460 collisions calculated by use of the kernel b_{ij}^{DLA} (equation (15)), was applied to
461 the 1-D aquifer, assuming that nanoparticles with diameter $(d_p)_1 = 25$ nm enter
462 the aquifer at $x=0$ m, in a form of a broadpulse over the duration of $t_p = 28$ hr.
463 The forward reversible attachment rate for $k=1$ was set to $r_{n_1-n_1^*(r)} = 0.229$ 1/hr,
464 and irreversible attachment was neglected ($r_{n_k-n_k^*(i)} = 0$ 1/hr). The collision
465 efficiency was calculated as the average of multiple experimental values
466 reported by Syngouna & Chrysikopoulos (2012), $\alpha = 0.0048$ [-]. All other
467 required model parameter values were those listed in Table 1. In addition, the
468 model developed by Katzourakis and Chrysikopoulos (2015) (subsequently,
469 this biocolloid transport model will be referred to as “KC model”) was also
470 applied to the 1-D aquifer under the same conditions with the exception that
471 the attachment rate was assumed independent of aggregate size and equal to
472 $r_{n-n^*(r)} = 0.229$ [1/hr]. Note that the KC model describes the transport of
473 colloids in three-dimensional, water saturated, homogeneous porous media,
474 accounting for particle attachment onto the solid matrix by the two-site kinetic
475 model, without considering particle aggregation.

476 In Figure 3a-f are shown the dimensionless concentrations as simulated
477 by both the present nanoparticle transport model and the KC model, at three
478 different locations within the 1-D aquifer ($x=0.2, 0.35,$ and 0.6 m) as a function
479 of time (see Figures 3a-c), and at three different times ($t=3, 28,$ and 32 hr) as
480 a function of distance within the aquifer (see Figure 3d-f). The concentrations

481 simulated by the present nanoparticle transport model reach peak
 482 concentrations faster, and exhibit less pronounced tailing than the KC model
 483 (see Figure 3a-c). Also, the nanoparticle distribution (snapshots) within the 1-
 484 D aquifer as simulated by the present nanoparticle transport model is higher
 485 at early times (t=3 hr) and lower at late times (t=32 hr) compared to the KC
 486 model (see Figure 3d, and f). As the aggregate diameters increase the
 487 various attachment rates $r_{n_k - n_k^{*(r)}}$ decrease (there is a different attachment rate
 488 for each cluster). When the nanoparticle attachment rate is reduced, fewer
 489 nanoparticles are retained by the solid matrix of the aquifer. It should be noted
 490 that for the simulations in Figures 3a-f the aggregate diameters did not
 491 exceed $(d_p)_1 = 386$ nm.

492 The simulations presented in Figure 3a-f were repeated for the case
 493 where only irreversible attachment was accounted for. In the present
 494 nanoparticle transport model, the reversible attachment and detachment rates
 495 were set to zero ($r_{n_k^{*(r)} - n_k} = r_{n_k - n_k^{*(r)}} = 0$ 1/hr), and the irreversible attachment to
 496 $r_{n_1 - n_1^{*(i)}} = 0.229$ 1/hr. In the KC model, the reversible attachment and
 497 detachment rates were set to zero, and the irreversible attachment was set to
 498 $r_{n - n^{*(i)}} = 0.229$ 1/hr. The simulations for the case where only irreversible
 499 attachment was accounted for, are presented in Figure 3g-l. Note that the
 500 results from the simulations obtained by the two models are quite different.
 501 The present nanoparticle transport model consistently yielded dimensionless
 502 total number concentrations significantly higher than those of the KC model.
 503 This discrepancy is attributed to nanoparticle aggregation, which is accounted
 504 for in the present nanoparticle transport model. As nanoparticles aggregate,
 505 new clusters with larger aggregates are created. A different $r_{n_k - n_k^{*(i)}}$ rate is
 506 assigned to each cluster, with a value which is decreasing with increasing
 507 cluster number. The effect of aggregation is more pronounced when
 508 irreversible attachment is accounted for, than when reversible attachment is
 509 considered (compare Figures 3a-f and 3g-l). This observation suggests that
 510 the nanoparticle aggregation effect on transport could be masked when
 511 reversible attachment occurs. This is similar to the findings reported in the

512 literature that reversible attachment may conceal the effects of the
513 geochemical heterogeneity of an aquifer (Katzourakis & Chrysikopoulos,
514 2018).

515 The dimensionless average size of the suspended aggregates,
516 $\bar{d}_p / (d_p)_1$ [-], for the exact conditions examined in Figures 3a-f, are presented
517 in Figure 4. The trend of the $\bar{d}_p / (d_p)_1$ for the case where there is reversible
518 attachment, shown in Figures 4a-c are very similar to those shown for n_1^T / n_1^0
519 in Figures 3a-c. Clearly, the aggregate size is directly proportional to the
520 nanoparticle concentration. The ratio $\bar{d}_p / (d_p)_1$ increases considerably, up to a
521 seven-fold. The increase in $\bar{d}_p / (d_p)_1$ with distance along the 1-D aquifer
522 observed in Figures 4d-f is expected, because as the nanoparticles move
523 downstream they aggregate and consequently increase in size. A temporary
524 increase in $\bar{d}_p / (d_p)_1$ appears immediately after the broad pulse injection of
525 nanoparticles is completed ($t > t_p = 28$ hr, see Figures 4a-c), because particles
526 previously attached onto the solid matrix with size greater or equal to the
527 injected nanoparticles ($(d_p)_k \geq (d_p)_1$) are starting to detach. This increase in
528 $\bar{d}_p / (d_p)_1$ fades away with time as the nanoparticle concentration reduces
529 rapidly. For the case where irreversible attachment is considered and at times
530 $t > t_p$ the ratio $\bar{d}_p / (d_p)_1$ becomes negligible after a temporary sharp increase.
531 This is a consequence of the faster irreversible attachment of smaller sized
532 nanoparticles, which in turn leads to an increase in the average size of the
533 suspended aggregates. Note that for relatively small nanoparticles, the
534 attachment rate is inversely proportional to their aggregate size (see Figure
535 2). Also, the suspended nanoparticle number concentrations eventually
536 become negligible due to irreversible attachment. In contrast, for the case
537 where reversible attachment is considered, the reduction of smaller
538 aggregates is less pronounced because there are continuously detached.
539 This is the reason that at late times ($t = 32$ hr, Figure 4f) the ratio $\bar{d}_p / (d_p)_1$ is
540 substantially higher for the case where irreversible attachment is considered.

541

542 4.4 Instantaneous source

543 The present nanoparticle model with instantaneous source (equations
544 (1)-(8),(10),(11),(13),(14),(19)) and the KC model with instantaneous source,
545 assuming diffusion-limited aggregation (DLA or fast aggregation) with
546 successful collisions calculated by use of the kernel b_{ij}^{DLA} (equation (15)),
547 were used to simulate nanoparticle transport in the 1-D aquifer. Two different
548 nanoparticle diameters were considered: $(d_p)_1 = 25$ nm, and $(d_p)_1 = 850$ nm.
549 Each size of nanoparticles was examined separately. The nanoparticles were
550 introduced instantaneously in the aquifer at $x_0 = 0.10$ m. The number of
551 nanoparticles injected was $(N_{inj})_1 = 3 \times 10^{10}$ [np1] for both nanoparticle sizes.
552 For the present nanoparticle transport model the forward reversible
553 attachment rate for $k=1$ was $r_{n_1-n_1^*(t)} = 0.257$ 1/hr for $(d_p)_1 = 25$ nm, and
554 $r_{n_1-n_1^*(t)} = 0.0227$ 1/hr for $(d_p)_1 = 850$ nm (see Figure 2). For the KC model the
555 forward reversible attachment rate was set to $r_{n-n^*(t)} = 0.257$ 1/hr for $(d_p)_1 = 25$
556 nm and $r_{n-n^*(t)} = 0.0227$ 1/hr for $(d_p)_1 = 850$ nm. All other required model
557 parameters were those listed in Table 1. The model simulations are presented
558 in Figure 5. As expected, the total number concentrations (n_1^T) decrease with
559 increasing time and distance from the source location. For the smaller
560 nanoparticles ($(d_p)_1 = 25$ nm) the simulated n_1^T curves were higher for the
561 present transport model than the KC model (see Figures 5a,b). However, for
562 the larger nanoparticles ($(d_p)_1 = 850$ nm) the simulated n_1^T curves were lower
563 for the present transport model than the KC model (see Figures 5c,d). For
564 both nanoparticle sizes considered here, the difference between the n_1^T
565 curves simulated with the present transport model and the KC model,
566 increases with increasing time and distance. These observations are
567 attributed to the aggregate diameter increase, which is only accounted by the
568 present model. Note that for the smaller nanoparticles the attachment rate

569 $r_{n_k - n_k^{*(r)}}$ decreases as the aggregate diameter increases, while for the larger
 570 nanoparticles the opposite is true (see Figure 2). Therefore, when the mean of
 571 the various $r_{n_k - n_k^{*(r)}}$ values decreases, n_1^T increases and when the mean of the
 572 various $r_{n_k - n_k^{*(r)}}$ values increases, n_1^T decreases.

573 The dimensionless average size distributions of suspended aggregates,
 574 $\bar{d}_p / (d_p)_1$ [-], for the exact conditions examined in Figures 5, are presented in
 575 Figure 6. The $\bar{d}_p / (d_p)_1$ trend for the smaller nanoparticles ($(d_p)_1 = 25$ nm)
 576 follows the trend of n_1^T shown in Figures 5a,b. Positive n_1^T slopes lead to
 577 increasing $\bar{d}_p / (d_p)_1$ ratios and negative n_1^T slopes to decreasing $\bar{d}_p / (d_p)_1$
 578 ratios. However, upstream from the source location ($x_0 = 0.1$ m), the dashed
 579 curve in Figure 6e exhibits a dip (minimum), which is not observed in the
 580 corresponding n_1^T curve in Figure 5b. Near the source, the $(d_p)_1 = 25$ nm
 581 nanoparticles, which have diffused upstream, attach onto the solid matrix of
 582 the 1-D aquifer with greater attachment rate than the constantly forming larger
 583 aggregates (see Figure 2). When the suspended nanoparticles migrate away
 584 from the source, the attached smaller nanoparticles detach, and in turn
 585 contribute to the reduction of the $\bar{d}_p / (d_p)_1$ ratio, as shown by the dip in the
 586 dashed curve of Figure 6e. At a subsequent point in time ($t = 1.1$ hr), this dip is
 587 smoothed because the $\bar{d}_p / (d_p)_1$ ratio upstream from the source location is
 588 reduced due to the nanoparticle migration (see dashed curve in Figure 6f). It
 589 should be noted that at late times, the $\bar{d}_p / (d_p)_1$ trend of the nanoparticles with
 590 diameter $(d_p)_1 = 850$ nm (see Figure 6a-c) deviates significantly from the trend
 591 of n_1^T shown in Figures 5c,d. This is attributed to the increasing attachment as
 592 nanoparticles with $(d_p)_1 = 850$ nm form larger aggregates (see Figure 2). At
 593 late times, when n_1^T decreases due to nanoparticle transport and attachment
 594 onto the solid matrix of the porous medium, some large aggregates detach
 595 and contribute to the observed increase in $\bar{d}_p / (d_p)_1$. Note that the snapshots

596 for $(d_p)_1 = 850$ nm (solid curves) in Figures 6e,f exhibit two distinct peaks. The
 597 second peak further downstream is expected, because it follows the n_1^T trend
 598 (Figure 5d). However, the first peak, near the source location ($x_0 = 0.1$ m), is
 599 attributed to formation of larger aggregates with attachment rates that
 600 increase as their size increases (see Figure 2). These aggregates detach
 601 from the solid matrix after the main concentration peak migrates downstream.
 602

603 **4.5 Nanoparticle size-dependent dispersivity**

604 The hydrodynamic dispersion is an important transport parameter and
 605 for aggregating nanoparticles should not be considered as an invariant
 606 parameter, but different clusters should be assigned different values:

$$607 \quad (D_x)_k = \alpha_L U \quad (28)$$

608 where α_L [L] is the longitudinal dispersivity. As the size of nanoparticles
 609 increase their dispersivity is also increasing, because as the size of particles
 610 increases: (1) the particle effective porosity is reduced, and (2) particles are
 611 excluded from lower-velocity regions of the parabolic velocity profile within the
 612 pore throats (Chrysikopoulos & Katzourakis, 2015).

613 To illustrate the effect of size-dependent dispersivity the simulations
 614 presented in Figure 5c were repeated under the exact same conditions with
 615 only one difference, the present nanoparticle transport model was modified to
 616 account for size-dependent dispersivity. It was assumed that aggregate
 617 dispersivity is increasing with particle diameter based on the following
 618 empirical relationship (Chrysikopoulos & Katzourakis, 2015):

$$619 \quad \alpha_L [\text{cm}] = 0.29 + 5.06 \times 10^{-5} d_p [\text{nm}] \quad (29)$$

620 In present nanoparticle transport model the dispersion was estimated by
 621 equations (28) and (29) (i.e. $k=1$, $(d_p)_1 = 850$ nm, $(D_x)_1 = 1 \times 10^{-3}$ m/hr²),
 622 whereas in the KC model the dispersion coefficient was set to $D_x = 1 \times 10^{-3}$
 623 m/hr². The simulated number concentration of suspended nanoparticle, n_1^T ,
 624 breakthrough curves are presented in Figure 7. It is shown that simulations
 625 conducted with the present nanoparticle transport model, which accounts for
 626 size-dependent dispersivity exhibit early breakthrough, more spreading,
 627 extended tailing, and lower concentrations compared to the KC model. This

628 result is expected, because formation of aggregates with progressively
629 increasing diameter size result in increasing dispersion coefficients.

630

631 **4.6 Comparison between DLA and RLA**

632 The simulations presented in Figure 5c, under the assumption of
633 diffusion-limited aggregation (DLA or fast aggregation), were repeated for the
634 exact same conditions, but assuming reaction-limited aggregation (RLA or
635 slow aggregation) with successful collisions determined by use of the kernel
636 b_{ij}^{RLA} (equation (16)). For the RLA simulations, the surface potential of the
637 particles of cluster $k=1$ containing nanoparticles with diameter $(d_p)_1=850$ nm
638 was set to $\Psi_{p1}=8.7$ [mV]. Also, the dispersivity was assumed to be invariant
639 with aggregate size.

640 The simulated breakthrough curves of the total number concentration of
641 suspended nanoparticles, n_1^T , obtained by the present model assuming RLA
642 are presented in Figure 8, together with the corresponding breakthrough
643 curves obtained by the present model assuming DLA, and the KC model.
644 Clearly, the breakthrough curves simulated under the assumption of RLA are
645 higher than those simulated under the assumption of DLA, but lower than
646 those obtained by the KC model. This is an expected result because fewer
647 aggregates are formed with RLA than DLA, and the KC model neglects
648 aggregation. Note that for nanoparticles with diameter $(d_p)_1=850$ nm
649 aggregate formation leads to higher attachment rates (see Figure 2).

650

651

652

653

654

655

656

657

658

659

660 **4.7 Impact of fractal dimension D_F on nanoparticle transport**

661 To further investigate the effect of nanoparticle aggregation on
662 nanoparticle transport, the simulations presented in Figure 5c for diffusion-
663 limited aggregation (DLA or fast aggregation), were repeated for different
664 fractal dimension values (D_F). The number of nanoparticles injected was
665 $(N_{inj})_1 = 3 \times 10^{10}$ [np1] with diameter $(d_p)_1 = 850$ nm. The results are presented
666 in Figure 9.

667 It is evident from Figure 9 that as the value of D_F decreases the
668 average concentration decreases as well. This is expected because larger D_F
669 values correspond to smaller cluster diameters (see equation (19)), which in
670 turn leads to smaller average attachment rates (see Figure 2). Therefore,
671 smaller D_F values yield higher attachment rates and smaller concentrations.
672 Note that the KC model concentrations can differ from the current model
673 concentrations up to an order of magnitude. Consequently, the effects of
674 aggregation cannot be overlooked.

675

676 **4.8 Comparison to other studies**

677 The results presented in this work are in agreement with other studies
678 published in the literature. Raychoudhury et al. (2012) performed various
679 nanoparticle transport experiments in columns packed with sand, and pointed
680 out that the particle single collector contact efficiency changes with particle
681 diameter. It was reported that initially the increasing particle size led to
682 decreasing collector efficiency; subsequently, as the particle size increased
683 further, the collector efficiency increased, following a trend similar to the one
684 shown in Figure 2. Using this relationship between particle size and collector
685 efficiency, model simulations with the Smoluchowski equation were
686 performed, which indicated, as in the present study (Figure 3c,i), that
687 breakthrough concentrations of small aggregating particles were higher than
688 non-aggregating particles. Also, Taghavy et al. (2015) obtained the same
689 result by developing a Lagrangian model that accounted for aggregation and
690 incorporated the population balance equation (2). Contrarily, Babakhani
691 (2019) reported that for a specific size range of nanoparticles when
692 aggregation was accounted for, the breakthrough concentration decreased

693 (as also shown in Figure 5c). The differences in the results presented by the
694 various authors are caused by the attachment behaviour, because an
695 increase in nanoparticle aggregate size may lead to either increased or
696 decreased attachment (see Figure 2). Finally, despite some differences in the
697 modelling of the attachment process (kinetic, equilibrium, DLVO interactions),
698 all of these studies concluded that aggregation can change the average
699 attachment rate and in turn can affect the mobility of nanoparticles, as
700 reported in this work.

701

702

703 **5. Summary and conclusions**

704 The novel nanoparticle transport model was developed in this work
705 accounts for advection, dispersion, reversible and irreversible attachment, and
706 aggregation. Both DLA and RLA conditions were considered. For the
707 numerical solution, the transport and attachment processes were decoupled
708 from the aggregation process using an adaptive splitting operator method and
709 then were solved separately. The results from numerous simulations
710 suggested that nanoparticle aggregation affects significantly nanoparticle
711 transport in porous media. It was shown that due to aggregation the size of
712 nanoparticles increases, which in turn can lead to an increased or decreased
713 average attachment rate, depending on the initial particle diameter. An
714 increase in average attachment causes late breakthrough, while a decrease
715 yields early breakthrough. Particle size-dependant dispersivity enhances
716 spreading and leads to early breakthrough of nanoparticles. The effect of
717 nanoparticle aggregation was more pronounced for irreversible than
718 reversible attachment. Also, it was shown that the effects of aggregation were
719 more significant under DLA than RLA conditions. The discrepancies between
720 the transport with and without aggregation varied in time and space and were
721 more evident as the evolution of aggregation progressed further. Therefore,
722 for the simulation of nanoparticle transport in porous media, neglecting to
723 account for aggregation, particle-size dependent dispersivity or particle
724 surface charges, can lead to erroneous and unrealistic results.

725

726 **Acknowledgments:** This research has received funding from the Partnership
727 for Research and Innovation in the Mediterranean Area (PRIMA), under grant
728 agreement number: 1923-InTheMED. FAIR data policy statement: All figures
729 and tables can be directly reproduced from the equations presented in this
730 manuscript.

731

732 6. References

- 733 Arosio, P., Rima, S., Lattuada, M., & Morbidelli, M. (2012). Population balance modeling of
734 antibodies aggregation kinetics. *The Journal of Physical Chemistry B*, 116(24), 7066–
735 7075. <https://doi.org/10.1021/jp301091n>
- 736 Axford, S. D. (1997). Aggregation of colloidal silica: Reaction-limited kernel, stability ratio and
737 distribution moments. *J. Chem. Soc., Faraday Trans.*, 93(2), 303–311.
738 <https://doi.org/10.1039/A606195H>
- 739 Babakhani, P. (2019). The impact of nanoparticle aggregation on their size exclusion during
740 transport in porous media: One-and three-dimensional modelling investigations.
741 *Scientific Reports*, 9(1), 1–12. <https://doi.org/10.1038/s41598-019-50493-6>
- 742 Babakhani, P., Bridge, J., Phenrat, T., Fagerlund, F., Doong, R., & Whittle, K. R. (2019).
743 Comparison of a new mass-concentration, chain-reaction model with the population-
744 balance model for early-and late-stage aggregation of shattered graphene oxide
745 nanoparticles. *Colloids and Surfaces A: Physicochemical and Engineering Aspects*, 582,
746 123862. <https://doi.org/10.1016/j.colsurfa.2019.123862>
- 747 Babakhani, P., Doong, R., & Bridge, J. (2018). Significance of early and late stages of
748 coupled aggregation and sedimentation in the fate of nanoparticles: Measurement and
749 modeling. *Environmental Science & Technology*, 52(15), 8419–8428.
750 <https://doi.org/10.1021/acs.est.7b05236>
- 751 Barry, D., Bajracharya, K., Crapper, M., Prommer, H., & Cunningham, C. J. (2000).
752 Comparison of split-operator methods for solving coupled chemical non-equilibrium
753 reaction/groundwater transport models. *Mathematics and Computers in Simulation*,
754 53(1–2), 113–127. [https://doi.org/10.1016/S0378-4754\(00\)00182-8](https://doi.org/10.1016/S0378-4754(00)00182-8)
- 755 Bedrikovetsky, P., Siqueira, F. D., Furtado, C. A., & Souza, A. L. S. (2011). Modified particle
756 detachment model for colloidal transport in porous media. *Transport in Porous Media*,
757 86(2), 353–383. <https://doi.org/10.1007/s11242-010-9626-4>
- 758 Bedrikovetsky, P., Zeinijahromi, A., Siqueira, F. D., Furtado, C. A., & de Souza, A. L. S.
759 (2012). Particle detachment under velocity alternation during suspension transport in
760 porous media. *Transport in Porous Media*, 91(1), 173–197.
761 <https://doi.org/10.1007/s11242-011-9839-1>
- 762 Benn, T., & Westerhoff, P. (2008). Nanoparticle silver released into water from commercially
763 available sock fabrics. *Environmental Science and Technology*, 42(11), 4133–4139.
764 <https://doi.org/10.1021/es7032718>

765 Botchev, M., Faragó, I., & Havasi, Á. (2004). Testing weighted splitting schemes on a one-
766 column transport-chemistry model. *International Journal of Environment and Pollution*,
767 22(1–2), 3–16. <https://doi.org/10.1504/ijep.2004.005473>

768 Brar, S., Verma, M., Tyagi, R. D., & Surampalli, R. Y. (2010). Engineered nanoparticles in
769 wastewater and wastewater sludge - Evidence and impacts. *Waste Management*, 30(3),
770 504-520. <https://doi.org/10.1016/j.wasman.2009.10.012>

771 Chatterjee, J., & Gupta, S. K. (2009). An agglomeration-based model for colloid filtration.
772 *Environmental Science and Technology*, 43(10), 3694–3699.
773 <https://doi.org/10.1021/es8029973>

774 Chen, G., Liu, X., & Su, C. (2011). Transport and retention of TiO₂ rutile nanoparticles in
775 saturated porous media under low-ionic-strength conditions: Measurements and
776 mechanisms. *Langmuir*, 27(9), 5393–5402. <https://doi.org/10.1021/la200251v>

777 Chowdhury, I., Hong, Y., Honda, R. J., & Walker, S. L. (2011). Mechanisms of TiO₂
778 nanoparticle transport in porous media: Role of solution chemistry, nanoparticle
779 concentration, and flowrate. *Journal of Colloid and Interface Science*, 360(2), 548–555.
780 <https://doi.org/10.1016/j.jcis.2011.04.111>

781 Chrysikopoulos, C. V., & Katzourakis, V. E. (2015). Colloid particle size-dependent
782 dispersivity. *Water Resources Research*, 51(6), 4668–4683.
783 <https://doi.org/10.1002/2014WR016094>

784 Chrysikopoulos, C. V., & Syngouna, V. I. (2012). Attachment of bacteriophages MS2 and
785 ΦX174 onto kaolinite and montmorillonite: Extended-DLVO interactions. *Colloids and*
786 *Surfaces B: Biointerfaces*, 92, 74–83. <https://doi.org/10.1016/j.colsurfb.2011.11.028>

787 Compère, F., Porel, G., & Delay, F. (2001). Transport and retention of clay particles in
788 saturated porous media. Influence of ionic strength and pore velocity. *Journal of*
789 *Contaminant Hydrology*, 49(1–2), 1–21. [https://doi.org/10.1016/S0169-7722\(00\)00184-4](https://doi.org/10.1016/S0169-7722(00)00184-4)

790 Elimelech, M., & O'Melia, C. R. (1990). Kinetics of Deposition of Colloidal Particles in Porous
791 Media. *Environmental Science and Technology*, 24(10), 1528–1536.
792 <https://doi.org/10.1021/es00080a012>

793 Family, F., Meakin, P., & Vicsek, T. (1985). Cluster size distribution in chemically controlled
794 cluster-cluster aggregation. *The Journal of Chemical Physics*, 83(8), 4144–
795 4150. <https://doi.org/10.1063/1.449079>

796 Fang, J., Shan, X. Q., Wen, B., Lin, J. Ming, & Owens, G. (2009). Stability of titania
797 nanoparticles in soil suspensions and transport in saturated homogeneous soil columns.
798 *Environmental Pollution*, 157(4), 1101–1109.
799 <https://doi.org/10.1016/j.envpol.2008.11.006>

800 Feder, J. (1988). *Fractals* New York. Plenum Press.

801 Feke, D. L., Prabhu, N. D., Mann, J. A., & Mann, J. A. (1984). A formulation of the short-range
802 repulsion between spherical colloidal particles. *Journal of Physical Chemistry*, 88(23),
803 5735–5739. <https://doi.org/10.1021/j150667a055>

804 Fuchs, N. (1934). About the stability and loading of aerosols. *Journal of Physics*, 89(11–12),

805 736–743.

806 Gaudreault, R., Di Cesare, N., Van De Ven, T. G. M., & Weitz, D. A. (2015). Structure and
807 Strength of Flocs of Precipitated Calcium Carbonate Induced by Various Polymers Used
808 in Papermaking. *Industrial and Engineering Chemistry Research*, *54*(24), 6234–6246.
809 <https://doi.org/10.1021/acs.iecr.5b00818>

810 Godinez, I. G., & Darnault, C. J. G. (2011). Aggregation and transport of nano-TiO₂ in
811 saturated porous media: Effects of pH, surfactants and flow velocity. *Water Research*,
812 *45*(2), 839–851. <https://doi.org/10.1016/j.watres.2010.09.013>

813 Goldberg, E., Scheringer, M., Bucheli, T. D., & Hungerbühler, K. (2014). Critical Assessment
814 of Models for Transport of Engineered Nanoparticles in Saturated Porous Media.
815 *Environmental Science & Technology*, *48*(21), 12732–12741.
816 <https://doi.org/10.1021/es502044k>

817 Gottschalk, F., Sonderer, T., Scholz, R. W., & Nowack, B. (2009). Modeled environmental
818 concentrations of engineered nanomaterials (TiO₂, ZnO, Ag, CNT, fullerenes) for
819 different regions. *Environmental Science and Technology*, *43*(24), 9216–9222.
820 <https://doi.org/10.1021/es9015553>

821 Gregory, J. (1981). Approximate expressions for retarded van der Waals interaction. *Journal*
822 *of Colloid and Interface Science*, *83*(1), 138–145. [https://doi.org/10.1016/0021-](https://doi.org/10.1016/0021-9797(81)90018-7)
823 [9797\(81\)90018-7](https://doi.org/10.1016/0021-9797(81)90018-7).

824 Heidmann, I. (2013). Metal oxide nanoparticle transport in porous media – an analysis about
825 (un)certainities in environmental research. *Journal of Physics: Conference Series*, *429*,
826 012042. <https://doi.org/10.1088/1742-6596/429/1/012042>

827 Hogg, R., Healy, T. W., & Fuerstenau, D. W. (1966). Mutual coagulation of colloidal
828 dispersions. *Transactions of the Faraday Society*, *62*, 1638–1651.
829 <https://doi.org/10.1039/tf9666201638>

830 IARC (2010) Monographs on the evaluation of carcinogenic risks to humans, v. 93, World
831 Health Organization, 2010

832 Kanney, J. F., Miller, C. T., & Kelley, C. T. (2003). Convergence of iterative split-operator
833 approaches for approximating nonlinear reactive problems. *Advances in Water*
834 *Resources*, *26*(3), 247–261. [https://doi.org/10.1016/S0309-1708\(02\)00162-8](https://doi.org/10.1016/S0309-1708(02)00162-8)

835 Katzourakis, V. E., & Chrysikopoulos, C. V. (2014). Mathematical modeling of colloid and
836 virus cotransport in porous media: Application to experimental data. *Advances in Water*
837 *Resources*, *68*, 62–73. <https://doi.org/10.1016/j.advwatres.2014.03.001>

838 Katzourakis, V. E., & Chrysikopoulos, C. V. (2015). Modeling dense-colloid and virus
839 cotransport in three-dimensional porous media. *Journal of Contaminant Hydrology*, *181*,
840 102–113. <https://doi.org/10.1016/j.jconhyd.2015.05.010>

841 Katzourakis, V. E., & Chrysikopoulos, C. V. (2018). Impact of Spatially Variable Collision
842 Efficiency on the Transport of Biocolloids in Geochemically Heterogeneous Porous
843 Media. *Water Resources Research*, *54*(6), 3841–3862.
844 <https://doi.org/10.1029/2017WR021996>

845 Kinzelbach, W., Schäfer, W., & Herzer, J. (1991). Numerical modeling of natural and
846 enhanced denitrification processes in aquifers. *Water Resources Research*, 27(6),
847 1123–1135. <https://doi.org/10.1029/91WR00474>

848 Lattuada, M., Sandkühler, P., Wu, H., Sefcik, J., & Morbidelli, M. (2003). Aggregation kinetics
849 of polymer colloids in reaction limited regime: experiments and simulations. *Advances in*
850 *Colloid and Interface Science*, 103(1), 33–56. [https://doi.org/10.1016/S0001-
851 *8686\(02\)00082-9*](https://doi.org/10.1016/S0001-8686(02)00082-9)

852 Lee, D. G., Bonner, J. S., Garton, L. S., Ernest, A. N. S., & Autenrieth, R. L. (2000). Modeling
853 coagulation kinetics incorporating fractal theories: A fractal rectilinear approach. *Water*
854 *Research*, 34(7), 1987–2000. [https://doi.org/10.1016/S0043-1354\(99\)00354-1](https://doi.org/10.1016/S0043-1354(99)00354-1)

855 Lin, M. Y., Klein, R., Lindsay, H. M., Weitz, D. A., Ball, R. C., & Meakin, P. (1990). The
856 structure of fractal colloidal aggregates of finite extent. *Journal of Colloid And Interface*
857 *Science*, 137(1), 263–280. [https://doi.org/10.1016/0021-9797\(90\)90061-R](https://doi.org/10.1016/0021-9797(90)90061-R)

858 Lin, M. Y., Lindsay, H. M., Weitz, D. A., Ball, R. C., Klein, R., & Meakin, P. (1989).
859 Universality in colloid aggregation. *Nature*, 339(6223), 360–362.
860 <https://doi.org/10.1038/339360a0>

861 Litton, G. M., & Olson, T. M. (1996). Particle size effects on colloid deposition kinetics:
862 Evidence of secondary minimum deposition. *Colloids and Surfaces A: Physicochemical*
863 *and Engineering Aspects*, 107, 273–283. [https://doi.org/10.1016/0927-7757\(95\)03343-2](https://doi.org/10.1016/0927-7757(95)03343-2)

864 Liu, H. H., Surawanvijit, S., Rallo, R., Orkoulas, G., & Cohen, Y. (2011). Analysis of
865 nanoparticle agglomeration in aqueous suspensions via constant-number Monte Carlo
866 simulation. *Environmental Science and Technology*, 45(21), 9284–9292.
867 <https://doi.org/10.1021/es202134p>

868 Loveland, J. P., Ryan, J. N., Amy, G. L., & Harvey, R. W. (1996). The reversibility of virus
869 attachment to mineral surfaces. In *Colloids and Surfaces A: Physicochemical and*
870 *Engineering Aspects*, 107, 205–221. [https://doi.org/10.1016/0927-7757\(95\)03373-4](https://doi.org/10.1016/0927-7757(95)03373-4)

871 Marray, J. P., & Parks, G. A. (1978). *Particulates in Water: Characterization, Fate, Effects and*
872 *Removal*, edited by MC Kavanaugh and JO Leckie. Adv. Chem. Ser., Amican Chemical
873 Society, Washington, DC, USA, 189.

874 Mueller, N. C., & Nowack, B. (2008). Exposure modelling of engineered nanoparticles in the
875 environment. *Environmental Science & Technology*, 42(12), 44447–53.
876 <https://doi.org/10.1021/es7029637>

877 Nicoud, L., Arosio, P., Sozo, M., Yates, A., Norrant, E., & Morbidelli, M. (2014). Kinetic
878 analysis of the multistep aggregation mechanism of monoclonal antibodies. *Journal of*
879 *Physical Chemistry B*, 118(36), 10595–10606. <https://doi.org/10.1021/jp505295j>

880 Nowack, B., & Bucheli, T. D. (2007). Occurrence, behavior and effects of nanoparticles in the
881 environment. *Environmental pollution*, 150(1), 5-22.
882 <https://doi.org/10.1016/j.envpol.2007.06.006>

883 Petosa, A. R., Jaisi, D. P., Quevedo, I. R., Elimelech, M., & Tufenkji, N. (2010). Aggregation
884 and deposition of engineered nanomaterials in aquatic environments: role of

885 physicochemical interactions. *Environmental Science & Technology*, 44(17), 6532–
886 6549. <https://doi.org/10.1021/es100598h>

887 Quik, J.T., de Klein, J.J. and Koelmans, A.A., (2015). Spatially explicit fate modelling of
888 nanomaterials in natural waters. *Water research*, 80, .200-
889 208. <http://dx.doi.org/10.1016/j.watres.2015.05.025>

890 Rajagopalan, R., & Tien, C. (1976). Trajectory analysis of deep-bed filtration with the
891 sphere-in-cell porous media model. *AIChE Journal*, 22(3), 523–533.
892 <https://doi.org/10.1002/aic.690220316>

893 Raychoudhury, T., Tufenkji, N., & Ghoshal, S. (2012). Aggregation and deposition kinetics of
894 carboxymethyl cellulose-modified zero-valent iron nanoparticles in porous media. *Water*
895 *research*, 46(6), 1735-1744.

896 Reerink, H., & Overbeek, J. T. G. (1954). The rate of coagulation as a measure of the stability
897 of silver iodide sols. *Discussions of the Faraday Society*, 18, 74.
898 <https://doi.org/10.1039/df9541800074>

899 Ruckenstein, E., & Prieve, D. C. (1976). Adsorption and desorption of particles and their
900 chromatographic separation. *AIChE Journal*, 22(2), 276–283.
901 <https://doi.org/10.1002/aic.690220209>

902 Russell, T., & Bedrikovetsky, P. (2018). Colloidal-suspension flows with delayed fines
903 detachment: Analytical model & laboratory study. *Chemical Engineering Science*, 190,
904 98–109. <https://doi.org/10.1016/j.ces.2018.05.062>

905 Ryan, J. N., & Gschwend, P. M. (1994). Effects of ionic strength and flow rate on colloid
906 release: Relating kinetics to intersurface potential energy. *Journal of Colloid and*
907 *Interface Science*, 164, 21–34. <https://doi.org/10.1006/jcis.1994.1139>

908 Sabelfeld, K., & Kolodko, A. (2002). Stochastic Lagrangian models and algorithms for
909 spatially inhomogeneous Smoluchowski equation. *Mathematics and Computers in*
910 *Simulation*, 61(2), 115–137. [https://doi.org/10.1016/S0378-4754\(02\)00141-6](https://doi.org/10.1016/S0378-4754(02)00141-6)

911 Sandkühler, P., Sefcik, J., & Morbidelli, M. (2004). Kinetics of aggregation and gel formation
912 in concentrated polystyrene colloids. *The Journal of Physical Chemistry B*, 108(52),
913 20105–20121. <https://doi.org/10.1021/jp046468w>

914 Shamir, U. Y., & Harleman, D. R. F. (1967). Numerical solutions for dispersion in porous
915 mediums. *Water Resources Research*, 3(2), 557–581.
916 <https://doi.org/10.1029/WR003i002p00557>

917 Sim, Y., & Chrysikopoulos, C. (1998). Three-dimensional analytical models for virus transport
918 in saturated porous media. *Transport in Porous Media*, 30(1), 87–112.
919 <https://doi.org/10.1023/A:1006596412177>

920 Sim, Y., & Chrysikopoulos, C. V. (1995). Analytical Models for One-Dimensional Virus
921 Transport in Saturated Porous Media. *Water Resources Research*, 31(5), 1429–1437.
922 <https://doi.org/10.1029/95WR00199>

923 Sim, Y., & Chrysikopoulos, C. V. (1999). Analytical solutions for solute transport in saturated
924 porous media with semi-infinite or finite thickness. *Advances in Water Resources*, 22(5),

925 507–519. [https://doi.org/10.1016/S0309-1708\(98\)00027-X](https://doi.org/10.1016/S0309-1708(98)00027-X)

926 Smoluchowski, M. V. (1916). Über Brownsche Molekularbewegung unter Einwirkung äußerer
927 Kräfte und deren Zusammenhang mit der verallgemeinerten Diffusionsgleichung.
928 *Annalen Der Physik*, 353(24), 1103–1112. <https://doi.org/10.1002/andp.19163532408>

929 Smoluchowski, M. V. (1917). Versuch einer mathematischen theorie der koagulation kinetic
930 kolloider losungen. *Zeit. Phys. Chem*, 92, 129–168.

931 Smoluchowski, Marian. (1916). Drei Vortrage uber Diffusion. Brownsche Bewegung und
932 Koagulation von Kolloidteilchen. *Z. Phys.*

933 Solovitch, N., Labille, J., Rose, J., Chaurand, P., Borschneck, D., Wiesner, M. R., & Bottero,
934 J. Y. (2010). Concurrent aggregation and deposition of TiO₂ nanoparticles in a sandy
935 porous media. *Environmental Science and Technology*, 44(13), 4897–4902.
936 <https://doi.org/10.1021/es1000819>

937 Steefel, C. I., & MacQuarrie, K. T. B. (1996). Approaches to modeling of reactive transport in
938 porous media. *Reviews in Mineralogy*, 34(1), 83-129.

939 Syngouna, V. I., & Chrysikopoulos, C. V. (2012). Transport of biocolloids in water saturated
940 columns packed with sand: Effect of grain size and pore water velocity. *Journal of*
941 *Contaminant Hydrology*, 129–130, 11–24. <https://doi.org/10.1016/j.jconhyd.2012.01.010>

942 Syngouna, V. I., & Chrysikopoulos, C. V. (2013). Cotransport of clay colloids and viruses in
943 water saturated porous media. *Colloids and Surfaces A: Physicochemical and*
944 *Engineering Aspects*, 416(1), 56–65. <https://doi.org/10.1016/j.colsurfa.2012.10.018>

945 Taghavy, A., Pennell, K. D., & Abriola, L. M. (2015). Modeling coupled nanoparticle
946 aggregation and transport in porous media: A Lagrangian approach. *Journal of*
947 *contaminant hydrology*, 172, 48-60.

948 Tufenkji, N., & Elimelech, M. (2004). Correlation Equation for Predicting Single-Collector
949 Efficiency in Physicochemical Filtration in Saturated Porous Media. *Environmental*
950 *Science and Technology*, 38(2), 529–536. <https://doi.org/10.1021/es034049r>

951 Vasiliadou, I. A., & Chrysikopoulos, C. V. (2011). Cotransport of *Pseudomonas putida* and
952 kaolinite particles through water-saturated columns packed with glass beads. *Water*
953 *Resources Research*, 47(2), W02543. <https://doi.org/10.1029/2010WR009560>

954 Wang, M., Gao, B., Tang, D., & Yu, C. (2018). Concurrent aggregation and transport of
955 graphene oxide in saturated porous media: roles of temperature, cation type, and
956 electrolyte concentration. *Environmental Pollution*, 235, 350–
957 357. <https://doi.org/10.1016/j.envpol.2017.12.063>

958 Weast, R. C. (1984). *Handbook of Chemistry and Physics 64th Edition 1983-1984*. Boca
959 Raton, FL, USA: CRC Press.

960 Weitz, D. A., & Lin, M. Y. (1986). Dynamic Scaling of Cluster-Mass Distributions in Kinetic
961 Colloid Aggregation. *Physical Review Letters*, 57(16), 2037–2040.
962 <https://doi.org/10.1103/PhysRevLett.57.2037>

963 Weitz, D. A., Lin, M. Y., & Lindsay, H. M. (1991). Universality laws in coagulation.
964 *Chemometrics and Intelligent Laboratory Systems*, 10(1–2), 133–140.

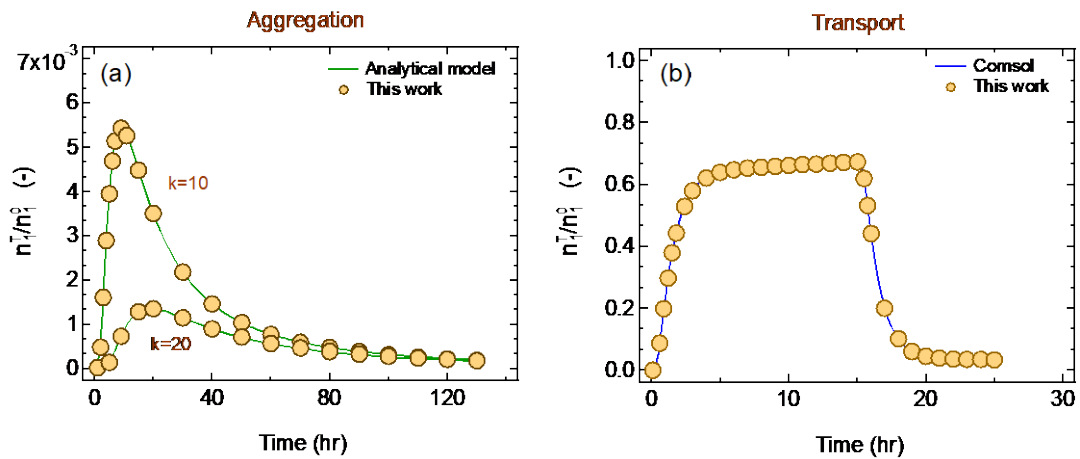
965 [https://doi.org/10.1016/0169-7439\(91\)80043-P](https://doi.org/10.1016/0169-7439(91)80043-P)
966 Wiesner, M. R., Lowry, G. V., Alvarez, P., Dionysiou, D., & Biswas, P. (2006). Assessing the
967 risks of manufactured nanomaterials. *Environmental Science and Technology*, 40(14),
968 4336-4345. <https://doi.org/10.1021/es062726m>
969 Wijnen, P. W. J. G., Beelen, T. P. M., Rummens, C. P. J., & van Santen, R. A. (1991).
970 Diffusion- and reaction-limited aggregation of aqueous silicate solutions. *Journal of Non-*
971 *Crystalline Solids*, 136(1–2), 119–125. [https://doi.org/10.1016/0022-3093\(91\)90127-R](https://doi.org/10.1016/0022-3093(91)90127-R)
972 Wood, T. M., & Baptista, A. M. (1993). A model for diagnostic analysis of estuarine
973 geochemistry. *Water Resources Research*, 29(1), 51–71.
974 <https://doi.org/10.1029/92WR02126>
975 Yao, K., Habibian, M. T., & O'Melia, C. R. (1971). Water and Waste Water Filtration:
976 Concepts and Applications. *Environmental Science & Technology*, 5(11), 1105–1112.
977 <https://doi.org/10.1021/es60058a005>
978
979

Table 1. Model parameters

Broadpulse Simulations			Instantaneous Simulations		
Parameter	Value (units)	Reference	Parameter	Value (units)	Reference
$(D_x)_k^a$	9×10^{-3} (m ² /hr)	–	U	0.3 (m/hr)	(Chrysikopoulos & Katzourakis, 2015)
U	0.2 (m/hr)	(Syngouna & Chrysikopoulos, 2013)	$(N_{in})_1$	3×10^{10} (np ₁)	–
t_p	28 (hr)	–	$r_{n_k^{*(r)}-n_k}$	0.03 (1/hr)	(Vasiliadou & Chrysikopoulos, 2011)
$(d_p)_1$	25×10^{-9} (m)	–	$r_{n_k^{*(i)}-n_k}$	0 (1/hr)	–
n_1^b	1×10^{15} (np ₁ /m ³)	–	X ₀	0.1 (m)	–
<i>With reversible attachment</i>			<i>Instantaneous simulations (d₁=25 nm)</i>		
$r_{n_1-n_1^{*(r)}}^c$	0.229 (1/hr)	Equation (24)	$(D_x)_k$	8×74^{-4} (m ² /hr)	Equation (28)
$r_{n_k-n_k^{*(i)}}^d$	0 (1/hr)	–	$(d_p)_1$	25×10^{-9} (m)	–
$r_{n_k^{*(r)}-n_k}^e$	0.3 (1/hr)	(Vasiliadou & Chrysikopoulos, 2011)	$r_{n-n^{*(r)}}$	0.257 (1/hr)	Equation (24)
$r_{n-n^{*(r)}}^f$	0.229 (1/hr)	Equation (24)	$r_{n_1-n_1^{*(r)}}$	0.257 (1/hr)	Equation (24)
<i>With irreversible attachment</i>			<i>Instantaneous simulations (d₁=850 nm)</i>		
$r_{n_k-n_k^{*(r)}}^g$	0 (1/hr)	–	$(D_x)_k$	1×10^{-3} (m ² /hr)	Equation (28)
$r_{n_1-n_1^{*(i)}}^h$	0.229 (1/hr)	–	$(d_p)_1$	850×10^{-9} (m)	–
$r_{n_k^{*(r)}-n_k}^i$	0 (1/hr)	–	$r_{n-n^{*(r)}}$	0.0227 (1/hr)	Equation (24)
$r_{n-n^{*(i)}}^j$	0.229 (1/hr)	–	$r_{n_1-n_1^{*(r)}}$	0.0227 (1/hr)	Equation (24)
Common physicochemical parameters			Common physicochemical parameters		
k_B^j	1.78×10^{-16} (kg m ² /(hr ² K))	(Weast, 1984)	L _x	0.6 (m)	–
A ₁₂₃ ^k	9.72×10^{-14} (kg m ² /hr ²)	(Murray & Parks, 1978)	α ^r	0.0048 (–)	(Syngouna & Chrysikopoulos, 2012)
A ₁₂₁ ^l	9.72×10^{-14} (kg m ² /hr ²)	–	g ^s	1.271×10^8 (m/hr ²)	–
d _c	6×10^{-4} (m)	–	θ	0.42 (–)	(Syngouna & Chrysikopoulos, 2012)
ρ _w ^m	999.7 (kg/m ³)	–	DLVO		
μ _w ⁿ	3.2 (kg/(m hr))	–	l _s ^t	0.1 (mol/m ³)	(Chrysikopoulos & Syngouna, 2012)
T ^o	298 (K)	–	N _A ^u	6.022×10^{23} (1/mol)	(Weast, 1984)
A _c	4.91×10^{-4} (m ²)	–	e ^v	1.602×10^{-19} (C)	(Weast, 1984)
ζ	0.637 (–)	(Feder, 1988)	ε ₀ ^w	8.854×10^{-12} (C ² /(J m))	(Weast, 1984)
D _F	2.1 (–)	(Lin et al., 1989)	ε _r ^x	78.4 (–)	(Weast, 1984)
ρ _n ^p	1420 (kg/m ³)	–	Ψ _{p1} ^y	8.7 (mv) [d ₁ =850 nm]	–
ρ _b ^q	1610 (kg/m ³)	–	σ _{Born} ^z	5×10^{-10} (m)	(Ruckenstein & Prieve, 1976)

- 981
- 982 ^a $(D_x)_k$ [L^2/t] is longitudinal hydrodynamic dispersion coefficient of suspended nanoparticles that belong to cluster k.
- 983 ^b n_1^0 [np_k] is the initial constant aqueous phase concentration of first cluster, used in (12).
- 984 ^c $r_{n_1-n_1}^{s(i)}$ [$1/t$] rate coefficient of reversible nanoparticle attachment onto the solid matrix, of first cluster k=1.
- 985 ^d $r_{n_k-n_k}^{s(i)}$ [$1/t$] rate coefficient of irreversible nanoparticle attachment onto the solid matrix, that belong to cluster k.
- 986 ^e $r_{n_k-n_k}^{s(r)}$ [$1/t$] rate coefficient of reversible nanoparticle detachment from the solid matrix, that belong to cluster k.
- 987 ^f $r_{n-n}^{s(i)}$ [$1/t$] rate coefficient of reversible nanoparticle attachment onto the solid matrix used by the KC model.
- 988 ^g $r_{n_k-n_k}^{s(i)}$ [$1/t$] rate coefficient of reversible nanoparticle attachment onto the solid matrix, that belong to cluster k.
- 989 ^h $r_{n_1-n_1}^{s(i)}$ [$1/t$] rate coefficient of irreversible nanoparticle attachment onto the solid matrix, of first cluster k=1.
- 990 ⁱ $r_{n-n}^{s(i)}$ [$1/t$] rate coefficient of irreversible nanoparticle attachment onto the solid matrix used by the KC model.
- 991 ^k_B [$M \cdot L^2/(t^2 \cdot T)$] Boltzmann constant), used in (21).
- 992 ^k_{A123} [ML^2/t^2] Complex Hamaker constant (nanoparticle-water-collector), used in (23).
- 993 ^l_{A121} [ML^2/t^2] Complex Hamaker constant (nanoparticle-water-nanoparticle), used in (23).
- 994 ^m _{ρ_w} [M/L^3] water density, used in (26)
- 995 ⁿ μ_w [$M/(L \cdot t)$] absolute water viscosity, used in (26)
- 996 ^o_T [K] Temperature, used in (21).
- 997 ^p _{ρ_n} [M/L^3] nanoparticle density, used in (26)
- 998 ^q _{ρ_b} [M/L^3] bulk density of the solid matrix, used in (26)
- 999 ^r _{α} [-] collision efficiency, used in (26).
- 1000 ^s_g [m/hr^2] acceleration of gravity, used in (26).
- 1001 ^t_s [mol/L] ionic strength, used in (23).
- 1002 ^u_{N_A} [1/mol] Avogadro's number, used in (23).
- 1003 ^v_e [C] elementary charge, used in (23).
- 1004 ^w _{ϵ_0} [$C^2/(J \cdot L)$] permittivity of free space, used in (23).
- 1005 ^x _{ϵ_r} [-] relative dielectric constant of the suspending liquid, used in (23).
- 1006 ^y _{Ψ_{p1}} [mV] surface potential of a particle, used in (23).
- 1007 ^z _{σ_{Born}} [L] Born collision parameter, used in (23).
- 1008
- 1009
- 1010
- 1011
- 1012
- 1013
- 1014
- 1015
- 1016
- 1017
- 1018
- 1019
- 1020
- 1021
- 1022
- 1023
- 1024
- 1025
- 1026
- 1027
- 1028
- 1029
- 1030
- 1031
- 1032
- 1033
- 1034

1035



1036

1037 **Figure 1.** Dimensionless total number concentrations (\bar{n}_T/\bar{n}_0^0) as a
1038 function of time for nanoparticle: (a) aggregation based on analytical and
1039 numerical solutions for simple kernel $k_{ij}=1$ and two different clusters ($k=10$ and
1040 20), and (b) transport based on the commercial software ComsolTM and the
1041 present numerical model at $x=0.6$ m.

1042

1043

1044

1045

1046

1047

1048

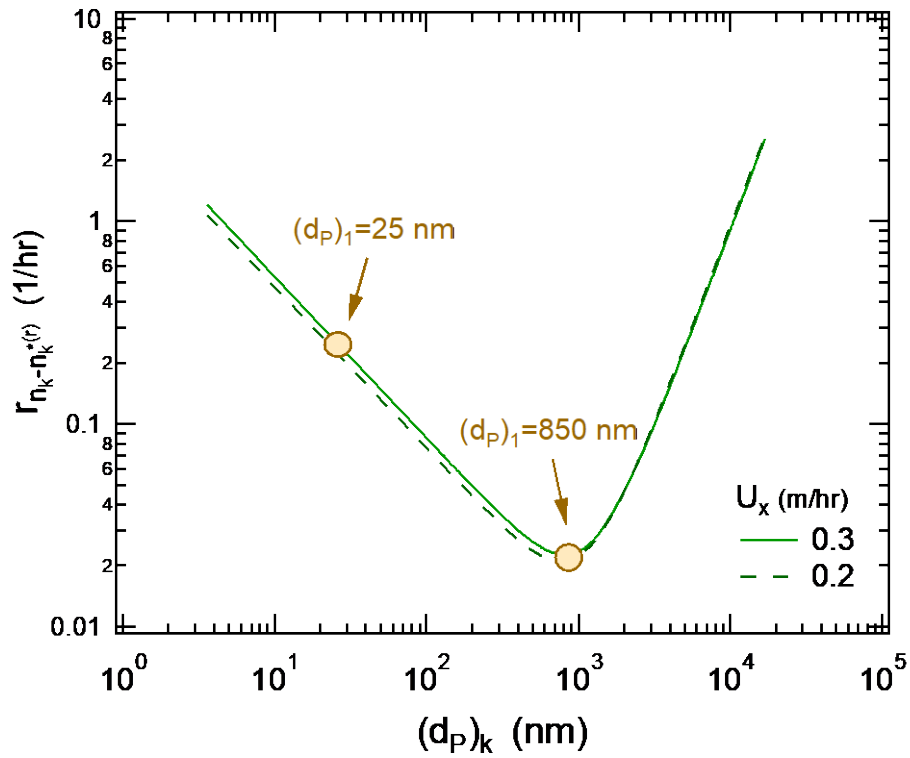
1049

1050

1051

1052

1053



1054

1055

1056

1057

1058

1059

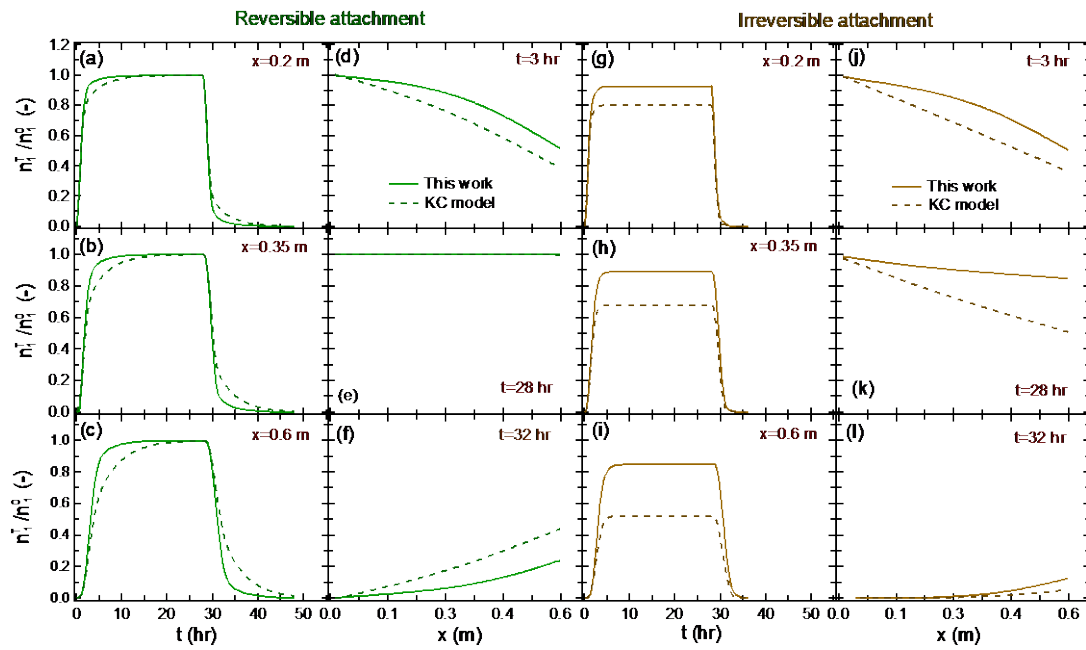
1060

1061

1062

1063

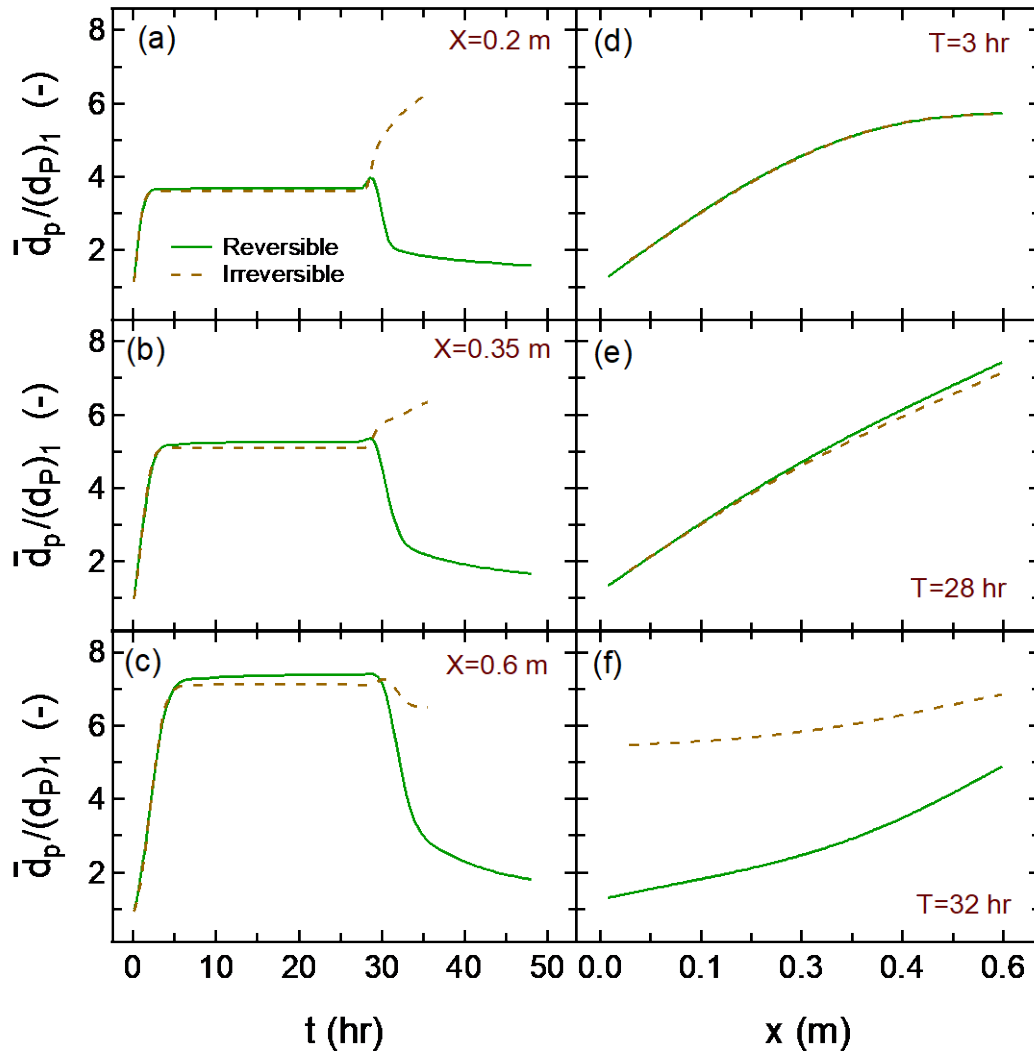
Figure 2. Forward attachment rate as a function of particle diameter, for two different interstitial velocities. The two aggregate diameters (d_{p1}) used in this study are shown.



1064
1065

1066 **Figure 3.** Dimensionless nanoparticle total number concentrations of cluster
 1067 $k=1$, for both cases of reversible attachment and irreversible attachment, for a
 1068 source of nanoparticles in the form of a broad pulse with $t_p=28$ hr, as a
 1069 function of time at three different locations: (a)&(g) $x=0.2$ m, (b)&(h) $x=0.35$ m
 1070 and (c)&(i) $x=0.6$ m, and a function of space for three different times: (d)&(j)
 1071 $t=3$ hr, (e)&(k) $t=t_p=28$ hr, and (f)&(l) $t=32$ hr. The continuous curves are
 1072 simulated by the present nanoparticle transport model and the dashed curves
 1073 by the KC model.

1074
1075
1076
1077
1078



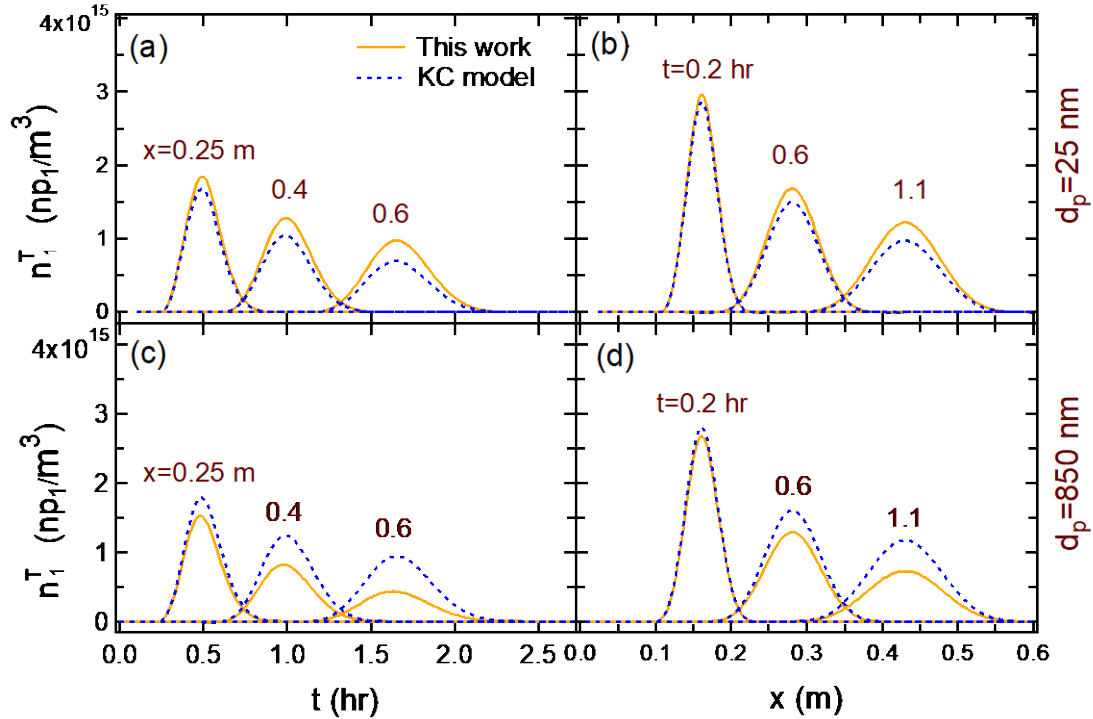
1079

1080 **Figure 4.** Dimensionless average size of suspended aggregates, for both
 1081 cases of reversible attachment (solid curves) and irreversible attachment
 1082 (dashed curves), for a source of nanoparticles in the form of a broad pulse
 1083 with $t_p=28$ hr, as a function of time at three different locations: (a) $x=0.2$ m, (b)
 1084 $x=0.35$ m and (c) $x=0.6$ m, and a function of space for three different times:
 1085 (d) $t=3$ hr, (e) $t=t_p=28$ hr, and (f) $t=32$ hr.

1086

1087

1088



1089

1090 **Figure 5.** Total number concentrations of suspended nanoparticles of cluster
 1091 $k=1$, introduced instantaneously in the 1-D aquifer, as a function of: (a)&(c)
 1092 time at three different locations ($x=0.25, 0.4, \text{ and } 0.6$ m), and (b)&(d) space at
 1093 three different times ($t=0.2, 0.6, \text{ and } 1.1$ hr). Two different nanoparticle sizes
 1094 are considered ($(d_p)_1=25$ and 850 nm). The continuous curves are simulated
 1095 by the present nanoparticle transport model and the dashed curves by the KC
 1096 model.

1097

1098

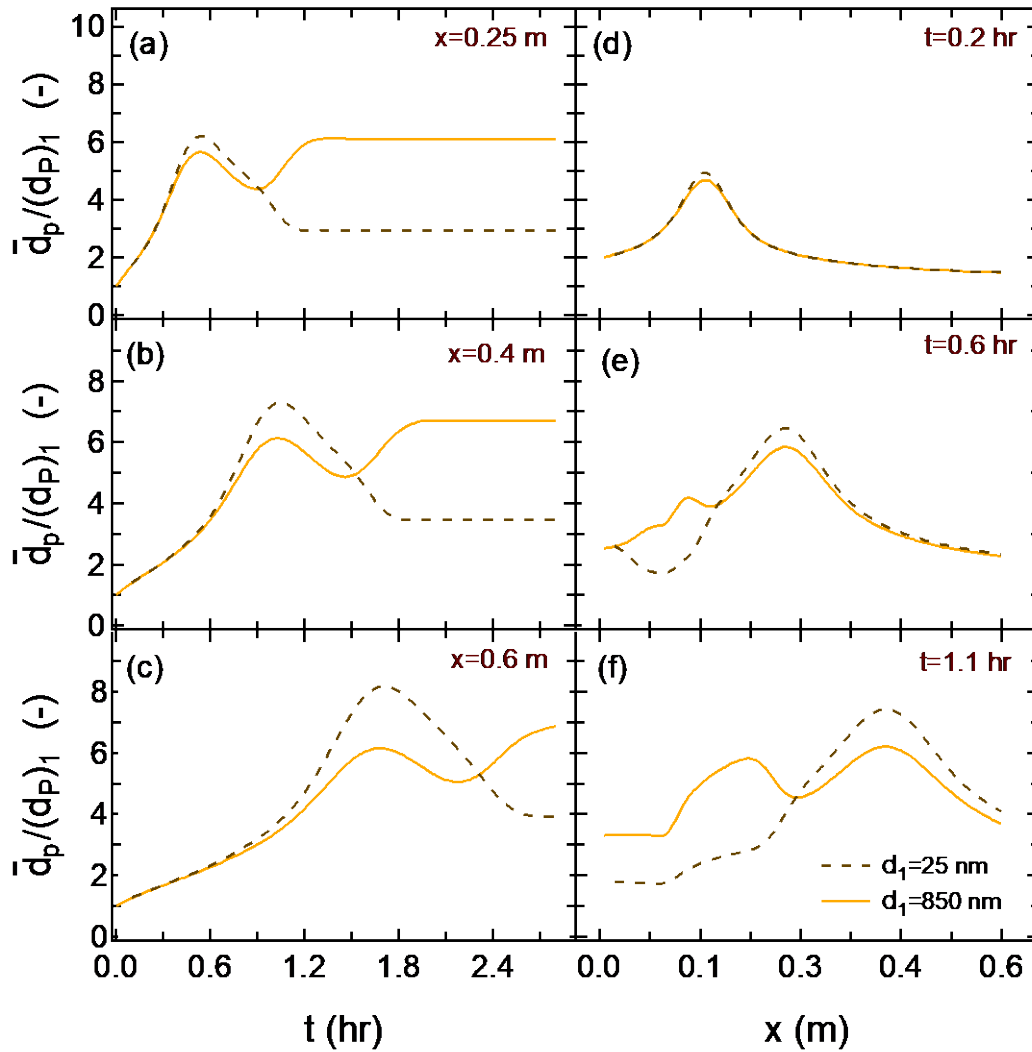
1099

1100

1101

1102

1103



1104

1105

1106 **Figure 6.** Dimensionless average size of suspended aggregates of

1107 nanoparticles introduced instantaneously in the 1-D aquifer, undergoing

1108 reversible attachment, simulated by the present nanoparticle transport model,

1109 as a function of time at three different locations: (a) $x=0.25$ m, (b) $x=0.4$ m and

1110 (c) $x=0.6$ m, and a function of space for three different times: (d) $t=0.2$ hr, (e)

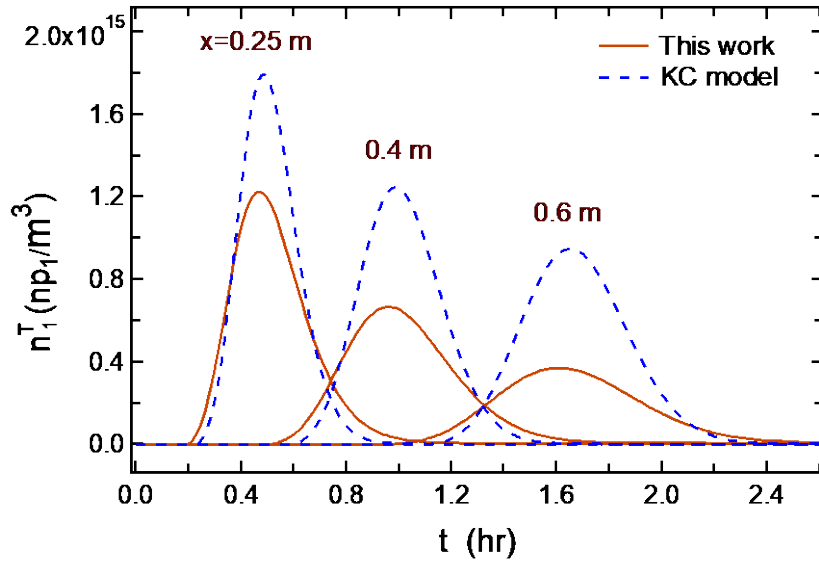
1111 $t=0.6$ hr, and (f) $t=1.1$ hr. The dashed curves correspond to nanoparticles with

1112 diameters $(d_p)_1 = 25$ nm, and the continuous curves correspond to

1113 nanoparticles with diameters $(d_p)_1 = 850$ nm.

1114

1115



1116

1117 **Figure 7.** Breakthrough curves, at three different locations ($x=0.25$, 0.4 , and
 1118 0.6 m), of total number concentrations of suspended nanoparticles of cluster
 1119 $k=1$, for nanoparticles with $(d_p)_1=850$ nm, introduced instantaneously in the
 1120 1-D aquifer. The continuous curves are simulated by the present nanoparticle
 1121 transport model accounting for size-dependent dispersivity, and the dashed
 1122 curves by the KC model with an invariant dispersion coefficient.

1123

1124

1125

1126

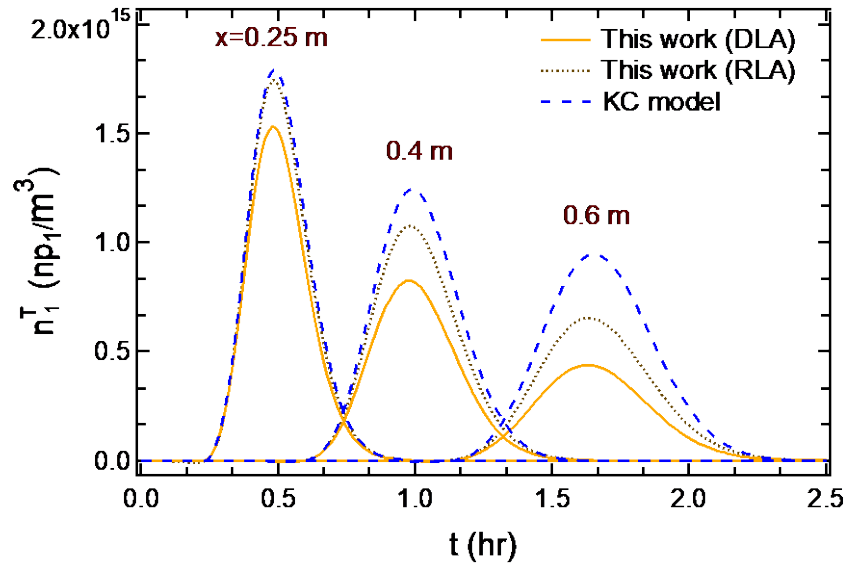
1127

1128

1129

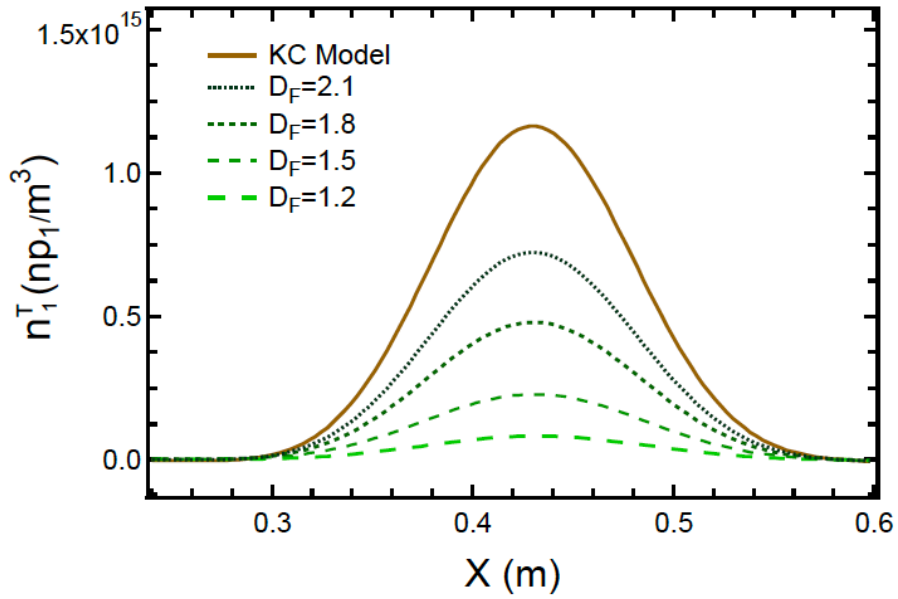
1130

1131



1132
 1133
 1134
 1135
 1136
 1137
 1138
 1139
 1140
 1141
 1142
 1143
 1144
 1145
 1146
 1147
 1148
 1149
 1150
 1151
 1152

Figure 8. Breakthrough curves, at three different locations ($x=0.25$, 0.4 , and 0.6 m), of total number concentrations of suspended nanoparticles of cluster $k=1$, for nanoparticles with $(d_p)_1=850$ nm, introduced instantaneously in the 1-D aquifer. The continuous curves are simulated by the present nanoparticle transport model assuming diffusion-limited aggregation (DLA), the dotted curves by the present nanoparticle transport model assuming reaction-limited aggregation (RLA with $\Psi_{p1}=8.7$ [mV]), and the dashed curves by the KC model.



1153
 1154 **Figure 9.** Total number concentrations of suspended nanoparticles of cluster
 1155 $k=1$, introduced instantaneously in the 1-D aquifer, as a function of space at
 1156 time $t=1.1$ hr, for several D_F [-] values. The continuous curve is simulated by
 1157 the KC model, whereas all other curves are simulated by the present
 1158 nanoparticle transport model. Initial nanoparticle size was $(d_p)_1 = 850$ nm.
 1159