TITLE: Detoxification of paralytic shellfish poisoning toxins in naturally contaminated mussels, clams and scallops by an industrial procedure.

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ABSTRACT

Paralytic shellfish poisoning (PSP) episodes cause important economic impacts due to closure of shellfish production areas in order to protect human health. These closures, when they are frequent and persistent, can seriously affect shellfish producers and the seafood industry, among others. In order to find a solution to this limitation, the obtention of an alternative processing method for bivalves with PSP content above the legal limit, which allows to reduce PSP toxins to acceptable levels, would be desirable. In this study, an industrial protocol, based on Decision 96/77/EC of the European Union which defines a protocol for PSP reduction in Acanthocardia tuberculatum, was developed and implemented for PSP mussel, clams and scallop detoxification. The procedure was applied to 6 batches of PSP-contaminated mussels, 2 batches of PSPcontaminated clams and 2 batches of PSP-contaminated scallops. Detoxification was around 85%, and a viable industrial protocol was developed allowing the transformation of a product at risk into a safe product. Although a significant reduction was obtained, in a sample circa 9000 µg STX diHCl equiv/kg, the final toxin level in these highly toxic mussels did not fall below the European limit. The processing protocol described confirms it may be applied efficiently to mussels, clams and scallops and it may be a major solution to counteract the closure of shellfish harvesting areas, especially if persistent.

Highlights

- In this study, an industrial protocol for PSP toxin reduction to safe levels, based on Decision 96/77/EC, was developed and applied for PSP mussel, clams and scallops detoxification.
- The procedure was applied to 6 batches of PSP-contaminated mussels, 2 batches of clams and 2 batches of scallops obtaining ± 85 % detoxification and a safe product.
- A significant reduction was obtained in a sample with 9000 µg STX diHCl equiv/kg, although it did not fall below the European limit.
- A viable industrial mussel canning processing was developed guaranteeing the manufacture of a safe product.

Keywords: Paralytic shellfish poisoning, detoxification, industrial protocol, mollusks, LC-FLD, MBA. (maximum 6 descriptive keywords)

Abbreviations:

C1-4, N-sulfo-carbamoyl gonyautoxins

CRM, certified reference material dcGTX decarbamoyl gonyautoxin dcNEO, decarbamoyl neosaxitoxin dcSTX, decarbamoyl saxitoxin equiv, equivalent GTX1-6, gonyautoxins HABs, harmful algal blooms *i.p.*, intraperitoneal LC-FLD, liquid chromatography-fluorescence detection MBA, mouse bioassay NEO, neosaxitoxin PSP, paralytic shellfish poisoning STX, saxitoxin TEF, toxicity equivalency factor

INTRODUCTION

Paralytic shellfish poisoning (PSP) is caused by consumption of shellfish containing PSP toxins of the family of saxitoxins (STX). These toxins are produced by microalgae, mainly toxic marine dinoflagellates such as species of the genera *Alexandrium* and *Gymnodinium*, and also by certain freshwater cyanobacteria. These toxins are accumulated and eventually metabolized into toxin derivatives in many species of filter-feeding bivalves, as mussels, clams and scallops, making them potentially toxic to humans. Harmful algal blooms (HABs) can also induce other ecological damage and adverse effects to living marine resources. In fact, some bivalves can be impaired during intense toxic episodes. For instance, a population of the surf clam *M. donacium* with high PSP toxic levels, died due to the desiccation caused by the incapability of the clams to burrow (Álvarez et al., 2019). In addition, severe canine intoxications, due to the consumption of nontraditional (non-bivalve shellfish) marine organisms, such as crabs or starfish containing high amounts of PSP toxins, was recently reported (Turner et al., 2018).

To protect public health and ensure the quality of seafood, monitoring programs are implemented worldwide in order to detect and quantify these toxins, and eventually forbidding shellfish harvesting when levels of toxins exceed the legal limit laid down in current regulations. In Europe for example, harvesting and commercialization of bivalves is prohibited above the threshold of 800 µg STX diHCI equiv/kg of shellfish tissues (EC, 2004). Closure of shellfish production areas has an important economic impact for producers and other associated industries. No solutions have been found to prevent these important episodes

which are seldom predictable, and despite the influence of PSP events on human health and fisheries, studies on shellfish detoxification to mitigate this problem are still very scarce.

Natural detoxification occurs very slowly and it is conditioned by the presence of toxin producing microalgae in the water column. Lipophilic toxins are retained longer than the hydrophilic toxins, such as PSP toxins, although the detoxification rate depends on the species, concentration of toxins and environmental conditions (Lee et al., 2008). Several studies described that some PSP toxins, but not all, can be reduced exposing shellfish to a non-toxic diet (Reis Costa et al., 2018). Nevertheless, mitigating or modulating the presence of microalgae in the field is not possible, so this eventual solution should be applied by maintaining large stocks of shellfish in a closed space for several days, and the feasibility of this would be dubious.

Once harvested, toxin reduction or elimination from shellfish is mainly affected by the chemical properties of the toxins. In the particular case of PSP toxins, a regulation was published after performing scientific studies which proved that a suitable heat treatment decreased the levels of PSP toxins and guaranteed the safety of the cockle *Acanthocardia tuberculata* (EC, 1996).

A detoxification procedure would result in an economically feasible solution for a shellfish canning industry in locations where PSP toxic episodes occur very often or are persistent, and large amounts of shellfish are affected. Besides, in view of the changing environmental conditions related to climate change, a rise in the incidence of these episodes could take place in the near future. It is important to mention that it would not be necessary to perform important modifications in factory installations to accomplish the PSP detoxification

protocol, since the required equipment is the same usually employed by the canning industry. Only the duration of the whole thermal process would be slightly increased.

In this paper, naturally PSP contaminated mussels, clams and scallops were specifically harvested in order to implement the thermal procedure described in the EU decision. Slight modifications were applied, in order to obtain a better efficiency of detoxification and yield of mussels, clams and scallops.

MATERIALS AND METHODS

Sampling of contaminated mussels and scallops

Samples were obtained from different sampling points along the Spanish and Portuguese coasts from July 2018 to March 2019. Mussels (*Mytilus galloprovincialis*) were acquired from several mussel raft cultures in: a) Galicia, (samples coming from two different floating rafts in the Ría of Vigo, Pontevedra); b) Andalucía, (one batch of mussels from Benalmádena, Málaga), and c) Portugal, (one batch of mussels from Portinho da Costa, near Lisbon). In addition, other mussel batches were obtained in Catalonia, one sample, after **exposure** to a toxic bloom of the dinoflagellate *Alexandrium minutum*, inside a harbour, as explained in this article. Special permissions from the local authorities were obtained in order to harvest the toxic molluscs from the closed areas. Two batches of Japanese littleneck clams (*Ruditapes philippinarum*) were obtained from Pontevedra, Galicia (Spain) and both batches of scallops (*Pecten maximus*) were obtained from Málaga, Andalucia (Spain). Sampling zones where toxic mussels, clams and scallops were harvested are depicted in Figure 1.



Fig 1: sampling points (marked by arrows) where PSP contaminated mussels and scallops were obtained during the study.

Samples were refrigerated in thermally isolated boxes with cold accumulators after collection and shipped to the laboratory. Upon arrival, samples were processed as indicated in "Detoxification study" and analyzed as described below. Some subsamples of the different batches of mollusks were frozen at - 20 °C and processing and analysis was performed after days or weeks until a maximum of 10 weeks.

Mussel exposure to a toxic bloom of the dinoflagellate Alexandrium minutum

A controlled field study was carried out in the Catalonian coast exposing 50 kg of edible mussel for 5 days to a toxic bloom of *Alexandrium minutum*, a known producer of PSP toxins. The objective was to allow high levels of PSP toxins to bioaccumulate in the mussels. Levels of *A. minutum* were always above 200000

cells/L and, as a result, the concentration of PSP toxins in mussels was higher than 4000 µg STX diHCl equiv/kg.

Procedure for PSP mussels, clams and scallops detoxification

The regulated procedure (EC, 1996), was applied to all different batches of PSP naturally contaminated mussels, clams and scallops with some modifications:

- Preliminary cleaning in running fresh water for two minutes.
- Pre-cooking in fresh water for three minutes at a temperature of 95 ± 5 °C.
- Separation of flesh and shells.
- Second cleaning in fresh water for 30 seconds.
- \circ Cooking in fresh water for nine minutes at a temperature of 98 ± 5 °C.
- Cooling in running fresh water for approximately 90 seconds.
- Conditioning in containers closed hermetically in a non-acidified liquid medium.
- Sterilization in autoclave at 116 °C for 51 min (referred as "Canning") or Pasteurization at 90 °C for 10 min.

Separation of the edible parts (foot) from the non-edible parts (gills, viscera and mantle), in mussels and clams, was omitted in order to increase the yielding of the process. In the case of scallops, edible parts is the sum of adductor muscle and roe. Samples subjected to the detoxification method are identified along the text as "EC". Aliquots of the same batches of mussels, clams or scallops were sterilized or pasteurized without applying the detoxification procedure and are identified along the text as "normal".

Toxin extraction

Two laboratories, ANFACO and IRTA, were involved in the extraction and the analysis of PSP toxins in the samples, either processed or not. Both laboratories performed the same extraction and analysis protocol described below, only with variations related to the chromatographic columns and LC equipment used.

Chemicals

Milli-Q ultrapure water, acetonitrile LC-MS grade (Scharlau), methanol LC-MS grade (Fisher), ammonium formate HiPerSolv Chromanorm® for LC-MS (VWR), glacial acetic acid reagent grade (Scharlau), periodic acid analytical reagent AnalaR NORMAPUR (VWR), Na₂HPO₄ analysis grade (MERCK), sodium hydroxide reagent grade (Scharlau), hydrogen peroxide solution 30% (v/v) reagent grade (Scharlau), ammonium acetate reagent grade (Scharlau), sodium chloride for analysis (MERCK).

SPE cartridges: SPE C18 sep-pack (3 mL, 500 mg) (Waters), SPE COOH (3 mL, 500 mg) (Bakerbond).

Certified PSP standards used at ANFACO: CRM-00-STX, gonyautoxins 1-5 (CRM-00-GTX1&4, CRM-00-GTX2&3, CRM-00-GTX5), neosaxitoxin (CRM-00-NEO), decarbamoylneosaxitoxin (CRM-00-dcNEO), decarbamoylsaxitoxin (CRM-00-dcSTX), N-sulfocarbamoyl gonyautoxin-2&3 (CRM-00-C1&2), and decarbamoylgonyautoxin-2&3 (CRM-00-dcGTX2&3) and gonyautoxins 6 (CRM-00-GTX6) were purchased from Cifga (Lugo, Spain).

Certified standards PSP used at IRTA: CRM-00-dcGTX2&3, CRM-00-C1&2, CRM-00-dcSTX, CRM-00-GTX2&3, CRM-00-GTX5, CRM-00-STX, CRM-00-GTX1&4, CRM-00-NEO and CRM-00-dcNEO were purchased from the National

Research Council (NRC, Halifax, NS, Canada) and CRM-00-GTX6 was purchased from Cifga (Lugo, Spain).

Standard solutions

Standard mixtures were prepared from the commercial standards, at ANFACO: MIX I (STX, dcSTX, GTX2,3, dcGTX2,3, GTX5 and C1,2), MIX II (NEO and GTX1,4), MIX III (dcNEO), MIX IV (dcSTX) and MIX V (GTX6). For each MIX, 6 standard calibration levels were prepared by dilution with Milli-Q water in the range 0.006 – 1 μ M (MIX I) and 0.015 – 1 μ M (the remaining MIX). These standard solutions were preserved at -20°C. Individual LQ were 45 µg equiv. STX diHCI/kg for dcGTX2,3; 5 µg equiv. STX diHCI/kg for C1,2; 40 µg equiv. STX diHCI/kg for dcSTX; 25 µg equiv. STX diHCI/kg for GTX2,3; 5 µg equiv. STX diHCI/kg for GTX5; 40 µg equiv. STX diHCI/kg for STX; 150 µg equiv. STX diHCI/kg for GTX1,4; 140 µg equiv. STX diHCI/kg for NEO; 80 µg equiv. STX diHCI/kg for dcNEO; 13 µg equiv. STX diHCI/kg for GTX6 and 20 µg equiv. STXdiHCl/kg for C3,4. Sum of individual LQ of PSP toxins at ANFACO-CECOPESCA was 563 µg equiv. STX diHCI/kg. At IRTA, standard mixtures were prepared from the commercial standards: MIX I (STX, dcSTX, GTX2,3, dcGTX2,3, GTX5 and C1,2), MIX II (NEO and GTX1,4), MIX III (dcNEO), MIX IV (dcSTX) and MIX V (GTX6). For each MIX, 6 standard calibration levels were prepared by dilution with Milli-Q water in the range LQ-800 µg STX diHCI equiv/kg. Individual LQ were 46 µg equiv. STX diHCl/kg for dcGTX2,3; 6 µg equiv. STX diHCl/kg for C1,2; 40 µg equiv. STX diHCl/kg for dcSTX; 26 µg equiv. STX diHCI/kg for GTX2,3; 5 µg equiv. STX diHCI/kg for GTX5; 40 µg equiv. STX diHCl/kg for STX; 150 µg equiv. STX diHCl/kg for GTX1,4; 140 µg equiv. STX diHCI/kg for NEO; 80 µg equiv. STX diHCI/kg for dcNEO and 13 µg equiv. STX diHCI/kg for GTX6. Sum of individual LQ of PSP toxins was 546 µg equiv. STX diHCI/kg. These standard solutions were preserved at -20°C.

Extraction, clean-up, hydrolysis and oxidation

The method was based on the HPLC-FLD Official Method (AOAC, 2005; Lawrence et al., 2005), and refined as described by Turner et al. and Ben-Gigirey et al. (Ben-Gigirey et al., 2012; Turner et al., 2009). The method involves an acetic acid extraction through clean-up with SPE C18 cartridge extraction followed by periodate oxidation and analysis by HPLC-FLD. If the presence of any toxin is observed, peroxide oxidation and/or fractionation (F1, F2, F3) are then carried out by using COOH ion exchange SPE cartridges with periodate oxidation, injecting the obtained extracts in the HPLC-FLD.

PSP toxin quantitation

The toxins dcGTX2,3, C1,2, dcSTX, GTX2,3, GTX5, STX were quantified in the C18 extract after peroxide oxidation; GTX1,4 and GTX6 in F2 fraction; NEO and dcNEO in F3 fraction and C3,4 in F1 hydrolizated fraction after periodate oxidation.

Total PSP toxin content, expressed as STX diHCl equivalents/kg, is calculated by summing individual toxin concentrations and applying toxicity equivalents factors (TEF) that are established for each toxin according to EFSA Scientific Opinion (EFSA, 2009).

In the samples where no PSP toxins has been detected (below LODs), the histograms were left blank.

HPLC-FLD equipment and chromatographic conditions

PSP toxins analyses, at ANFACO, were carried out using an HPLC Alliance 2695 model and fluorescence detector 2474 model (Waters Corporation). A XSelect CSH C18 3.5 μ m, 4.6 mm x 150 mm column and a XSelect CSH C18 3.5 μ m, 3.9 mm x 5 mm precolumn from Waters were used. Chromatography conditions are described in the AOAC Method (Lawrence et al. 2005).

At IRTA, PSP toxins analyses were carried out using an UPLC Acquity H-Class model and FLR Acquity fluorescence detector (Waters Corporation). A Kinetex C18 4.5 μ m, 4.6 x 150 mm column and a XSelect CSH C18 4.5 μ m guard column from Phenomenex were used. Chromatography conditions used are those described in the rapid method by Hatfield et al. (Hatfield et al., 2016).

Method performance

The method acceptability criteria were selected to ensure the performance of the method, according to the International Organization for Standardization (ISO) 17025:2005 standards and the screening and semi-quantitation of PSP toxins EURLB-SOP quality requirements (EURLMB, 2019). The minimum performance criteria were checked out throughout the study such as retention time deviation \pm 0.2 min, peak area deviation (RSD ≤3.0%), linearity (R² ≥0.98), sensitivity (individual toxin LOD should be equal or lower than 1:20th of regulatory level), precision intra-batch ≤20% and inter-batch ≤25%.

PSP Mouse Bioassay (MBA)

Some raw mussel samples were also analyzed with the MBA, as a reference, before the recent regulation on PSP detection methods was established (EC, 2017), since this was the reference method to quantify PSP toxins at the moment of the study. Briefly: 100 g of shellfish were homogenized on 100 ml HCl 0.18 N, pH adjusted between 2 and 4 with NaOH and boiled for 5 min. Sample was cooled down and 1 ml of filtered extract *i.p.* injected into 3 male albino mice (20 g aprox). Mice were monitored for 1 h to quantify the toxin according to the time of death. The limit of detection of this technique was 350 µg STX diHCl equiv/kg (AOAC, 1995). MBA assays were performed at ANFACO facilities, under European guidelines and legislation on laboratory animals welfare. ANFACO is a registered user of laboratory animals with identification number ES360570160901.

RESULTS

The different batches of cultivated mussels (*M. galloprovincialis*), clams (*R. philippinarum*) and scallops (*P. maximus*), origin and sampling place, date of harvesting and analytical results initially obtained in the raw mollusks, are summarized in Table 1. In this table, mean values ± standard error of the mean (SEM) obtained for each sample analyzed by both laboratories are included.

The different batches were split and samples were processed by the different thermal treatments as described above. A standard canning or pasteurization, as usually performed in an industrial situation, and the detoxification procedure followed by sterilization or pasteurization were carried out. These treatments are referred, respectively, as EC canning, EC Pasteurization, Normal canning and Normal pasteurization. All samples were analyzed by HPLC by the two laboratories (ANFACO and IRTA); some raw samples were analyzed by MBA. The HPLC results obtained by both laboratories showed good agreement, and good correlation was obtained, as shown in Figure 2.



Fig 2: Correlation chart for the total content of PSP toxins present in raw bivalves set analyzed by HPLC-FLD (µg STX diHCl equiv/kg) (n=13) showing a good correlation between results obtained at ANFACO and IRTA laboratories.

Table 1. Origin and date of harvesting of live PSP contaminated mussels and scallops. Results
of PSP toxins (mean values ± standard deviation, n=2) in raw bivalves analyzed by HPLC-FLD,
at both laboratories.

Species	Location	Harvesting date	Average result (μg STX diHCl equiv/kg) (n=2)
Mussel (Mytilus galloprovincialis)	Ría of Vigo (Vigo A)	09/07/2018	1072 ± 11
Mussel (<i>M. galloprovincialis</i>)	Ría of Vigo (Redondela C)	23/07/2018	1604 ± 330
Mussel (M. galloprovincialis)	Ría of Vigo (Redondela C)	23/07/2018	737 ± 134
Mussel (<i>M. galloprovincialis</i>)	Andalucía (Benalmádena)	02/08/2018	812 ± 270
Mussel (<i>M. galloprovincialis</i>)	Portinho da Costa (Lisbon)	22/10/2018	9001 ± 345
Mussel (<i>M. galloprovincialis</i>)	Catalonia	05/03/2019	4205 ± 43
Mussel (<i>M. galloprovincialis</i>), frozen	Catalonia	05/03/2019	2317 ± 261
Clam (R. philippinarum)	Ría of Pontevedra	27/07/2018	1041 ± 23
Clam (R. philippinarum), frozen	Ría of Pontevedra	27/07/2018	903 ± 204
Scallop (Pecten maximus)	Andalucía	22/08/2019	3232 ± 466
Scallop (P. maximus) eviscerated	Andalucía	22/08/2019	1976 ± 117
Scallop (P. maximus)	Andalucía	22/08/2019	3171 ± 30
Scallop (P. maximus), eviscerated	Andalucía	22/08/2019	1779 ± 126

Fig 3A, shows the levels of PSP toxins, expressed as μ g STX diHCl equiv/kg, in raw or thermally processed mussels from Redondela, Galicia. Results show that, with these levels of toxins, around 1604 μ g STX diHCl equiv/kg, the normal sterilization procedure or the application of the detoxification process (EC), previously to retorting or pasteurization, is able to decrease PSP levels below the limit of detection.



Fig 3A. PSP toxins in naturally contaminated mussels from Redondela, NW Spain analyzed by HPLC. "Raw" corresponds to mussels without processing. Mean values (n=2) are represented, and vertical bars indicate the SEM. The raw product was also analyzed by mouse bioassay (MBA Raw, in the figure).

Fig 3B shows the quantification of the different PSP analogues detected in the mussels from Redondela, expressed as STX diHCI equiv/kg. It is worth to mention that in the sample, GTX1,4 was the dominant toxin. Very low concentration of other analogues, such as dcSTX, GTX2,3 and C3,4 toxins were detected as well.



Fig 3B. PSP toxins in naturally contaminated mussels from Redondela, NW Spain. Figure represents all the PSP analogues detected in each sample, raw or processed under different conditions. Mean values (n=2) are represented, and vertical bars indicate the SEM

Similar results were obtained in other batch of mussels from the same location, with a PSP concentration near the legal limit, 800 μ g STX diHCl equiv/kg as shows Fig 3C.



Fig 3C. Naturally contaminated mussels from Redondela, NW Spain analyzed by HPLC. Mean values (n=2) are represented, and vertical bars indicate the SEM. The raw product was also analyzed by mouse bioassay (MBA Raw, in the figure).

A similar behavior was found with a new lot of mussels from the same areas but from a different location, Vigo, also in Galicia, as illustrates Fig 4.



Fig 4. Naturally contaminated mussels from Vigo, NW Spain analyzed by HPLC. "Raw" corresponds to mussels without processing. Mean values (n=2) are represented, and vertical bars indicate the SEM. The raw product was also analyzed by mouse bioassay (MBA Raw, in the figure).

A batch of mussels from the South of Spain, Málaga, in Andalucía, was also processed, obtaining an important reduction of PSP toxins concentration, as expected. All the protocols applied allowed to decrease initial levels of PSP toxins below the legal limit, as shows Fig 5A.



Fig 5A. Naturally contaminated mussels from Andalucía, S Spain analyzed by HPLC. Mean values (n=2) are represented, and vertical bars indicate the SEM. The raw product was also analyzed by mouse bioassay (MBA Raw, in the figure).

Fig 5B represents results obtained with the raw or thermally processed samples showing all PSP analogues identified in the mussels from Andalucia. Raw sample contained mostly GTX1,4 and in lower concentration GTX2,3, dcSTX, C1,2 and GTX5. However, it is remarkably that after processing dcSTX is the dominant analogue.



Fig 5B Naturally contaminated mussels from Andalucia, Spain. Figure represents all the PSP analogues detected in each sample, raw or processed under different conditions. Mean values (n=2) are represented, and vertical bars indicate the SEM.

A new batch of mussels with an extremely high concentration of PSP toxins was harvested from Portugal, as shows Fig 6A. In this case, mussels exposed to the toxic episode for a long time, presented levels of toxins obtained with HPLC (9000 μ g STX diHCl equiv/kg), that exceeded more than 10 times the legal limit. Although after application of the detoxification procedure, levels of PSP toxins decreased in a significant way, reaching more than 85 % of detoxification, no safe products were attained in this case, since PSP toxins levels in EC canned mussels, were still higher (1054±33 μ g STX equiv/kg) than the legal limit (800 μ g STX equiv/kg). Surprisingly, after application of a pasteurization following the EC procedure, levels of PSP toxins were lower than those reached after sterilization and below the legal limit (783±183 μ g STX diHCl equiv/kg).



Fig 6A. Naturally contaminated mussels from Lisbon, Portugal, analyzed by HPLC. Mean values (n=2) are represented, and vertical bars indicate the SEM. The raw product was also analyzed by mouse bioassay (MBA Raw, in the figure).



Fig 6B. Naturally contaminated mussels from Lisbon, Portugal, analyzed by HPLC. Figure represents all the PSP analogues detected in each sample, raw or processed under different conditions. Mean values (n=2) are represented, and vertical bars indicate the SEM.

Fig 6B shows the results showing all PSP analogues detected in the sample from Lisbon, Portugal. The raw sample contained several toxins of the group, mainly

dcSTX; GTX6; GTX5, C1,2 and dcGTX2,3. Again, dcSTX was the dominant analogue in the processed samples, even at higher levels than in the raw sample after "Normal Canning". This fact suggests that a transformation of other toxins to dcSTX takes place due to the canning process.

Similar results were obtained in the batch of mussels from Catalonia containing a final concentration of 4206 µg STX diHCl equiv/kg (Fig 7A). It is worth mentioning that, the same sample, after frozen storage at -20°C for 3 weeks, contained 2318 µg STX diHCl equiv/kg (Fig 8A). Application of both treatments: the normal sterilization procedure and the detoxification process (EC), previously to retorting or pasteurization, produce a decrease in PSP levels, below the legal limit.



Fig 7A. Naturally contaminated mussels after controlled immersion into an area with *A minutum* in Catalonia, Spain analyzed by HPLC. Mean values (n=2) are represented, and vertical bars indicate the SEM.

The raw sample of immersed mussels coming from Catalonia contained several toxins of the group, mainly GTX1,4; GTX2,3 and STX. STX was the dominant analogue in all four processed samples, and was not present in the raw sample. This fact suggests that a transformation to STX takes place due to the thermal process (Figure 7B).



Fig 7B. PSP toxins in naturally contaminated mussels after controlled immersion into an area with *A minutum* from Catalonia, Spain, fresh sample. Figure that represents all the PSP analogues detected in each sample, raw or processed under different conditions. Mean values (n=2) are represented, and vertical bars indicate the SEM.

Analysis of the same sample after preservation at -20°C, resulting in a an important decrease in total PSP toxin content, showed a similar behavior regarding decrease of toxins after processing (Figure 8A) and the same toxins of the group were present, mainly GTX1,4; GTX2,3 and STX (Figure 8B). STX was again the dominant analogue in the processed samples, even at higher levels than in the raw sample after all thermal processing. This fact suggests again that a transformation to STX takes place due to the canning and pasteurization process (Figure 7B).



Fig 8A. Naturally contaminated mussels from Catalonia after storage at -20°C analyzed by HPLC. Mean values (n=2) are represented, and vertical bars indicate the SEM.



Fig 8B. PSP toxins in naturally contaminated mussels from Catalonia after storage at - 20°C. Figure that represents all the PSP analogues detected in each sample, raw or processed under different conditions. Mean values (n=2) are represented, and vertical bars indicate the SEM.

Regarding the clams, similar results of mussels were obtained containing concentration of $1041 \pm 22 \ \mu g$ STX diHCl equiv/kg (Fig 9A). The same sample after frozen storage at -20°C for 7 weeks, contained 903 \pm 204 μg STX diHCl equiv/kg (data not shown). Application of both treatments: the normal sterilization procedure and the detoxification process (EC), previous to retorting or pasteurization, produce a decrease in PSP levels, to not detectable levels (Fig.9B).



Fig 9A. PSP toxins in naturally contaminated clams from Pontevedra, NW Spain analyzed by HPLC. Mean values (n=2) are represented, and vertical bars indicate the SEM.



Fig 9B. Naturally contaminated scallops from Pontevedra, Galicia, Spain, analyzed by HPLC. Figure represents all the PSP analogues detected in each sample, raw, eviscerated or processed under different conditions. Mean values (n=2) are represented, and vertical bars indicate the SEM.

Regarding the scallops, evisceration of raw specimens reduced significantly the content of the total PSP toxin content, (Fig 10A). Evisceration of scallops during the cleaning steps and processing after precooking reduced drastically all the congeners to non-detectable levels when retorting was performed, and to levels well below the legal limit when pasteurization was performed, only dcSTX ($327\pm69 \ \mu g$ STX diHCl equiv/kg) and GTX5 ($28\pm9 \ \mu g$ STX diHCl equiv/kg) were detectable when pasteurization was carried out after a conventional pre-cooking step, and only dcSTX was detectable to a lower level ($144\pm19 \ \mu g$ STX diHCl equiv/kg) when pasteurization was carried out after the detoxification protocol (Fig 10B).



Fig 10A. Naturally contaminated scallops (*Pecten maximus*) from Marbella, Andalucía, Spain. Mean values (n=2) are represented, and vertical bars indicate the SEM.



Fig 10B. Naturally contaminated scallops from Marbella, Andalucía, Spain, analyzed by HPLC. Figure represents all the PSP analogues detected in each sample, raw, eviscerated or processed under different conditions. Mean values (n=2) are represented, and vertical bars indicate the SEM.

DISCUSSION

Some studies have been conducted to reduce or eliminate PSP toxins in mollusks and other invertebrates. The influence of thermal processing in naturally contaminated bivalves has already been studied by our group and by other authors, finding PSP detoxification in shellfish after application of high temperatures (Lawrence et al., 1994; Reboreda et al., 2010; Vieites et al., 1999).

In this study, an approved thermal procedure to decrease PSP toxins in the giant cockle, *Acanthocardia tuberculata*, (EC, 1996), was evaluated on mussels, clams and scallops naturally contaminated with PSP toxins). We applied the heat treatment (so called "detoxification procedure") that consists on several cleaning and cooking steps, as establishes the European legislation but with certain modifications (EC, 1996). Mainly, in this legislated procedure, the edible parts are separated from the non-edible parts. In the case of mussels, the digestive viscera constitute 30 % of the total tissue weight (FAO, 2004), and removing it is not realistic regarding the commercial practices by the industry for this species (as well as for clams), while it is accepted for scallops. Hence, we did eviscerate the scallops, since this is a common practice.

A standard canning or pasteurization, as usually performed in an industrial situation, without previous washes and cooking, was carried out to compare the results with those obtained after application of the "detoxification procedure".

Different batches of mussels, clams and scallops containing PSP levels above the regulated limit (800 μ g STX equiv/kg) and coming from several areas of Spain and Portugal, were thermally processed. In our hands, scallops, clams and almost all mussels unless one, underwent an important detoxification process. Only one sample of mussels from Portugal, containing a total PSP

toxin content (9000 µg STX diHCl equiv/kg) more than 10 times the legal concentration (Fig 6) showed significant decrease in total PSP toxin content after processing, except for normal canning with a decrease of 22% in total PSP toxin content. Otherwise, the rest of samples, either after standard canning or pasteurization, independent on the application of the "detoxification procedure", reached levels of detoxification of 85 %. This was a much unexpected finding since the levels of toxins were in all samples above the legal limit. These results suggest that in some circumstances, if concentrations of toxins in shellfish are not very elevated, it is not necessary to apply the "detoxification procedure", since a normal canning or pasteurization seems sufficient to reduce PSP levels. Nevertheless, in our opinion, detoxification of mussels depends not only on the initial concentration, but also on the exposure time to toxins and the water conditions as well. So, the detoxification procedure should be applied in all cases. In contrast, when shellfish are contaminated with high levels (>5300 μ g STX diHCl equiv/kg), even the "detoxification procedure" application is not enough to reduce PSP levels below the legislated limit. Taking this into account, a threshold level should be established in mussels if a detoxification legislative proposal is expected. Based on data found in this work, with detoxification levels of 85%, a maximum level of 5300 µg STX diHCl equiv/kg should be stablished for the application of this procedure. The highest level allowed in the European Legislation authorized for the harvesting of A. tuberculata is 3000 µg STX diHCI equiv/kg, if the product is intended to the canning industry (EC, 1996). In addition to this Decision, applied in Spain for the giant cockle, a legislation in Canada allows canning of soft shell clams and mussels with levels between 800 and 1600 µg STX equiv/kg (Fernández et al., 2003). Also, butter clams

containing 3000 to 5000 µg STX equiv/kg may be commercialized after removing the entire siphon, whereas butter clams containing 800 to 3000 µg STX equiv/kg may be marketed after removing the distal half of the siphon (Fernández, 1998), cited in (FAO, 2004).

A preliminary article shows that the standard canning process resulted in a significant and reproducible reduction of PSP toxicity in mussel meat (> 50%), decreasing toxin levels under the limit of detection (Vieites et al., 1999). In this paper, authors observed that the decrease of toxicity was not dependent on toxin levels of raw material, although it is worth to mention that raw samples in this work contained PSP toxin levels below 4000 μ g STX diHCl equiv/kg. These results are in good agreement with ours, since only when the concentrations of PSP toxins are very high (9000 μ g STX diHCl equiv/kg), we find that the traditional canning did not allow detoxifying mussels below the legal limit. A subsequent work confirmed the detoxification after canning, although no very relevant results were obtained in that study due to the unavailability of highly PSP contaminated shellfish (Reboreda et al., 2010).

In lobsters, boiling or steaming reduced toxicity by approximately 65% compared to values obtained in the raw samples (Lawrence et al., 1994).

Our results suggest that thermal treatment induces chemical modifications of toxins, changing the toxic potency of processed shellfish, converting toxins into more or less toxic analogues, as occurs by biotransformation or metabolic transformations in shellfish (Reis Costa et al., 2018).

In the present study, GTX1,4 was the dominant detected toxin in raw samples of mussels, except for the most toxic sample from Portugal where a mixture of analogues was identified and dcSTX and GTX6 were the major toxins with

similar levels. Among the different PSP toxins, dcSTX was the major analogue obtained in processed samples, including the one with the highest PSP concentration except for the mussel from Catalonia where the major analogues obtained in the processed samples was STX. GTX1 and GTX4 together are less toxic (TEF=1 and 0.7, respectively) than dcSTX (TEF=1) (EFSA, 2009). This was taken into account for the calculation of total PSP toxin content since HPLC results were expressed as STX diHCI equiv/kg.

More studies will be necessary to assess if toxins in cooked mussels and scallops are removed through chemical decomposition, leached out during the loss of water or transferred to packing medium.

Regardless of these possibilities, our study demonstrates that processing of PSP contaminated shellfish, significantly reduces the PSP toxin content and industrial processes may be a solution in shellfish harvesting areas affected by intense and frequent PSP closures. The proposed method has been already implemented in the transformation of *Acanthorcardia tuberculatum* and should be easily transferred to other bivalve transforming facilities.

CONCLUSIONS

In conclusion, an efficient and unexpensive "detoxification procedure" can be applied in PSP contaminated mussels, clams and scallops to decrease PSP toxins below the legal limit (800 µg STX diHCl equiv/kg). However, a maximum threshold level in raw material should be previously established to define if the processing will efficiently reduce PSP toxins below the legal limit. Based on our data, 5300 µg STX diHCl equiv/kg would be the highest level. Although it is still necessary that the industry should proceed with quality controls of the final

product to ensure that it responds to the legal requirements and levels of PSP toxins are safe, in the same way as stated in the reference legislation for *Acanthocardia tuberculatum*.

ACKNOWLEDGEMENTS

This work has been financed by the European Union's Horizon 2020 research and innovation programme under Grant Agreement no. 773400 SEAFOOD^{TOMORROW} project.

The authors acknowledge the Departament d'Agricultura, Ramaderia, Pesca i Alimentació (DARP) of the Generalitat de Catalunya for financing of the Shellfish Harvesting Areas Monitoring Program. We also appreciate the support of the CERCA program of the Generalitat de Catalunya. The authors acknowledge Margarita Fernández-Tejedor for her assistance in the mussel exposure to a toxic bloom of the dinoflagellate *Alexandrium minutum*.

REFERENCES

Álvarez, G., Díaz, P.A., Godoy, M., Araya, M., Ganuza, I., Pino, R., Álvarez, F., Rengel, J., Hernández, C., Uribe, E., Blanco, J., 2019. Paralytic Shellfish Toxins in Surf Clams *Mesodesma donacium* during a Large Bloom of *Alexandrium catenella* Dinoflagellates Associated to an Intense Shellfish Mass Mortality. Toxins (Basel) 11, 188.

AOAC, 1995. Paralytic Shellfish Poison, Biological method. Final Action. Method 959.08, Official Methods of Analysis of the Association of Official Analytical Chemists (AOAC), Chapter 49, Natural Toxins, p. 46.

AOAC, 2005. Official Method 2005.06. Paralytic Shellfish Poisoning Toxins in Shellfish. Prechromatographic Oxidation and Liquid Chromatography with fluorescence detection. First Action 2005. J AOAC INT 88, 1714.

Ben-Gigirey, B., Rodriguez-Velasco, M.L., Gago-Martinez, A., 2012. Extension of the validation of AOAC Official Method 2005.06 for dc-GTX2,3: interlaboratory study. J AOAC INT 95, 111-121.

EC, 1996. Commission Decision of 18 January 1996 establishing the conditions for the harvesting and processing of certain bivalve molluscs coming from areas where the paralytic shellfish poison level exceeds the limit laid down by Council Directive 91/492/EEC (96/77/EC). Off. J. Eur. Communities L 15, 46-47.

EC, 2004. REGULATION (EC) No 853/2004 OF THE EUROPEAN PARLIAMENT AND OF THE COUNCIL of 29 April 2004 laying down specific hygiene rules for food of animal origin. Off. J. Eur. Communities L 226, 55-205.

EC, 2017. Commission Regulation (EU) 2017/1980 of 31 October 2017 amending Annex III to Regulation (EC) No 2074/2005 as regards paralytic shellfish poison (PSP) detection method. Off. J. Eur. Communities L 285, 8-9.

EFSA, E.F.S.A., 2009. Marine biotoxins in shellfish – Saxitoxin group. EFSA Journal 7, 1019.

EURLMB, 2019. AOAC 2005.06 Standard Operating Procedure. Simplified version for the screening and semi-quantitation of PSP toxins. Available at http://www.aecosan.msssi.gob.es/en/CRLMB/docs/docs/metodos analiticos de desarrollo/S OP AOAC 2005 06 semi.pdf.

FAO, 2004. Marine Biotoxins. In FAO Food and Nutrition Paper 80. p. 295. Available at <u>http://www.fao.org/3/y5486e/y5486e00.htm</u> in: F.a.A. Organization (Ed.), Rome, Italy.

Fernández, M.L., 1998. Phycotoxins: Regulatory Limits and Effects on Trade. , in: M. Miraglia, H.P.v. Egmond, C. Brera, J. Gilbert (Eds.), Mycotoxins and Phycotoxins – Developments in Chemistry, Toxicology and Food Safety. Alaken, Inc., Fort Collins, Colorado, pp. 503-516.

Fernández, M.L., Shumway, S.E., J., B., 2003. Managementof shellfish resources., in: H. M., A.A. D., A.D. M. (Eds.), Manual on Harmful Marine Microalgae. UNESCO Publishing, Paris, pp. 657-692.

Hatfield, R.G., Punn, R., Algoet, M., Turner, A.D., 2016. A Rapid Method for the Analysis of Paralytic Shellfish Toxins Utilizing Standard Pressure HPLC: Refinement of AOAC 2005.06. Journal of AOAC International 99, 475-480.

Lawrence, J.F., Maher, M., Watson-Wright, W., 1994. Effect of cooking on the concentration of toxins associated with paralytic shellfish poison in lobster hepatopancreas. Toxicon 32, 57-64.

Lawrence, J.F., Niedzwiadek, B., Menard, C., 2005. Quantitative determination of paralytic shellfish poisoning toxins in shellfish using prechromatographic oxidation and liquid chromatography with fluorescence detection: collaborative study. J AOAC INT 88, 1714-1732.

Lee, R., Lovatelli, A., Ababouch, L., 2008. Bivalve Depuration: Fundamental and Practical Aspects. FAO. Fisheries Technical Paper No 511. Rome, FAO., 139.

Reboreda, A., Lago, J., Chapela, M.-J., Vieites, J.M., Botana, L.M., Alfonso, A., Cabado, A.G., 2010. Decrease of marine toxin content in bivalves by industrial processes. Toxicon 55, 235-243.

Reis Costa, P., Braga, A.C., Turner, A.D., 2018. Accumulation and Elimination Dynamics of the Hydroxybenzoate Saxitoxin Analogues in Mussels Mytilus galloprovincialis Exposed to the Toxic Marine Dinoflagellate *Gymnodinium catenatum*. Toxins (Basel) 10, 428.

Turner, A.D., Dhanji-Rapkova, M., Dean, K., Milligan, S., Hamilton, M., Thomas, J., Poole, C., Haycock, J., Spelman-Marriott, J., Watson, A., Hughes, K., Marr, B., Dixon, A., Coates, L., 2018. Fatal Canine Intoxications Linked to the Presence of Saxitoxins in Stranded Marine Organisms Following Winter Storm Activity. Toxins (Basel) 10, 94.

Turner, A.D., Norton, D.M., Hatfield, R.G., Morris, S., Reese, A.R., Algoet, M., Lees, D.N., 2009. Refinement and extension of AOAC Method 2005.06 to include additional toxins in mussels: single-laboratory validation. J AOAC INT 92, 190-207.

Vieites, J.M., Botana, L.M., Vieytes, M.R., Leira, F.J., 1999. Canning Process that Diminishes Paralytic Shellfish Poison in Naturally Contaminated Mussels (*Mytilus galloprovincialis*). Journal of Food Protection 62, 515-519.