An ecological niche modelling approach to assess present and future suitability areas of
 *Quercus coccifera* L. in the Levant under climate change scenarios

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#### 19 Abstract

20 *Ouercus coccifera* L. (Kermes oak) is an evergreen oak, typical of the maguis in the eastern and south eastern part of the Mediterranean. It occurs almost continuously along the Syrian-Lebanese 21 coast up to 1500 m, and more scattered inland, until the southernmost arid area of Petra in Jordan. 22 Coupled human impacts and global warming are strongly affecting the species natural distribution 23 leading to a widespread forest fragmentation in the whole region. In this study, we aim at 24 25 investigating the current bioclimatic range of Kermes oak, and forecast the potential suitability area to the 21<sup>st</sup> century. Ecological niche modelling was used to retrieve the environmental envelope of 26 the species according to 23 topographic and climate variables. Five algorithms and three General 27 28 Circulation Models were applied to provide the potential distribution of Kermes oak at present time and project it to future. Results pointed out a current suitability area in the Middle East extending 29 from NW of Syria, rather continuously along the Lebanese coasts and inland until the 30 31 Mediterranean western slopes of Palestine and the Golan area (Israel), encompassing the Jordan Valley toward Dana and Wadi Rum (Jordan), with an isolated patch in Jabal Al-Arab (South Syria). 32 Future scenarios depict a significant fragmentation and restriction of Kermes oak range, especially 33 in the north of Syria and Golan, with a general shifting in altitude. These information might 34 facilitate the foresters to cope with the challenge of climate changes by identifying the most suitable 35 36 areas climatically effective for successful ecosystem restoration, including reforestation programmes. 37

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Keywords: *Quercus coccifera*, Ecological Niche Modelling, Climate change, Forest conservation
and reforestation, Middle East.

#### 41 Introduction

42 The portion of the Mediterranean biome of the Old World spans the coastal and the inland of the homonymous Basin, and it is recognized as one of the most threatened hotspot of biodiversity 43 worldwide (Underwood et al., 2009). The richness and the composition of the vegetation 44 assemblages are also the result of millennia of human interactions with the environment, especially, 45 in the eastern part of the Basin. The most typical form of vegetation increasingly affected by 46 anthropogenic impacts is the Mediterranean maquis, which hosts a considerable number of tree 47 species, including deciduous and evergreen oaks as key elements (Tomaselli, 1977). The genus 48 Quercus L. (Fagaceae) comprises around 30 species in the Euro-Mediterranean region(Govaerts and 49 50 Frodin, 1998, Denk and Grimm, 2010, Simeone et al., 2013), one of them being a prominent 51 evergreen component of the Levantine maquis: the Kermes oak (Quercus coccifera L.). Its global 52 distribution encompasses the Mediterranean Basin and it is the only one evergreen oak with such a range (Blondel et al., 2010, Toumi and Lumaret, 2010). Moreover, it defines the easternmost limit 53 of the Mediterranean maquis in the Middle East (Al-Eisawi, 1996), together with other few shrubs 54 and trees (e.g. Ceratonia siliqua L., Cercis siliquastrum L., Laurus nobilis L., Olea europaea L., 55 Phillyrea latifolia L., Pinus halepensis Mill., Pistacia palaestina Boiss., Quercus ithaburensis 56 Decne., Rhamnus alaternus L., and Styrax officinalis L.). Although oaks are deeply investigated 57 58 from decades, their taxonomy is still a matter of debate producing a long list of subspecies, varieties, ecotypes and synonyms. Kermes oak is not excluded from this contest and the two latest 59 reviews of its nomenclature lead to the following main classifications: one single species, namely 60 61 Quercus coccifera L., with a plethora of synonyms (e.g. O. calliprinos Webb, O. palaestina Kotschy) (Govaerts and Frodin, 1998), and one species with two distinct subspecies, i.e. Quercus 62 coccifera subsp. coccifera (distributed in Europe and North Africa), and Quercus coccifera subsp. 63 64 calliprinos (distributed in Cyprus, Anatolia, Middle East) (Tutin et al., 2001, Menitsky, 2005). In this paper we will refer to Kermes oak as Quercus coccifera L. according to the currently accepted 65 66 species list in Govaerts and Frodin (1998). A further reason to consider the species as a single entity relies on the genetic results by Toumi and Lumaret (2010) who identify some differences only todistinguish two morphotypes.

Focusing on the Middle East, Q. coccifera can be found in pure stands or mixed with other eastern 69 entities, for example Arbutus andrachne L., Pinus brutia Ten., Pistacia palaestina Boiss., Quercus 70 71 infectoria subsp. boissieri (Reuter) O. Schwarz, Quercus ithaburensis Decne., Styrax officinialis L. (Whyte, 1950, Danin, 1992, Danin, 2001, Ghattas et al., 2005, Jomaa et al., 2009, Al-Eisawi, 2012). 72 It occurs on poor and rocky areas on different parent rocks marl, limestone, basalt and green rocks, 73 on terra rossa, rendzina, sandy loam, and even on some podzolic soils (Nahal et al., 1989, Nahal and 74 Zahoueh, 2005), from the Thermo-mediterranean (coastal) to the Presteppic Supra-mediterranean 75 76 zone (up to 1800 m a.s.l.) (Zohary, 1960, Abi-Saleh et al., 1976, Al-Eisawi, 1996, UNDP, 2011). 77 Despite its widespread distribution from the northwest Syria to the Ma'an Governorate in the Petra 78 area (Jordan), and from the Mediterranean coast to the surrounding of As Suwayda (southern inland 79 of Syria), the Kermes oak forests are suffering from degradation by human impacts: population growth, the creeping of urban areas and cultivated lands, overgrazing, forest fires, coal production 80 and pollution are some of the threats affecting the natural occurrence of the species, with 81 consequences also on the phytosociological structure, as most of the climax vegetation has 82 disappeared (ARIJ, 2007, Jomaa et al., 2008). 83

84 In addition, the increasing frequency of climate extremes, augments the stress on those ecosystems and consequently aggravate the effects of the abovementioned impacts. In fact, climate changes are 85 widely acknowledged as the major drivers for adverse consequences on terrestrial and marine 86 87 ecosystems (Bellard et al., 2012, Alberto et al., 2013), thus understanding the framework of climatespecies interactions is of paramount importance to assess the vegetation dynamics and its related 88 feedbacks on biotic and abiotic factors (Diaz et al., 2007, Thuiller et al., 2011, Bellard et al., 2012). 89 90 The eastern part of the Mediterranean Basin is already experiencing a strong increase in temperature and drought, with ongoing effects on forests and ecosystem services (Kelley et al., 2015), high risk 91

92 of massive species extinction, landscape modifications (Kitoh et al., 2008), habitat fragmentation,

over-exploitation and invasive alien species diffusion (UNDP, 2011). The latest projections for the
21<sup>st</sup> century are not encouraging, especially for the Levant, which is predicted to undergo a severe
surface reduction of the Mediterranean vegetation (Klausmeyer and Shaw, 2009). The situation is
even more exacerbate by human exploitation of the natural resources (ARIJ, 2007, Jomaa et al.,
2008), civil disorders and geopolitical instability (Schoenfeld, 2010, Hens, 2012).

In view of this, novel prosing approaches helpful to provide information for decision-makers are welcomed. The Ecological Niche Modelling (ENM) might be considered an effective method to create maps of predicted suitable areas of a target vegetation unit (e.g. species, coenoses, ecosystems), dealing with responses to present climate features and future global warming scenarios. Several positive feedbacks might outcome from the application of ENM for conservation, reforestation and management purposes (Parmesan, 2006, Hidalgo et al., 2008, Vessella and Schirone, 2013, Vessella et al., 2015).

The present study is the first application of different ENM algorithms targeting the occurrence of *Quercus coccifera* in the east Mediterranean Basin to its easternmost extent. More specifically, it aims at (i) predicting the species current potential distribution, (ii) forecasting its suitability areas in the 21<sup>st</sup> century under different greenhouses emissions scenarios, and (iii) evaluating the effects of climate change on species bioclimatic range stability, contraction or extension of its area of occupancy, and (iv) identify accordingly proper sites for in situ conservation and ecosystem restoration activities.

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# 113 Materials and Methods

114 Study area and data collection

In this paper, we focused on the Levant countries facing the easternmost part of the Mediterranean
Basin where a Csa/Csb climate rules (Peel et al., 2007), namely: Syria, Lebanon, Israel, State of
Palestine, Jordan (the same area is also known as Bilād al-Shām, بلاد الشام in Arabic).

The actual distribution of Kermes oak in the study area was puzzled out using heterogeneous data 118 from different but mostly updated sources: Lebanon Reforestation Initiative (LRI)and Forest Map 119 of Lebanon (FAO and MoA, 2005) for Lebanon, Reinforcing Capacity Building for Defending 120 Biodiversity in the Palestinian Territories (DEBPAL2 EU Project) for Palestine, Israel Biodiversity 121 122 Information System (BioGIS) for Israel and Palestine, bibliographic research and field surveys for Jordan, Syria, Palestine (Zohary, 1960, Al-Eisawi, 1996, ARIJ, 2007, Ghazal, 2008, Jomaa et al., 123 2009, Al-Eisawi, 2012, Jawarneh et al., 2012). The retrieved distribution was standardized into a 124 presence point dataset 30 arc-seconds resolution to match the ENMs requirements employed 125 hereafter and to avoid pseudo-replications or to lessen the effect of variation in sampling effort. A 126 127 total of 7,739 spatially unique points represent the actual distribution of *Q. coccifera* in the study 128 area (Figure 1a).

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#### 130 Environmental variables and modelling algorithms

Twenty-three environmental variables were chosen as main determinants to model the niche of Q. 131 coccifera based on its present distribution. In details, 19 climatic raster layers were retrieved from 132 WorldClim 1.4 (release 3) at 30 arc-seconds resolution to represent the actual climatic envelope 133 (Hijmans et al., 2005). Elevation was obtained from ASTER Global Digital Elevation Model 134 135 (http://gdem.ersdac.jspacesystems.or.jp) and re-scaled to30 arc-seconds resolution to correspond with WorldClim data; slope and aspect were handled from elevation using ArcMap 9.3.1. In 136 the Emberger Ouotient was calculated and employed to summarized the 137 addition. evapotranspiration and draughtiness in a single user-friendly parameter (Emberger, 1930, Daget, 138 1977). The full list of the environmental layers is shown in Table 1. Additional biotic and abiotic 139 factors (e.g. soil nutritional factors, species competition, regeneration patterns) were excluded 140 141 mainly because of lack of data at broad scale (Pearson and Dawson, 2003).

A selection of 5 ENM algorithms commonly employed in modelling studies were applied to define
the potential suitability range for *Q. coccifera* based on its present distribution and the current

climate conditions. The list of models includes 'machine learning techniques' (Genetic Algorithm 144 for Rule set Prediction - GARP; Maximum Entropy - MaxEnt), 'multivariate analysis' (Climate 145 Space Model – CSM) and 'profile methods' (BIOCLIM; Envelope Score). All algorithms were run 146 using openModeller 1.5.0 with default options and only presence-data with pseudo-absence points 147 randomly generated from the background (Muñoz et al., 2011). Each algorithm produced a raster 148 map 30 arc-seconds resolution as output, showing the suitability areas for the study species in terms 149 of probability of occurrence ranging from 0 (no suitability) to 100% (optimal conditions). The 150 goodness-of-fit of the results was evaluated toward the generated confusion matrix, which counts 151 for the observed and predicted presence/absence events. The Receiver Operating Characteristic 152 153 (ROC) curve approach was followed, and the Area Under the Curve AUC was calculated to 154 estimate the prediction success (Fielding and Bell, 1997). Correlation and similarity among the models outputs were also assessed by Principal Component Analysis (PCA) and UPGMA 155 Clustering (Diniz-Filho et al., 2010). Once the statistical robustness was assessed, the level of 156 agreement among the models was computed by calculating the weighted average of the raster value 157 per grid cell, taking into account the AUC value of each model (Vessella et al., 2015). 158

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#### 160 Forecasting projections under multiple scenarios

Predictive modelling is becoming an important tool in ecology, biogeography, conservation and management of tree species, and an increase number of studies is focusing on prominent hotspots of biodiversity, such as the Mediterranean Basin (Thuiller et al., 2005, Benito Garzon et al., 2008, Hidalgo et al., 2008, Casazza et al., 2014, López-Tirado and Hidalgo, 2014).

165 The consensus map pointed out at the current time from the five models was used to assess and 166 extract the potential climatic niche of *Q. coccifera* to be later projected under future climate 167 scenarios in the  $21^{st}$  century.

168 The depiction of the mechanistic links of variables affecting by the global change was attempted 169 and climate projections from three downscaled and bias corrected GCMs (Global Circulation

Models, CCSM4, HADGEM2-ES and MIROC5) were retrieved from WorldClim website (original 170 data from CMIP Phase 5 home page, http://cmip-pcmdi.llnl.gov/cmip5). Two greenhouse gas 171 emission scenarios based on Representative Concentration Pathways (RCP) were considered for 172 each GCM, namely: RCP 4.5 (intermediate emission scenarios achieving an impact of 4.5 watts per 173 square metre by 2100) and RCP 8.5 (hard emission scenario accomplishing an increase of 8.5 watts 174 per square metre by 2100) (van Vuuren et al., 2011, IPCC, 2014). Changes were investigated 175 separately for two temporal frameworks representing the 21<sup>st</sup> century: 2041-2060 (average on 2050) 176 and 2061-2080 (average on 2070). Topographic variables were excluded from the forecasting 177 procedure because future changes are expected in the bioclimatic features linked to topography and 178 179 not topography per se. Undeniably, those variables would contribute to a robust current distribution, 180 but they would alter future projections (Kumar, 2012).

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#### 182 **Results**

#### 183 *Quercus coccifera present potential distribution*

The five algorithms employed gave back potential distribution maps with discriminatory capacity 184 far from random results, as stated by high AUC values over 0.88. MaxEnt provided the most robust 185 prediction at present, with the smallest suitability area, while CSM and Envelope Score yielded 186 187 lager areas but weaker goodness-of-fit (Table 2). The Spearman matrix pointed out strong positive correlation among the cell probability values calculated by MaxEnt, GARP and BIOCLIM, while 188 Envelope Score and CSM slightly diverge (Table 3). This pattern is reinforced by the dendrogram 189 190 of similarity, which showed an asymmetric clustered tree with two subgroups (Figure 2). The 191 uncertainties related to the intrinsic structures of the differences among models were also assessed by PCA (Figure 2). The first principal component explains 73.78% of the correlation structure, and 192 the factors values mapped in Figure 2 depict the level of agreement among the models in the same 193 areas where they match with the highest probabilities. In addition, the plot of loadings shows that 194 195 the five models are similarly oriented along the first axes, thus tending to be almost analogous.

Thus, the majority of discrepancy pointed out is related to the relative position of 'MaxEnt - GARP 196 197 - BIOCLIM' and 'Envelope Score - CSM' groups along the second axes, which explains only 12.99% of the correlation (Figure 2). Despite such a divergence among the models about the extent 198 of the predicted habitable surface and the probability magnitude associated to each cell (Figure S1), 199 the majority consensus map mostly disentangles those sources of variation by producing a more 200 conservative prediction map. This solution also reinforces the high suitability for *Q. coccifera* in the 201 202 coastal part of the study area, further inland in Syria (Aleppo and Idlib District), in the Golan region and around the Jordan Valley slopes (Figure 1b; cf. the physical map in Figure S2). The present 203 predicted habitable surface is about 38,300 km<sup>2</sup> mostly distributed within Csa/Csb climates extent, 204 205 and representing the theoretical continuous species range, part of which is occupied by the actual 206 distribution.

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## 208 *Quercus coccifera distribution in the 21st century*

The three-tested General Circulation Models project a future potential distribution of Q. coccifera 209 progressively reduced in extent across the 21st century (Figure 3). Such a contraction is also 210 different in magnitude, depending on the GCM and the considered emission scenario. Overall, two 211 areas seem to undergo a strong change in Q. coccifera occurrence: the inland part of northern Syria 212 213 and the Golan region. The intermediate emissions (RCP 4.5) would affect the species suitability in the lowlands, on the coasts and in the eastern slopes of the Jordan Valley. If we focus on the surface 214 with probability values over 50%, Table 3 shows a global reduction ranging from about 4,000 to 215 216 24,000 km<sup>2</sup>, geographically displayed in Figure 3. Among the GCMs, HADGEM2-ES resulted to be the most severe, CCSM4 the most lenient. This pattern is confirmed both in 2050 and in 2070. 217 Where the species is still predicted to occur, the associated probability values are generally lower; 218 this is true especially around the Jordan Valley and the Western Mediterranean slopes. The high 219 emissions scenario (RCP 8.5) confirms the abovementioned trend, projecting a more marked 220 reduction and range fragmentation. For example, the southern part of the species potential 221

distribution is predicted to mostly disappear according to HADGEM2-ES (Israel, State of Palestine and Jordan), as well as in Syria where the Aleouite Mts (Jabal An Nusayriyah reliefs)would host the last remnants of the largest habitable area predicted at the present time. The estimated loss of potential area under the most pessimistic scenario would range from about 18,000 to 30,000 km<sup>2</sup>. Nevertheless, a core area placed around the Mt Lebanon, Mt Hermon and Anti-Lebanon Mts resulted to survive to every future scenario here investigated.

Notably, the bioclimatic envelope retrieved from the niche modelling at the present condition would not be altered in time. Most of the variables used in this study would nearly keep constant their values ranges with the exception of the Emberger Quotient and Annual Precipitation - Bio12 (upper limit), Maximum Temperature of the Warmest Month - Bio5 and Temperature Annual Range -Bio7 (lower limit). The major change in those ranges is observed in elevation, resulted in an upward shift of a minimum value of 350 (CCSM4, RCP4.5 in 2050) up to a maximum of 700 m (HADGEM2-ES, RCP 8.5 in 2070) (Table 3).

In view of this, the range dynamics of Q. coccifera affected by climate change would lead to an 235 unbalanced variation when projected to the 21<sup>st</sup> century (Figure 4 and Table 5). The analysis 236 focused on the probability values over 50% reveals as the gained surface concentrated around the 237 Lebanon and Anti-Lebanon Mts, Hermon Mt, Damascus and As Suwayda Districts, Petra region, 238 239 scarcely replaces the large loss of suitable area observed in the remaining parts of the potential species range. This pattern is reflected into a stable surface, with respect to the present area, of 240 about 39.9 - 67.9% in 2050, reduced up to 11.7 - 23.8% in 2070 (RCP4.5); as expected, under the 241 242 RCP 8.5 the stable surface is further reduced up to 10.9 - 17.6% in 2050, and 3 - 11.9% in 2070. Globally, every scenario depicts a negative variation of Q. coccifera future potential distribution 243 with extreme values from -47.6% to -83.6% (HADGEM2-ES in 2050 under RCP4.5, and in 2070 244 245 under RCP8.5, respectively).

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#### 247 Discussion

The output map of the potential distribution of *Q. coccifera* at the present climate conditions confirms as the species could encompass the whole Levant defined by a Mediterranean climate, with the exception of the highest reliefs where deciduous broadleaves and conifers dominate (e.g. the upper part of Lebanon Mts. and Anti-Lebanon) (Abi-Saleh et al., 1976, Blondel et al., 2010).

Moreover, our findings are consistent with recent updates on vegetation decline by global warming 252 in the Middle East (Kitoh et al., 2008, Klausmeyer and Shaw, 2009). A general reduction of the 253 254 Mediterranean environment, threaten by desertification and expansion of arid and semi-arid regions has been stated (IPCC, 2014) and striking contractions with no suitable expansion areas nearby 255 might further limit the species capacity to persist in the Levant. The GCMs and the emissions 256 257 scenarios considered in this study evidence a coherent response of the species in terms of suitable 258 area that would be reduced especially in the lowlands, with a consequent constraint and upward shift at higher elevations (Figure 4, 5; Table 4). This phenomenon has been already assessed for 259 260 other species with recent climate change (Parmesan, 2006, Kelly and Goulden, 2008, Lenoir et al., 2008) and rather stressed under future global warming (Hayhoe et al., 2004, Gonzalez et al., 2010). 261 According to our results, higher temperatures, drought and evapotranspiration (synthesized by the 262 Emberger Quotient) are the main drivers for such an upward shift by Q. coccifera and they would 263 affect the species adaptability as well (Rever et al., 2013, Bussotti et al., 2014). 264

265 Water deficits and heat stress resulted from prolonged drought period might be predicted to induce extensive tree mortality (Allen et al., 2010, Sarris et al., 2011), produce ecophysiological and 266 phenological shifts (Gordo and Sanz, 2010), alter the seasonal cycles of insects and pathogens 267 268 (Rafferty et al., 2015), modify soil thermal cycles and soil moisture (Petroselli et al., 2013), limit seed production (Connolly and Orrock, 2015). Since the induced spatiotemporal modification by 269 270 climate mainly strikes the southernmost part of the species range (Alberto et al., 2013, Bussotti et al., 2014), the vegetation composition, the inter- and intraspecific competition and the biogeography 271 of forests in the Levant might be drastically undergoing a radical change (Kelly and Goulden, 272 273 2008). Under another perspective, the upward shift of Q. coccifera and related species in response

to warming may induce a progressive isolation and degradation of Montane Mediterranean and 274 275 Oro-Mediterranean forest ecosystems up to replacement. This phenomenon has been already observed in many parts of the world pointing out cascading effects leading to habitat and biome 276 reassessment even at broad scale (Parmesan and Yohe, 2003, Peñuelas and Boada, 2003, Jump et 277 al., 2009). In view of this, the future distribution of Mediterranean and montane/temperate 278 ecosystems is strictly interlinked, and the results of modelling projections applied to a 279 280 Mediterranean species might be the starting point for evaluating the dynamics of the montane environments as well (Beniston, 2003, Sanz-Elorza et al., 2003, Xu et al., 2009, Sarris et al., 2011, 281 Ruiz-Labourdette et al., 2013). 282

283 The results achieved could be also informative to detect priority areas and future putative refuge for 284 Q. coccifera forests conservation, highlighting where to intervene and plan strategies to restrain the effects of global change. However, the results from the models might be mitigated and less 285 286 pessimistic for the future, due to some ecophysiological features of the species. Among them, the plasticity of the reproductive phenology with recurrent flowerings and cycles of acorns maturation 287 from less than one year up to two years (Bianco and Schirone, 1985), coupled with the high 288 resprouting capacity (e.g. Trabaud, 1991). The effectiveness of ENM approach to rethink the 289 network of protected areas and readdress the forest management has been already discussed 290 291 (Hannah et al., 2007, Hannah, 2008, Klausmeyer and Shaw, 2009, Loarie et al., 2009, Araújo et al., 2011), but this study represents the first attempt to Levant using several ENMs and GCMs. 292

All the scenarios considered depict a marked reduction in species suitability area, close to 50% in the most optimistic one (Table 5). The geographical location of such a contraction mainly interests the northern and southern limits of the potential range (Syria, Israel, State of Palestine and Jordan). In those areas, the species would seem to face a severe fragmentation or even run an extinction risk. Efforts to face forest fragmentation, to protect, recover and extend the species occurrence in those areas would seem a hard challenge, mainly due to geopolitical reasons, while species translocation and *ex situ* strategies might ensure at least an extreme conservation measure for those provenances, although it is essential to assess the risks beyond those actions and when the benefits outweigh the costs (Hunter, 2007, Thomas, 2011). On the other hand, *Q. coccifera* would persist and extend around the mountainous regions of Lebanon and Anti-Lebanon Mts (Figure 4). This area represents the present core of the species distribution in the Levant, and it could be considered as future putative refugia where to focus the *in situ* conservation programmes.

Overall, a multidisciplinary approach is welcomed, coupling for example ENMs, ecophysiological and genetic studies to better understand the processes behind *Q. coccifera* adaptation, survival, seed dispersion capacity, gene flow, local phenology, etc. This might help to evaluate the repercussions of a changing climate for the species, and to secure its future persistence in the Levant by means of appropriate guidelines and recommendations for the decision-makers.

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Table 1. Environmental variables used in this study to model present potential distribution of *Q. coccifera*.
All of them are raster data at 30 arc-seconds of resolution. Asterisks indicate those layers not used in
forecasting modelling (see Materials and Methods).

Variable	Label	Source
Annual Mean Temperature (°C)	Bio1	WorldClim
Mean Diurnal Range (°C)	Bio2	WorldClim
Isothermality (Bio2/Bio7) x 100	Bio3	WorldClim
Temperature Seasonality (SDx100) (°C)	Bio4	WorldClim
Max Temperature of Warmest Month (°C)	Bio5	WorldClim
Min Temperature of Coldest Month (°C)	Bio6	WorldClim
Temperature Annual Range (°C)	Bio7	WorldClim
Mean Temperature of Wettest Quarter (°C)	Bio8	WorldClim
Mean Temperature of Driest Quarter (°C)	Bio9	WorldClim
Mean Temperature of Warmest Quarter (°C)	Bio10	WorldClim
Mean Temperature of Coldest Quarter (°C)	Bio11	WorldClim
Annual Precipitation (mm)	Bio12	WorldClim
Precipitation of Wettest Month (mm)	Bio13	WorldClim
Precipitation of Driest Month (mm)	Bio14	WorldClim
Precipitation Seasonality (Coeff. of variation)	Bio15	WorldClim
Precipitation of Wettest Quarter (mm)	Bio16	WorldClim
Precipitation of Driest Quarter (mm)	Bio17	WorldClim
Precipitation of Warmest Quarter (mm)	Bio18	WorldClim
Precipitation of Coldest Quarter (mm)	Bio19	WorldClim
Emberger Quotient	Emb	This work
Elevation (m) *	DEM	ASTERGDEM
Slope (%) *	SL	Derived from DEM
Aspect (degrees) *	AS	Derived from DEM
Aspect (degrees) *	AS	Derived from DEM

Table 2. Statistical scores for the goodness-of-fit of the algorithms used in this paper. Values of predicted
surfaces refer to the full range of suitability (1-100%).

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	AUC	Surface predicted presence (km <sup>2</sup> )	Accuracy (%)	Commission error
Bioclim	0.92	71,913	100.00	2.1x10 <sup>-3</sup>
Climate Space Model	0.88	137,449	59.74	0.056
Envelope Score	0.89	332,794	100.00	1.54x10 <sup>-3</sup>
GARP – Best Subsets	0.93	52,238	96.54	0.017
Maximum Entropy	0.97	34,297	81.82	0.035

Algorithm	BIOCLIM	CSM	ENVSCORE	GARP	MAXENT
BIOCLIM	1	0.603	0.714	0.843	0.821
CSM	0.603	1	0.596	0.545	0.476
ENVSCORE	0.714	0.596	1	0.587	0.599
GARP	0.843	0.545	0.587	1	0.879
MAXENT	0.821	0.476	0.599	0.879	1

**Table 3.** Spearman correlation matrix resulted from PCA analysis among the algorithms.

Table 4. Ranges of the environmental layers (minimum-maximum) of *Q. coccifera* over the threshold (50%) for each studied scenario in the present and future.

Topographic variables are also included to quantify the species shifting in altitude, orientation and slope. Surfaces refer to the global predicted areas with suitability values over 50%.

RCP 4.5 (moderate emissions)							RCP 8.5 (hard emissions)						
			2050			2070			2050			2070	
	Present	CCSM4	HADGEM2-ES	MIROC5	CCSM4	HADGEM2-ES	MIROC5	CCSM4	HADGEM2-ES	MIROC5	CCSM4	HADGEM2-ES	MIROC5
Bio1	10.8/21.3	10.8/21.3	11.1/21.3	11.2/21.3	10.9/21.4	11.2/21.4	11.4/21.3	10.9/21.4	11.3/21.3	11.4/21.3	10.9/21.5	11.9/21.4	11.7/21.3
Bio2	7.7/13.4	7.7/13.4	7.8/13.4	7.7/13.4	7.8/13.5	7.9/13.5	7.8/13.4	7.7/13.6	7.9/13.4	7.8/13.5	7.7/13.7	8/13.5	7.9/13.5
Bio3	31/48	31/46	31/43	31/45	31/47	31/43	31/44	31/45	31/43	31/44	31/44	31/41	31/42
Bio4	4782/8139	5070/8138	5421/8103	5479/8137	5106/8138	5444/8038	5538/8017	5214/8058	5521/7860	5446/8122	5176/7886	5694/7692	5659/7860
Bio5	24.7/35.8	24.8/35.8	25.7/35.8	25.5/35.8	24.7/35.9	26.3/35.8	26.4/35.8	24.9/35.9	26.3/35.8	25.7/35.8	24.9/36	27.5/35.9	26.2/35.8
Bio6	-2.1/9.1	-2.1/9.1	-2.1/9.1	-2.1/9.1	-2/9.2	-2/9.1	-2.1/9.2	-1.9/9.2	-2/9.1	-2/9.2	-1.9/9.3	-1.9/9.1	-1.9/9.2
Bio7	21.8/34.3	22.7/34.3	23.6/34.3	23.1/34.3	22.6/34.5	23.7/34.4	23.6/34.4	23/34.3	24/34.5	23.2/33.8	23/33.9	24.4/34.4	23.6/33.8
Bio8	2.5/13.9	2.6/13.9	2.5/13.9	2.5/13.8	2.5/14	2.5/14	2.5/13.9	2.6/14	2.5/14	2.5/13.9	2.7/14	2.8/14	2.6/13.9
Bio9	18.4/28.2	18.4/28.2	19.4/28.2	19.5/28.2	18.6/28.2	19.6/28.2	19.9/28.2	18.4/28.3	19.8/28.2	19.6/28.3	18.6/28.5	20.6/28.2	20.1/28.3
Bio10	18.6/28.3	18.6/28.3	18.4/28.3	19.5/28.3	18.7/28.5	19.6/28.3	19.9/28.3	19/28.5	19.8/28.3	19.5/28.4	19/28.6	20.7/28.6	20.2/28.2
Bio11	2.5/13.8	2.5/13.8	2.5/13.8	2.5/13.8	2.5/13.8	2.5/13.8	2.5/13.8	2.5/13.8	2.5/13.8	2.5/13.8	2.5/13.8	2.7/13.8	2.6/13.8
Bio12	205/1389	205/1312	221/1217	221/1306	205/1223	226/1187	219/1267	210/1108	221/1165	219/1296	223/1107	244/1083	234/1198
Bio13	45/299	45/291	45/285	45/299	45/284	45/297	48/293	45/245	45/292	46/284	47/259	52/272	52/299
Bio14	0/5	0/5	0/5	0/5	0/5	0/5	0/5	0/4	0/5	0/5	0/3	0/5	0/5
Bio15	73/114	74/113	75/115	74/115	76/110	77/109	75/111	82/108	81/115	80/110	82/110	82/117	81/109
Bio16	119/793	119/759	120/745	120/777	119/714	123/740	119/723	119/626	123/737	119/725	119/615	135/686	124/699
Bio17	0/32	0/32	0/29	0/31	0/32	0/31	0/30	0/24	0/30	0/31	0/32	0/26	0/31
Bio18	0/38	0/38	0/29	0/32	0/38	0/33	0/30	0/38	0/30	0/31	0/38	0/26	0/31
Bio19	117/791	119/759	120/745	120/777	120/714	123/740	119/723	117/626	123/737	120/725	119/614	135/686	124/694
Emb	23/202	23/182	23/146	23/155	23/161	23/133	23/136	23/137	23/129	23/144	23/132	23/107	23/130
DEM	2/1678	7/2019	178/2053	139/2001	7/2082	411/2202	377/2068	113/2068	457/2207	364/2131	405/2236	725/2384	581/2266
SL	0/22.4	0/22.3	0/22.3	0/22.3	0/22.3	0/22.3	0.22.4	0/22.3	0.22.3	0/22.3	0/22.4	0.22.3	0/22.3
AS	0/360	0/360	0./360	0/360	0/360	0/360	0/360	0/360	0/360	0/360	0/360	0/360	0/360
Surface (km <sup>2</sup> x10 <sup>3</sup> )	38.3	34.4	20.1	23.4	28.9	14.4	16.4	20.7	13.8	18.1	15.1	6.2	8.5

**Table 5.** Comparison among suitable surfaces with probability class over 50% at the present time and future scenarios. Values are given in  $\text{km}^2 \ge 10^3$  and in percentage (brackets) when referred to the forecasted portion of stable, gain, loss and net variation surfaces with respect to the present time. Bold values refer to the major expected variations, all of them belonging to HADGEM2-ES scenario.

		2050				2070				
		Stable	Gain	Loss	Variation	Stable	Gain	Loss	Variation	
N.	CCSM4	26.02 (67.9)	8.39 (21.9)	-12.26 (-32.0)	-3.87 (-10.1)	23.85 (62.3)	5.08 (13.3)	-14.43 (-37.3)	-9.35 (-24.4)	
SCP 4.	HADGEM2-ES	17.14 (36.9)	2.93 (7.6)	-21.14 (-55.2)	-18.21 ( <b>-47.6</b> )	11.70 (30.5)	2.7 (7.0)	-26.58 (-69.4)	-23.88 ( <b>-62.3</b> )	
R	MIROC5	19.31 (50.4)	4.09 (10.7)	-18.96 (-49.5)	-14.87 (-38.8)	13.65 (35.6)	2.71 (7.1)	-24.63 (-64.3)	-21.92 (-57.2)	
Ŋ	CCSM4	17.66 (46.1)	3.0 (7.8)	-20.62 (-53.8)	-17.62 ( <i>-46.0</i> )	11.97 (31.2)	3.11 (8.1)	-26.31 (-68.7)	-23.20 (-60.6)	
RCP 8.	HADGEM2-ES	10.96 (28.6)	2.83 (7.4)	-27.32 (-71.3)	-24.49 ( <b>-63.9</b> )	3.64 (9.5)	2.61 (6.8)	-34.64 (-90.4)	-32.03 ( <b>-83.6</b> )	
	MIROC5	14.73 (38.9)	3.35 (8.7)	-23.55 (-61.5)	-20.20 (-52.7)	5.69 (14.9)	2.82 (7.4)	-32.59 (-85.1)	-29.77 (-77.7)	
Present Time (reference surface value) = $38.30 \text{ km}^2 \times 10^3$										

## FIGURE 1



Quercus coccifera present potential distribution







FIGURE 3



RCP 4.5 (intermediate emissions scenario)

FIGURE 4



# 1 **References**

Abi-Saleh B, Barbero M, Nahal I, Quézel P. 1976. Les séries forestières de végétation au Liban Essai
 d'interprétation schématique. Bulletin De La Societe Botanique de France, 123: 541-560.

- 4 Al-Eisawi D. 1996. Vegetation of Jordan, UNESCO-Cairo office.
- 5 Al-Eisawi D. 2012. Conservation of Natural Ecosystems in Jordan. *Pakistan Journal of Botany*, 44: 95-99.
- Alberto FJ, Aitken SN, Alia R, Gonzalez-Martinez SC, Hanninen H, Kremer A, Lefevre F, Lenormand T,
   Yeaman S, Whetten R, Savolainen O. 2013. Potential for evolutionary responses to climate change
   evidence from tree populations. *Global Change Biology*, 19: 1645-1661.
- Allen CD, Macalady AK, Chenchouni H, Bachelet D, McDowell N, Vennetier M, Kitzberger T, Rigling A,
   Breshears DD, Hogg EH, Gonzalez P, Fensham R, Zhang Z, Castro J, Demidova N, Lim JH, Allard G,
   Running SW, Semerci A, Cobb N. 2010. A global overview of drought and heat-induced tree
   mortality reveals emerging climate change risks for forests. *Forest Ecology and Management*, 259:
   660-684.
- Araújo MB, Alagador D, Cabeza M, Nogués-Bravo D, Thuiller W. 2011. Climate change threatens European
   conservation areas. *Ecology Letters*, 14: 484-492.
- ARIJ. 2007. Status of the Environment in the Occupied Palestinian Territory, Jerusalem, Applied Research
   Institute of Jerusalem.
- Bellard C, Bertelsmeier C, Leadley P, Thuiller W, Courchamp F. 2012. Impacts of climate change on the
   future of biodiversity. *Ecology Letters*, 15: 365-377.
- Beniston M. 2003. Climatic change in mountain regions: a review of possible impacts. *Climate Variability* and Change in High Elevation Regions: Past, Present & Future. Springer.
- Benito Garzon M, Sanchez de Dios R, Sainz Ollero H. 2008. Effects of climate change on the distribution of
   Iberian tree species. *Applied Vegetation Science*, 11: 169-178.
- Bianco P, Schirone B. 1985. On Quercus coccifera L. sl: variation in reproductive phenology. *Taxon*: 436-439.
- Blondel J, Aronson J, Bodiou JY, Boeuf G. 2010. The Mediterranean region: Biological Diversity in Space and
   Time, Oxford, Oxford University Press.
- Bussotti F, Ferrini F, Pollastrini M, Fini A. 2014. The challenge of Mediterranean sclerophyllous vegetation
   under climate change: From acclimation to adaptation. *Environmental and Experimental Botany*,
   103: 80-98.
- Casazza G, Giordani P, Benesperi R, Foggi B, Viciani D, Filigheddu R, Farris E, Bagella S, Pisanu S, Mariotti
   MG. 2014. Climate change hastens the urgency of conservation for range-restricted plant species in
   the central-northern Mediterranean region. *Biological Conservation*, 179: 129-138.
- Connolly B, Orrock J. 2015. Climatic variation and seed persistence: freeze-thaw cycles lower survival via
   the joint action of abiotic stress and fungal pathogens. *Oecologia*: 1-8.
- 36 Daget P. 1977. Le bioclimat méditerranéen: analyse des formes climatiques par le système d'Emberger.
   37 Vegetatio, 34: 87-103.
- 38 Danin A. 1992. Flora and vegetation of Israel and adjacent areas. *Bocconea*, **3**: 18-42.
- 39 Danin A. 2001. Near East: Ecosystems, Plant Diversity. Encyclopedia of Biodiversity, 4: 353-364.
- Denk T, Grimm GW. 2010. The oaks of western Eurasia: Traditional classifications and evidence from two
   nuclear markers. *Taxon*, 59: 351-366.
- Diaz S, Lavorel S, de Bello F, Quetier F, Grigulis K, Robson M. 2007. Incorporating plant functional diversity
   effects in ecosystem service assessments. *Proceedings of the National Academy of Sciences of the* United States of America, 104: 20684-20689.
- Diniz-Filho JAF, Ferro VG, Santos T, Nabout JC, Dobrovolski R, De Marco P. 2010. The three phases of the
   ensemble forecasting of niche models: geographic range and shifts in climatically suitable areas of
   Utetheisa ornatrix (Lepidoptera, Arctiidae). *Revista Brasileira De Entomologia*, 54: 339-349.
- 48 **Emberger L. 1930**. Sur une formule climatique applicable en géographie botanique.
- 49 FAO, MoA. 2005. National Forest and Tree Assessment and Inventory. Final Report TCP/LEB/2903.

- Fielding AH, Bell JF. 1997. A review of methods for the assessment of prediction errors in conservation
   presence/absence models. *Environmental Conservation*, 24: 38-49.
- Ghattas R, Hrimat N, Isaac J. 2005. Palestine. In: Merlo M, Croitoru L eds. Valuing Mediterranean Forests:
   Towards Total Economic Value. Wallingford, UK/Cambridge, MA, USA, CABI International.

54 **Ghazal A. 2008**. Landscape Ecological, Phytosociological and Geobotanical Study of Eu-Mediterranean in 55 West of Syria, PhD. Thesis, University of Hohenheim, Stuttgart.

- Gonzalez P, Neilson RP, Lenihan JM, Drapek RJ. 2010. Global patterns in the vulnerability of ecosystems to
   vegetation shifts due to climate change. *Global Ecology and Biogeography*, 19: 755-768.
- Gordo O, Sanz JJ. 2010. Impact of climate change on plant phenology in Mediterranean ecosystems. *Global Change Biology*, 16: 1082-1106.
- Govaerts R, Frodin DG. 1998. World checklist and bibliography of Fagales. *Kew: Royal Botanic Gardens, Kew vii, 407p.-illus.. ISBN,* 1900347466.
- Hannah L. 2008. Protected areas and climate change. Annals of the New York Academy of Sciences, 1134:
   201-212.
- Hannah L, Midgley G, Andelman S, Araújo M, Hughes G, Martinez-Meyer E, Pearson R, Williams P. 2007.
   Protected area needs in a changing climate. *Frontiers in Ecology and the Environment*, 5: 131-138.
- Hayhoe K, Cayan D, Field CB, Frumhoff PC, Maurer EP, Miller NL, Moser SC, Schneider SH, Cahill KN,
   Cleland EE. 2004. Emissions pathways, climate change, and impacts on California. Proceedings of
   the National Academy of Sciences of the United States of America, 101: 12422-12427.
- 69 Hens L. 2012. Environmentally displaced people. *Encyclopedia of Life Support Systems (EOLSS)*, 2: 48-51.
- Hidalgo PJ, Marín JM, Quijada J, Moreira JM. 2008. A spatial distribution model of cork oak (Quercus suber) in southwestern Spain: A suitable tool for reforestation. *Forest Ecology and Management*, 255: 25-34.
- Hijmans RJ, Cameron SE, Parra JL, Jones PG, Jarvis A. 2005. Very high resolution interpolated climate
   surfaces for global land areas. *International Journal of Climatology*, 25: 1965-1978.
- Hunter ML. 2007. Climate change and moving species: furthering the debate on assisted colonization.
   *Conservation Biology*, 21: 1356-1358.
- **IPCC. 2014.** Climate Change 2014: Synthesis Report. In: Pachauri RK, Meyer LA eds. Contribution of Working
   Groups I, II and III to the Fifth Assessment Report of the Intergovernmental Panel on Climate
   Change. Geneva (Switzerland), IPCC.
- Jawarneh MS, Brake MH, Muhidat R, Migdadi HM, Lahham JN, El-Oqlah AE. 2012. Characterization of
   Quercus Species Distributed in Jordan Using Molecular Markers. International Conference on
   Applied Life Sciences. Turkey, InTech.
- Jomaa I, Auda Y, Hamze M, Abi saleh B, Safi S. 2009. Analysis of Eastern Mediterranean Oak Forests Over
   the Period 1965-2003 Using Landscape Indices on a Patch Basis. Landscape Research, 34: 105-124.
- Jomaa I, Auda Y, Saleh BA, Hamze M, Safi S. 2008. Landscape spatial dynamics over 38 years under natural
   and anthropogenic pressures in Mount Lebanon. *Landscape and Urban Planning*, 87: 67-75.
- Jump AS, Matyas C, Penuelas J. 2009. The altitude-for-latitude disparity in the range retractions of woody
   species. *Trends in Ecology & Evolution*, 24: 694-701.
- Kelley CP, Mohtadi S, Cane MA, Seager R, Kushnir Y. 2015. Climate change in the Fertile Crescent and
   implications of the recent Syrian drought. *Proceedings of the National Academy of Sciences*, 112:
   3241-3246.
- Kelly AE, Goulden ML. 2008. Rapid shifts in plant distribution with recent climate change. *Proceedings of the National Academy of Sciences of the United States of America*, 105: 11823-11826.
- Kitoh A, Yatagai A, Alpert P. 2008. First super-high-resolution model projection that the ancient" Fertile
   Crescent" will disappear in this century. *Hydrological Research Letters*, 2: 1-4.
- Klausmeyer KR, Shaw MR. 2009. Climate Change, Habitat Loss, Protected Areas and the Climate
   Adaptation Potential of Species in Mediterranean Ecosystems Worldwide. *Plos One*, 4.
- 98 Kumar P. 2012. Assessment of impact of climate change on Rhododendrons in Sikkim Himalayas using
   99 Maxent modelling: limitations and challenges. *Biodiversity and Conservation*, 21: 1251-1266.
- Lenoir J, Gegout JC, Marquet PA, de Ruffray P, Brisse H. 2008. A significant upward shift in plant species
   optimum elevation during the 20th century. *Science*, 320: 1768-1771.

- Loarie SR, Duffy PB, Hamilton H, Asner GP, Field CB, Ackerly DD. 2009. The velocity of climate change.
   *Nature*, 462: 1052-1055.
- López-Tirado J, Hidalgo PJ. 2014. A high resolution predictive model for relict trees in the Mediterranean mountain forests (Pinus sylvestris L., P. nigra Arnold and Abies pinsapo Boiss.) from the south of
   Spain: A reliable management tool for reforestation. *Forest Ecology and Management*, 330: 105 114.
- 108 **Menitsky YL. 2005**. *Oaks of Asia,* Enflied, New Hampshire, USA, Science Publishers.
- Muñoz MED, de Giovanni R, de Siqueira MF, Sutton T, Brewer P, Pereira RS, Canhos DAL, Canhos VP.
   2011. openModeller: a generic approach to species' potential distribution modelling.
   *Geoinformatica*, 15: 111-135.
- Nahal I, Rahma A, Chalabi MN. 1989. Forest and Forest Nurseries, Aleppo, Books and published
   department of Aleppo University, Faculty of Agriculture.
- Nahal I, Zahoueh S. 2005. Syria. In: Merlo M, Croitoru L eds. Valuing Mediterranean Forests: Towards Total
   *Economic Value.* Wallingford, UK/Cambridge, MA, USA, CABI International.
- Parmesan C. 2006. Ecological and evolutionary responses to recent climate change. Annual Review of
   *Ecology Evolution and Systematics*, 37: 637-669.
- Parmesan C, Yohe G. 2003. A globally coherent fingerprint of climate change impacts across natural
   systems. *Nature*, 421: 37-42.
- Pearson RG, Dawson TP. 2003. Predicting the impacts of climate change on the distribution of species: are
   bioclimate envelope models useful? *Global Ecology and Biogeography*, 12: 361-371.
- Peel MC, Finlayson BL, McMahon TA. 2007. Updated world map of the Koppen-Geiger climate
   classification. *Hydrology and Earth System Sciences*, 11: 1633-1644.
- Peñuelas J, Boada M. 2003. A global change-induced biome shift in the Montseny mountains (NE Spain).
   Global Change Biology, 9: 131-140.
- Petroselli A, Vessella F, Cavagnuolo L, Piovesan G, Schirone B. 2013. Ecological behavior of Quercus suber
   and Quercus ilex inferred by topographic wetness index (TWI). *Trees-Structure and Function*, 27:
   1201-1215.
- Rafferty NE, CaraDonna PJ, Bronstein JL. 2015. Phenological shifts and the fate of mutualisms. *Oikos*, 124:
   14-21.
- Reyer CPO, Leuzinger S, Rammig A, Wolf A, Bartholomeus RP, Bonfante A, de Lorenzi F, Dury M, Gloning
   P, Abou Jaoude R, Klein T, Kuster TM, Martins M, Niedrist G, Riccardi M, Wohlfahrt G, de Angelis
   P, de Dato G, Francois L, Menzel A, Pereira M. 2013. A plant's perspective of extremes: terrestrial
   plant responses to changing climatic variability. *Global Change Biology*, **19**: 75-89.
- Ruiz-Labourdette D, Schmitz MF, Pineda FD. 2013. Changes in tree species composition in Mediterranean
   mountains under climate change: Indicators for conservation planning. *Ecological Indicators*, 24:
   310-323.
- Sanz-Elorza M, Dana ED, Gonzalez A, Sobrino E. 2003. Changes in the high-mountain vegetation of the
   central Iberian peninsula as a probable sign of global warming. *Annals of Botany*, 92: 273-280.
- Sarris D, Christodoulakis D, Körner C. 2011. Impact of recent climatic change on growth of low elevation
   eastern Mediterranean forest trees. *Climatic Change*, 106: 203-223.
- 142Schoenfeld S. 2010. Environment and Human Security in the Eastern Mediterranean: Regional143Environmentalism in the Reframing of Palestinian-Israeli-Jordanian Relations. Achieving144Environmental Security: Ecosystem Services and Human Welfare, 69: 113.
- Simeone MC, Piredda R, Papini A, Vessella F, Schirone B. 2013. Application of plastid and nuclear markers
   to DNA barcoding of Euro-Mediterranean oaks (Quercus, Fagaceae): problems, prospects and
   phylogenetic implications. *Botanical Journal of the Linnean Society*, 172: 478-499.
- Thomas CD. 2011. Translocation of species, climate change, and the end of trying to recreate past
   ecological communities. *Trends in Ecology & Evolution*, 26: 216-221.
- Thuiller W, Lavergne S, Roquet C, Boulangeat I, Lafourcade B, Araujo MB. 2011. Consequences of climate
   change on the tree of life in Europe. *Nature*, 470: 531-534.

- Thuiller W, Lavorel S, Araújo MB, Sykes MT, Prentice IC. 2005. Climate change threats to plant diversity in
   Europe. Proceedings of the National Academy of Sciences of the United States of America, 102:
   8245-8250.
- **Tomaselli R. 1977**. The degradation of the Mediterranean maquis. *Ambio*: 356-362.
- Toumi L, Lumaret R. 2010. Genetic variation and evolutionary history of holly oak: a circum-Mediterranean
   species-complex [Quercus coccifera L./Q. calliprinos (Webb) Holmboe, Fagaceae]. *Plant Systematics and Evolution*, 290: 159-171.
- Trabaud L. 1991. Fire Regimes and Phytomass Growth Dynamics in a Quercus-Coccifera Garrigue. *Journal of Vegetation Science*, 2: 307-314.
- Tutin TG, Heywood VH, Burges NA, Valentine DH, Walters SM, Webb DA. 2001. Flora Europaea.
   Cambridge, Cambridge University Press.
- 163 Underwood EC, Viers JH, Klausmeyer KR, Cox RL, Shaw MR. 2009. Threats and biodiversity in the
   164 mediterranean biome. *Diversity and Distributions*, 15: 188-197.
- 165 UNDP. 2011. Lebanon's Second National Communication to the United Nations Framework Convention on
   166 Climate Change. Beirut, Lebanese Ministry of the Environment.
- van Vuuren DP, Edmonds J, Kainuma M, Riahi K, Thomson A, Hibbard K, Hurtt GC, Kram T, Krey V,
   Lamarque JF, Masui T, Meinshausen M, Nakicenovic N, Smith SJ, Rose SK. 2011. The
   representative concentration pathways: an overview. *Climatic Change*, 109: 5-31.
- Vessella F, Schirone B. 2013. Predicting potential distribution of Quercus suber in Italy based on ecological
   niche models: Conservation insights and reforestation involvements. *Forest Ecology and* Management, 304: 150-161.
- Vessella F, Simeone MC, Schirone B. 2015. Quercus suber range dynamics by ecological niche modelling:
   from the Last Interglacial to present time. *Quaternary Science Reviews*, 119: 85-93.
- 175 **Whyte RO. 1950**. The Phytogeographical Zones of Palestine. *Geographical Review*, **40**: 600-614.
- Xu J, Grumbine RE, Shrestha A, Eriksson M, Yang X, Wang Y, Wilkes A. 2009. The melting Himalayas:
   cascading effects of climate change on water, biodiversity, and livelihoods. *Conservation Biology*,
   23: 520-530.
- **Zohary M. 1960**. The maquis of Quercus calliprinos in Israel and Jordan. *Bulletin of the Research Council of Israel,* **90**: 51-72.
- 181 182