

Abstract

 Although knowledge of Arctic benthic biodiversity has increased considerably in recent decades, some regions, such as Northeast Greenland, remain poorly studied. The aim of this study was to complement a previous macrofaunal inventory carried out in Young Sound, a High-Arctic fjord in this region (74°N). We sampled shallow benthic assemblages along a small inner/outer fjord gradient, including one station previously prospected two decades ago and three new stations. This sampling strategy revealed highly diversified benthic assemblages (166 species identified on a total sampling area of 1.32 m²), which considerably increases the number of species recorded for the fjord (i.e. 225 species vs 100 previously recorded). The outermost station was dominated in abundance by various assemblages of bivalves, while the middle stations showed greater species evenness, including numerous species of polychaetes, bivalves and crustaceans. The innermost station was dominated by ostracods, gammarid amphipods and tube-dwelling polychaetes. Overall, benthic assemblages varied little between the four stations and the transect as a whole exhibited characteristics typical of outer fjord habitats, reflecting the rather moderate impact of meltwater inputs in this part of the fjord. Finally, trophic plasticity and omnivory were observed in most of the recorded macrobenthic species, highlighting the adaptability of these species to low trophic availability in the ecosystem. Future biodiversity studies will need to explore the innermost and deepest areas to provide a more comprehensive inventory and understanding of the influence of environmental conditions on the structure and functioning of Young Sound benthic habitats.

Key words

Benthic ecosystems • Macrofauna • Community structure • Spatial variability • Meltwater

inputs • Arctic fjord

Introduction

 To date, Arctic benthic biodiversity remains poorly studied although species colonizing the seafloor represent around 90 % of the total estimated species richness of Arctic invertebrates (Sirenko 2001; Gradinger et al. 2010). While Arctic benthic ecosystems have long been considered as areas of low species diversity (e.g. Knox and Lowry 1977), there is increasing evidence that this assumption has likely been biased by the scarcity of surveys conducted at high latitudes (Piepenburg 2005; Piepenburg et al. 2011). Recent inventories estimate the diversity of Arctic benthic species to be over 4600 species, a much higher number than previously estimated, while several thousand species are likely to remain to be discovered (Bluhm et al. 2011; Josefson et al. 2013). Although considerable progress has been made in Arctic biodiversity in recent decades, some geographic areas remain understudied such as East Greenland (Piepenburg et al. 2011).

 Under climate change, Arctic benthic ecosystems are expected to undergo major modifications driven by changes in sea ice cover, salinity and temperature, acidification, or sedimentation (AMAP 2017). It is very likely that these changes will impact benthic biodiversity, especially in coastal areas, which are expected to be the most exposed to future environmental changes (Węsławski et al. 2011). The expansion of boreal species into high latitudes is also expected to affect benthic biodiversity by bringing new species into Arctic areas (Węsławski et al. 2011; Kotwicki et al. 2021; Cottier-Cook et al. 2023), with likely consequences for species interactions and ecosystem functioning. Indeed, the recent expansion of several engineer species with boreal affinities into the Arctic (e.g. mussel species from the *Mytilus* spp. complex or the kelp *Laminaria hyperborea*) has had implications for the structure of shallow benthic ecosystems by increasing habitat heterogeneity and complexity, facilitating the arrival of species new to these ecosystems (Berge et al. 2005; Krause-Jensen and Duarte 2014; Mathiesen et al. 2017; Filbee-Dexter et al. 2019; Kotwicki et al. 2021). Many questions remain unresolved regarding how endemic benthic species will cope with new abiotic conditions (increases in temperature or sedimentation, decreases in seawater salinity) and associated changes on biotic interactions (e.g. the arrival of new predators or competitors, Renaud et al. 2015). Assessing the vulnerability of Arctic benthic communities experiencing such environmental and ecological changes thus remains challenging and highlights the strong need for initial references states in benthic biodiversity in all Arctic ecoregions.

 In this context, any increase in knowledge of the benthic biodiversity of poorly characterized Arctic ecoregions is crucial, especially in shallow coastal ecosystems. For example, the only studies conducted on the North-Eastern coasts of Greenland have focused on epifauna and/or megafauna shelf communities (e.g. Piepenburg and Schmid 1996; Brandt and Schnack 1999; Starmans et al. 1999; Fredriksen et al. 2020). To our knowledge, for such real *Terra Incognita* of Arctic ecoregions (Piepenburg et al. 2011), only Sejr et al. (2000) provided quantitative datasets on shallow macrozoobenthos assemblages in a high-Arctic fjord (Young Sound, NE Greenland). Few studies have also investigated trophic linkages in macrobenthic ecosystems from the Northeast Water Polynya providing only limited details on their biodiversity (Hobson et al. 1995; Ambrose and Renaud 1997; Piepenburg et al. 1997). Finally, although many older 91 studies from the mid-20th century have been conducted on the shallow habitats (e.g. Ockelmann 1958), they have mostly focused on the occurrence of individual taxa than providing insights into the structure and diversity of macrobenthic communities.

 The aim of the present study is to complement the inventory of benthic macrofaunal species carried out by Sejr et al. (2000) in Northeast Greenland (Young Sound, 74°N) with additional sampling sites. While Sejr et al. (2000) examined the macrobenthic communities in vertical 97 zonation (from 20 to 85 m), this study prospects shallow benthic macrofaunal assemblages (\approx 20 m) distributed along a small inner/outer gradient located in the most marine part of the fjord. We resampled the shallowest Daneborg station from Sejr et al. (2000) and investigated three new stations, one towards the outer fjord (Kap Breusing) and two towards the inner fjord (Basalt Island, Pass Hytten). The description of the assemblages, coupled with those of Sejr et al. (2000), aims to provide a baseline for future studies investigating the temporal variability of Arctic shallow coastal soft-bottom communities and their response to ongoing climate change.

Materials and Methods

Study sites

 Young Sound is a high Arctic fjord located in North-East Greenland (Fig. 1) of about 90 km 107 long and 2-7 km wide (Rysgaard et al. 2003). A sill is located at its outer section (\approx 45 m depth, see Fig. 1) limiting seawater exchanges between the inner parts of the fjord and the shelf (Bendtsen et al. 2007) where average and maximum depths reach 100 and 360 m, respectively (Rysgaard et al. 2003). This fjord is characterized by extreme seasonal variations, thick sea ice $(> 150 \text{ cm})$ and snow cover $(20 - 100 \text{ cm})$ covering the fjord for 9 to 10 months per year (Glud

 et al. 2007) and strongly limiting the annual primary production, estimated to be about 10 g C m⁻² yr⁻¹ (Rysgaard et al. 1999). During the summer, glacier and snow melt generate strong river 114 flows ranging annually from 0.9 to 1.4 km^3 (Bendtsen et al. 2014). Such freshwater inputs induce strong spatial salinity and temperature gradients in summer surface waters, with temperature decreasing from 9 to 2°C and salinity increasing from 8 to 30, from the inner to the outer part of the fjord, respectively (Bendtsen et al. 2007). In contrast, deeper waters (> 20 m) 118 generally remain stable throughout the year (i.e. $> 2^{\circ}$ C) along the entire fjord length (Holding et al. 2019). Vertical variations in salinity induce a strong permanent haline stratification at 5 to 10 m depth, which limits pelagic primary production by reducing nutrient exchange between surface and subsurface waters (Holding et al. 2019; Sejr et al. 2022). On a decadal scale, the sea ice conditions in the fjord have changed with a gradual increase in the duration of the ice- free season from about 80 days per year (1950-2000) to more than 90 days per year but combined with a significant increase in the year-to-year variability since 2000. At the same time, freshwater runoff (mostly glacial meltwater) is estimated to have increased 6-7 fold since the 1960s (Sejr et al. 2022). Annual monitoring of main Young Sound environmental parameters (physico-chemical, pelagic productivity and community structure) has been made for three decades by the Greenland Ecosystem Monitoring – Marine Basis program, making Young Sound ideal for studying the effects and feedback of climate change in the high Arctic. However, the macrozoobenthic benthic communities are not part of the monitoring program since the study by Sejr et al. (2000), highlighting the crucial need to reacquire biodiversity data.

Sampling and laboratory analyses

 Sampling was conducted in August 2016 at four stations distributed along an inner/outer marine fjord gradient: Pass Hytten, Basalt Island, Daneborg and Kap Breusing (Fig. 1, Table 1). Macrofauna samples were collected by scuba divers for each replicate using a modified Hiscock and Hoare suction dredge (Eleftheriou and McIntyre 2005), fitted with a suction pipe directly connected to a nylon mesh sample bag (square mesh = 1 mm). Each replicate was collected over an area of 0.09 m² (except for one replicate at Kap Breusing, which was collected over 0.0625 m²) and then fixed directly in 10% buffered formalin. Sediment samples were also collected at both Pass Hytten and Kap Breusing using small sediment cores for grain size analysis. Due to logistical problems, we were unable to collect sediment samples at either Daneborg or Basalt Island. Once in the laboratory, the animals were sorted, transferred to 70% ethanol and then identified to the lowest taxonomic level. Sediment samples were analyzed by the private company "*Eurofins – Analyses pour l'environnement*" (Saverne, France) using laser methods: samples were sieved through a 2 mm square mesh and then analyzed using a Malvern Mastersizer 2000 laser particle size analyzer to provide different size fractions ranging from 2 147 to 2000 µm.

Statistical analyses

 Spatial differences in benthic assemblage structure were visualized using Principal Component Analysis (PCA) performed on Hellinger-transformed fauna abundances. This transformation was selected to reduce the weight of rare species in the analysis (Legendre and Gallagher 2001). The first twenty species explaining the most spatial variation in the ordination (i.e. accounting for more than 60% of SIMPER dissimilarities) were plotted in a second graph to visualize which species contributed to the benthic structure of each station. Spatial differences in community structure were tested using one-way permutational multivariate analysis of variance (PERMANOVA) performed on Bray-Curtis transformed fauna abundances. Homoscedasticity was checked using a permutational analysis of multivariate dispersions (PERMDISP) test to ensure that data dispersion did not influence the PERMANOVA result (Anderson et al. 2008; Anderson and Walsh 2013). Species rarefaction curves were calculated for each replicate to compare trends in species richness between stations. Species richness was compared among replicates based on rarefaction curves by standardizing all data to the same abundance (corresponding to the number of individuals in the replicate with the lowest abundance) to remove any abundance bias on taxa richness (Gotelli and Colwell 2001). An additional rarefaction curve was plotted to estimate the specific richness for the entire Young Sound system and Hulbert's diversity index was calculated to estimate the expected number of species within a random sample of 35 individuals (based on the lowest number of individuals observed in a replicate). Jacknife1 Chao2 estimates were calculated and plotted as accumulation curves to estimate the total species richness in case species rarefaction curves did not reach their asymptotes (Colwell et al. 2004; Chao and Chiu 2016). Finally, we calculated the Pielou's evenness index to take into account the diversity observed at each station along with the associated species richness using the "vegan" R package (Oksanen et al. 2019). All other diversity indices were calculated using EstimateS v9.1 software.

- **Results**
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Structure of benthic community

 Significant differences in benthic community structure were observed between sampling stations (degrees of freedom for treatment and error [*df*] = 3 and 11, pseudo-*F* = 8.2529, *p*-perm $177 = 0.001$). Pairwise analyses revealed that such differences were observed between all stations (the smallest significant difference was observed between Basalt Island and Kap Breusing, *df* $= 1$ and 4, pseudo- $F = 7.4856$, *p*-perm $= 0.049$). PCA analysis showed that more than half of the spatial variability was explained by the first three principal components (PC1, PC2 and PC3 explained 24.36, 17.92 and 17.27 % of the variance, respectively). Low dispersion was observed among the replicates of each station (Fig. 2a, c). Significant differences in Pielou's evenness index were observed between stations (*df* = 3 and 14, pseudo-*F* = 36.4600, *p*-perm = 0.001). These differences were observed between all stations by pairwise PERMANOVA analyses (the smallest significant difference was observed between Daneborg and Kap 186 Breusing, $df = 1$ and 4, pseudo- $F = 16.2496$, *p*-perm = 0.048), except between Pass Hytten 187 and Kap Breusing $(df = 1$ and 4, pseudo- $F = 0.0386$, *p*-perm = 1).

 Benthic assemblages at Pass Hytten were strongly dominated by crustaceans (relative abundance = 52.9 - 78.0%, Fig. 3a, b), mainly due to high densities of ostracods and the amphipod *Metopa glacialis* (Fig. 2b). The community structure at Daneborg and Basalt Island was more evenly balanced between taxa (Fig. 3a, b), although polychaetes were the most dominant group (relative abundance = 32.6 - 58.3% at Daneborg and 44.9 - 55.9% at Basalt Island). *Euchone analis*, *E*. *incolor* and *Laphania boecki* were the most abundant polychaete species at Daneborg, while *Abyssoninoe* sp., *Laphania boecki* and *Nereimyra aphroditoides* dominated the polychaete assemblage at Basalt Island (Fig. 2b, d). Finally, bivalves were the dominant taxonomic group at Kap Breusing (relative abundance = 67.5 - 68.5%, Fig. 3a, b), largely represented by high densities of the filter-feeding bivalves *Hiatella arctica* and *Musculus discors* (> 50%, Fig. 2b, d).

 Total abundance showed large variation among stations and replicates (Fig. 3b). The highest 200 abundances were observed in the second replicate from Kap Breusing (reaching 6044 ind. m^2 , Fig. 2b). In contrast, the first replicate from Kap Breusing showed a much lower total abundance $(1824 \text{ ind. m}^{-2})$, which was in the same order of magnitude as the abundances observed at 203 Daneborg (from 933 to 1733 ind. m⁻²) and Pass Hytten (from 1467 to 2667 ind. m⁻², Fig. 2b).

204 Finally, the lowest total abundances were observed at Basalt Island (from 389 to 1611 ind. $m⁻²$, Fig. 2b).

Abiotic and biotic description of benthic habitats

Pass Hytten

 The sediment texture is dominated by fine silt particles (55% of the total size fraction for 209 particles $<$ 63 µm) mixed with fine sands (25%) and gravels (20%, Fig. 1 in Online Resource 210 1). Significant bioturbation activity probably occurs due to high abundances of epifauna (e.g. *Ophiocten sericeum*, Fig. 4a, Table 2) and burrowing shrimps (e.g. *Sclerocrangon boreas*, L. Chauvaud pers. obs.). The associated macrozoobenthic assemblage shows a low species evenness with ten taxa representing 83.5% of the total abundance (Table 2), including crustaceans such as ostracods, the amphipod *Metopa glacialis* and the cumacean *Diastylis scorpioides*(59 % of the total abundance). The strictly benthic species (epifauna and endofauna) were dominated by both tube-dwelling (*Euchone incolor* [see high abundances in Fig. 3a], *Lyssipe labiata*, *Maldane sarsi* and *Owenia borealis*) and errant polychaetes (*Abyssoninoe* sp. and *Leitoscoloplos mammosus*). Bivalves *Astarte moerchi*, *Hiatella arctica*, *Mya truncata* and *Musculus discors*) also occur but in much lower proportions (< 15 % of the total abundance).

Basalt Island

 Here, the sediment is a mixture of stone, sand and small gravel (2-5 mm) associated with detrital organic matter (mainly macroalgae and shell debris) accumulated on the seafloor (Fig. 4b). However, we were unable to sample the sediment at this station due to the dominance of stones. 224 The steep slope of the seafloor (slope gradient $\approx 80 - 100$ %, see Fig. 4b) probably makes the bottom unstable due to the downward export of mineral material towards the deep fjord basins. The species assemblage at Basalt Island shows greater species evenness than Pass Hytten and Kap Breusing, with the ten most dominant species accounting for only 51.3% of total abundance (Table 2). Polychaetes dominate the species assemblage with *Abyssoninoe* sp., *Clymenura* sp., *Laphania boecki*, *Nereimyra aphroditoides* and *Pholoe* sp. representing 26.4 % of the total abundance. Several bivalves such as *A*. *moerchi*, *H*. *arctica* or *M*. *truncata* were also found at the site but in low abundances, as well as the brittle star *Ophiocten sericeum* (Table 2, Table S1 in Online Resource 1). Among crustaceans, ostracods show a high abundance (5.9 %, Table 2).

Daneborg

Based on underwater pictures and previous data on particle size distribution (Sejr et al. 2000),

235 the seabed was characterized by a mixture of fine and very fine sand (particles $[63 - 125 \text{ µm}] =$

- 236 24.6 % and particles $> 125 \mu m = 44.2 \text{ %}$). The species evenness of the species assemblage was
- close to that observed at Basalt Island (Table 2). Daneborg is mainly dominated by two taxa,
- bivalves and in particular the species *Astarte moerchi* (18.6 %) and crustaceans (ostracods and
- *D*. *scorpioides* representing 11.8 % and 3.1 % of the total abundance, respectively). Polychaetes
- complete the assemblage, including *Euchone incolor*, *Laphania boecki*, *Euchone analis*, *Chone*
- sp. and *Clymenura* sp., which all account for 23.6% of the total abundance.
- *Kap Breusing*
- The grain size distribution of Kap Breusing was variable between replicates, but overall it was 244 a poorly sorted sandy sediment (the 63-2000 µm grain size range varied from 53 to 80 %, Fig. S1 in Online Resource 1). Heterogeneous medium/coarse sand (38 to 55 % of the particles 246 between 250 and 2000 µm) was found in two of the three replicates, the other was largely 247 composed of fine sand (61 % of particles between 63 and 250 μ m).
- The species assemblage here was associated with a rock wall close to the location where the sediment samples were collected and included Sertulariidae hydrozoans and *Musculus discors* bivalves (L. Chauvaud, pers. obs., see background of Fig. 4d). Both species form a complex three-dimensional structure hosting a highly diversified assemblage including the gastropod *Alvania scrobiculata* and the bivalves *A*. *moerchi*, *H*. *arctica*, *M*. *discors* and *M*. *truncata*, representing more than 68% of the total abundance (Table 2). Four vagile polychaete species were also found as *Harmothoe* sp., *Nereimyra aphroditoides*, *Nereis zonata* and *Syllis* sp. (11.3 % of total abundance) and the sedentary terebellid *Polycirrus medusa* (2.8 %; Table 2).
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Biodiversity estimates of benthic habitats

257 Based on a total area of 1.32 m^2 sampled by 15 suction dredge replicates, we collected a total of 2322 individuals belonging to 166 species (Table S1 in Online Resource 1). The polychaete phylum was by far the most diverse (67 species), followed by mollusks (33 species) and crustaceans (31 species). Echinoderms, nemerteans, sipunculans, pycnogonids, tunicates, nematodes, brachiopods and platyhelminths were also found but always displayed with much lower species richness. Macroalgae randomly collected in our samples were also identified, and the species list is included in the supplementary material as complementary observations (Table S2 in Online Resource 1).

 Species accumulation curves did not reach their asymptotes for all stations (Fig 5). The initial slope of the species accumulation curves was highly variable between stations, but more consistent between intra-site replicates. Based on the lowest number of individuals (i.e. 35 in 268 the 5th Basalt Island replicate), Hulbert's expected species richness within a random sample of 35 individuals (ES35) displayed the lowest levels in Pass Hytten and Kap Breusing (i.e. range ES₃₅ = 9 – 13 and 12 – 13 for Pass Hytten and Kap Breusing, respectively). Expected species 271 richness showed intermediate values in Daneborg (range $ES_{35} = 15 - 21$) and the highest levels 272 in Basalt Island (range $ES_{35} = 19 - 22$). Based on the total number of species per replicate (S_{tot}), 273 species richness was also lower at Pass Hytten (range $S_{\text{tot}} = 17 - 31$) than at Basalt Island (range S_{tot} = 19 – 43) and Daneborg (range S_{tot} = 32 – 47), while Kap Breusing had the highest intra-275 station variability (range $S_{tot} = 27 - 59$). A rarefaction curve based on the total set of fauna samples (i.e. all replicates from the four sampled stations combined) also does not reach an asymptote either, despite the much higher number of individuals (i.e. > 2000).

 The rarefaction curves of Chao2 and Jackkniffe1 estimates slightly differ from the species accumulation curves (Fig. 6a, b). Based on the lowest number of replicates (i.e. 2 sampled stations at Kap Breusing), biodiversity estimates were lowest at Pass Hytten (Chao2 = 48, 281 Jackknife1 = 51), intermediate at Basalt Island (Chao2 = 62, Jackknife1 = 64) and Daneborg 282 (Chao2 = 77, Jackknife1 = 78), and highest at Kap Breusing (Chao2 = 95, Jackknife1 = 91). Based on the entire dataset, Chao2 and Jackknife1 estimates predicted a total species richness of 204 and 221 species, respectively.

Discussion

Benthic species richness of the study area: historical and recent data comparison

 The present study complements the previous work of Sejr et al. (2000) conducted twenty years ago in Young Sound, which described gradual, spatially limited changes in species assemblages along a depth gradient (from 20 to 85 m) in the vicinity of Daneborg station (Fig. 1). Our study conducted along a small inner/outer fjord gradient in the most marine part of Young Sound shows little spatial variation in shallow (above 20 m depth) benthic community structure. This pattern suggests that our four sampling stations cover a relatively continuous and uniform habitat. The sampled habitat is mostly dominated by large tubicolous polychaetes (e.g. Maldanidae, *Euchone*) and some specific motile polychaetes (*Leitoscoloplos*, *Lumbreneris*), and has a high abundance of suspension-feeding bivalves (e.g. *Astarte*, *Bathyarca*, *Hiatella*). Turbidity and meltwater inputs are generally factors that control community structure in Arctic fjords, decreasing the presence of filter feeders towards the inner areas of fjords exposed to glacial melt (Włodarska-Kowalczuk and Pearson 2004; Sejr et al. 2010; Włodarska-Kowalczuk et al. 2012; Udalov et al. 2021). However, we doubt that turbidity and meltwater inputs account for the spatial variations observed between our four stations, as weak environmental forcing is observed along the small inner/outer fjord gradient studied in our survey. In fact, although turbidity reaches more than 50 NTU in the innermost parts of Young Sound, it only varies from 2-3 to less than 1 NTU between Pass Hytten and Kap Breusing (Meire et al. 2017). It is therefore more likely that the observed spatial variability in community structure reflects small-scale habitat variation (bottom varying abundance in bare rock, fine sand, mixed sand and gravel or mud) rather than large-scale environmental variation (i.e. differences in exposure to meltwater inputs).

 The present study provides a complementary census of 166 benthic species, including numerically dominant polychaetes (67 species), crustaceans (46 species) and mollusks (33 species). It complements the previous work by Sejr et al. (2000) which listed a total of 100 311 species despite a higher sampling effort (total sampled surface $= 2.24$ m² vs 1.3225 m² here). With only 41 species in common between both studies, the total number of benthic invertebrate species recorded from the Young Sound now reaches a total of 225. The relatively low number of taxa shared between the two inventories mainly reflects the different sampling strategies as Sejr et al. (2000) targeted species assemblages from increasing depths while our sampling focused on a limited spatial fjord gradient that covers contrasting shallow habitats (i.e. silt, sand/gravel, rocky wall) and their associated species assemblages.

 It should also be noted that identification bias cannot be excluded as a single species may be recorded under two different names over time, as different identification experts were involved and taxonomical descriptions have been produced since 2000. As an example, *Scoloplos armiger* (only in Sejr et al. 2000) and *Leitoscolopos mammosus* (only here) have been reported to taxonomic confusion (Mackie 1987; Renaud et al. 2007). Further census efforts will include harmonization of the species list with the involvement of benthic taxonomy experts. It could be assumed that the relatively low number of species common to this inventory and the study by Sejr et al. (2000) is due to long-term variability in benthic community structure. However, this seems unlikely as previous studies conducted in Arctic silled-fjord show a high stability of benthic community structure over several decades (Renaud et al. 2007), in contrast to open fjords (Beuchel et al. 2006; Kortsch et al. 2012).

 As indicated by the slopes of the species accumulation curves, the inventory of benthic biodiversity in this fjord is far from being complete, with many rare species likely yet to be 331 recorded at all study sites. The expected number of species per 35 individuals (ES_{35}) varies between 9 and 22 species, underlining the high biodiversity but also the heterogeneity across stations and replicates. Based on biodiversity estimates, total species richness of the outer fjord shallow sites should reach around 204-221 species, much higher than previous estimates based on Sejr et al. (2000) and this should be much higher as only a very limited part of the shallow Young Sound has been visited and a wide variety of habitats remains unexplored. Inventory studies on shallow and deep areas of both outer and inner fjord usually find much higher species diversity than in the present work. For example, Kędra et al. (2010) recorded more than 300 taxa after extensive multi-year sampling of all soft-bottom habitats in Hornsund (SW Svalbard, from 30 m to 250 m), and Voronkov et al. (2013) found 403 benthic taxa in shallow hard-bottom communities from Kongsfjorden (NW Svalbard). We thus think that the actual benthic species inventory of the Young Sound is far from complete, as it is limited to the only most marine section.

Ecological characteristics and insights about the functioning of the benthic fauna

Ecological characteristics of benthic assemblages

 The current inventory identifies four major benthic assemblages, all belonging to the Young Sound outer fjord benthic community: 1) the rocky outer fjord assemblage (Kap Breusing), 2) the sandy outer fjord assemblage (Daneborg), 3) the sand-gravelly outer fjord assemblage (Basalt Island) and 4) the silty outer fjord assemblage (Pass Hytten).

 The macrobenthic assemblage collected at Kap Breusing was composed of a unique and poorly described species assemblage dominated by two engineer species, the bivalve *Musculus discors* and the sertulariid colonial hydroids. Several studies have highlighted the ability of some bivalve species to secrete byssus threads to construct nests that host a high diversity of other benthic invertebrate species (e.g. *Arcuatula senhousia*, *Limaria hians*; Morton 1974; Hall- Spencer and Moore 2000). Byssus threads produced by these species transform the physical habitat into a complex three-dimensional structure that is then colonized by hydroids (Merrill and Turner 1963). To our knowledge, only one study has reported this ability for *M*. *discors* (Merrill and Turner 1963). The high biodiversity observed at Kap Breusing (where the highest number of species was observed within a single replicate) reflects the structuring effect of the *M*. *discors* / Sertulariidae assemblage on the community structure. The high abundance and diversity of motile polychaetes in this assemblage, especially for Nereididae, Polynoidae or Syllidae, is similar to assemblages observed in complex three-dimensional habitats such as maerl beds, kelp forests or artificial structures (Grall et al. 2006; Yakovis et al. 2007; Pabis and Sicinski 2010). The *M*. *discors* / Sertulariidae assemblage was also dominated by large suspension-feeding bivalves (*Hiatella arctica*, *Musculus discors* and *Mya truncata*), which is more in line with common species assemblages observed on seafloors protected from sediment disturbance caused either by freshwater inputs or by burrowing species (Włodarska-Kowalczuk 2007).

 The outer fjord sandy assemblage of Daneborg was dominated by the bivalve *Astarte moerchi*, ostracods and polychaetes of the genus *Euchone*. The dominance of *A*. *moerchi* among the bivalves collected at Daneborg is similar to that observed by Sejr et al. (2000) at the same station but community structure differs with our data: several abundant polychaete species found in 2016 (e.g. *Clymenura* sp., *Euchone* spp., *Laphania boecki*) were not observed by Sejr et al. (2000) while other species abundant in 1995 (e.g. *Dipolydora quadrilobate* and *Lumbriclymene minor*) were not collected in 2016. Such long-term differences could be explained either by long-term changes in the benthic assemblages but also to our weak sampling effort that is submitted to small scale habitat heterogeneity.

 The Basalt Island sand-gravel outer fjord assemblage is dominated by an association of motile polychaetes (*Abyssoninoe* sp., *Clymenura* sp., *L*. *boecki*, *Nereimyra aphroditoides*, *Pholoe* sp.), ostracods and the bivalve *A*. *moerchi*. Although the associated community structure differs with that of Daneborg, both display several common dominant species (e.g. *A*. *morchi*, *Clymenura* sp., *L*. *boecki*) indicating common ecological characteristics potentially related to their vicinity in the fjord and similar environmental constraints.

 Finally, the silty outer fjord assemblage collected at Pass Hytten is dominated by ostracods, the *Metopa glacialis* amphipod as well as tube-dwelling polychaete species (e.g. *Maldane sarsi*, *Owenia borealis*, *Euchone* sp.) and crustaceans. Surprisingly, several species described in the literature as typical of either outer or inner fjord assemblages are found in association in Pass Hytten. For instance, the sub-surface deposit feeder *M. sarsi* is the most dominant polychaete although it is reported to be typical of outer fjord assemblages (Gulliksen et al. 1985; Renaud et al. 2007; Kędra et al. 2013), whereas it is absent from inner fjord areas (Włodarska- Kowalczuk et al. 2012). Surface deposit-feeding thyasirids and nuculanids are also present, albeit in moderate abundance, at Pass Hytten, although they are generally representative of benthic assemblages from inner fjord areas (Włodarska-Kowalczuk and Pearson 2004). The presence of *Euchone* sp. or *Leitoscoloplos mammosus* polychaetes in Pass Hytten is also consistent with previous observations made in shallow and deep areas exposed to high sedimentation rates (Sejr et al. 2010; Kędra et al. 2013; McGovern et al. 2020). Therefore, the coexistence of several species typical of both outer and middle/inner fjord assemblages at Pass Hytten suggests that although this site is more similar to outer fjord habitats, the presence of a handful of species adapted to high sediment loads would indicate that this station is approaching the transition zone between the outer and inner parts of the fjord. This could be explained by the more interior location of Pass Hytten compared to the other three stations sampled, as well as its location closer to the Zackenberg River delta. We hypothesize that future sampling, either in the innermost part of the fjord or closer to the Zackenberg delta, will identify new species assemblages typical of benthic habitats from areas exposed to glacial and freshwater inputs.

New insights about the functioning of the shallow benthic outer fjord community

 The new data from our study provide a better understanding of the importance of trophic adaptations for benthic community functioning in outer fjord habitats. A complete analysis of the trophic structure of benthic assemblages has already been conducted at Pass Hytten and Kap Breusing by Bridier et al. (2021). It appears that several dominant species in our study are known to have high trophic plasticity (i.e. feeding on several carbon sources, De Cesare et al. 2017; Bridier et al. 2021). The bivalves *Astarte moerchi*, *Astarte elliptica*, *Hiatella arctica*, and *Musculus discors* in Young Sound appear to feed on a range of carbon sources such as phytoplankton, benthic organic matter (i.e. macroalgae and/or microphytobenthos) or sedimentary organic matter (detritus, Bridier et al. 2021). For some species (e.g. the astartidae *A*. *moerchi* and *A*. *elliptica*), the proportion of alternative organic matter resources can even reach almost 50% of their diet (Bridier et al. 2021). In addition, other species abundant in Young Sound are known to be omnivores (i.e. feeding on several trophic levels, McMeans et al. 2015; Bridier et al. 2021). This is particularly true for the brittle star *Ophiocten sericeum*, the amphipod *Metopa glacialis* and the polychaete *Nereis zonata*, which were observed in high abundance at all stations. Although the brittle star *O*. *sericeum* is generally considered a surface feeder, it falls between the primary and the secondary consumer levels in Young Sound (Bridier et al. 2021). Conversely, the polychaete *N*. *zonata* and the amphipod *M*. *glacialis* exhibit trophic levels below those of the strict predators with which they are generally associated (Bridier et al. 2021). These variations in trophic position show that these species are capable of feeding at multiple trophic levels, feeding on sources of primary organic matter when these are available in the ecosystem, or feeding at higher trophic levels when the availability of primary production is low (McMeans et al. 2015; Bridier et al. 2021). All these trophic adaptations likely reflect the low pelagic productivity of Young Sound, which ranks among the least productive coastal 429 ecosystems in the world, with pelagic primary production estimated at 10 g C m⁻² yr⁻¹ (Rysgaard et al. 1999; Cloern et al. 2014). Trophic plasticity might therefore allow species to adapt to this low pelagic primary production by optimizing carbon flows through food webs through the exploitation of all carbon sources available in the ecosystem (e.g. microphytobenthos, macroalgae, detritus, Mincks et al. 2008; Renaud et al. 2015). Omnivory may also help benthic species to adapt to low pelagic primary production by reusing carbon that has already passed through the food web a second time, thus participating in the recycling of carbon circulating in the community.

 The present study complements previous work by Bridier et al. (2021) on the functioning of benthic communities in the shallow habitats of Young Sound. Although Bridier et al. (2021) highlighted the existence of trophic adaptations for several benthic species in Young Sound, the paucity of information on benthic community structure prevents drawing conclusions on the functioning of benthic communities in outer fjord habitats. The new data presented here show that the observations on community trophic adaptations involve species that are numerically dominant in the inventoried benthic habitats (i.e. the rocky, sandy and silty outer fjord assemblages). For example, the species *H*. *arctica*, *M*. *discors* (characterized by high trophic plasticity) and *N*. *zonata* (characterized by high levels of omnivory) alone account for 66.5% of the total abundance at Kap Breusing. It could be argued that biomass data would have been more appropriate than abundance data to discuss the role of these latter species in community carbon fluxes. However, considering that all species mentioned as omnivorous or trophically plastic are larger than the rest of the species identified in this study, the biomass dominance of these individuals is likely greater than their numerical dominance, suggesting then an even greater role for community carbon fluxes. Trophic adaptations thus play a predominant role in the functioning of the outer fjord benthic communities, which feed only partially on fresh organic matter from pelagic primary production. These results also highlight the crucial role of benthic primary producers in the functioning of shallow Arctic coastal ecosystems, whose importance is often underestimated, although they represent a significant carbon source (Attard et al. 2016; Gaillard et al. 2017; Ørberg et al. 2022). With climate change, increased meltwater inputs will have multiple consequences for Young Sound's pelagic environments, including impacts on pelagic primary production, organic matter quality and organic matter transfers to zooplankton (Meire et al. 2017; Middelbo et al. 2018; Bridier et al. 2019). The existence of trophic adaptations in several species that dominate the benthic communities of the outer fjord (and therefore potentially represent major carbon transfer pathways) could thus be a crucial advantage for their resilience in the face of upcoming environmental changes. In the future, it would be important to investigate the community and trophic structures of the inner fjord to observe how the dominance of these trophic adaptations evolves in habitats exposed to stronger meltwater inputs and higher trophic constraints. In addition, it would also be interesting to describe the structure of benthic communities and food webs in the deeper basins of the fjord (up to 350 m), which do not benefit from local benthic primary production.

Conclusion and further directions of research

 The species list obtained in this study is a substantial addition to the inventory carried out two decades ago, adding 66 new species to the 100 already listed by Sejr et al. (2000). The sampling of three new stations (Pass Hytten, Basalt Island, Kap Breusing) in addition to the station previously described by Sejr et al. (2000) (i.e. Daneborg) also provides new insights into the spatial variability of benthic communities in this fjord. Although the community structure differs significantly between the four stations, these differences remain small overall, suggesting that these stations all belong to the same habitat type. Comparison of the community structure with that of the literature suggests that the four stations are characteristic of outer fjord assemblages.

 Young Sound is a unique location in the Arctic where annual monitoring has been conducted for over 20 years by the Greenland Ecosystem Monitoring - MarineBasis Program. This fjord is therefore an ideal candidate for future biodiversity inventories in northeast Greenland which is one of the Arctic regions with the least studied benthic communities (Piepenburg et al. 2011). Additional sampling will undoubtedly be needed to improve our knowledge of the structure, dynamics and functioning of the Young Sound benthic habitats and associated biodiversity. This sampling should be prioritized in the deeper silty basins (up to 350 m), but also in the inner (i.e. Tyrolerfjord) and intermediate sections of the fjord, which are more exposed to strong freshwater inputs, turbidity and sedimentation. On the opposite, the fjord mouth should also be explored to sample various habitats made up of coarse sands, gravels or potentially biogenic habitats. According to the literature, the innermost parts of the fjord should be dominated by opportunistic species (e.g. the cirratulid polychaete *Chaetozone* spp.) and mollusks feeding on surface deposits such as small thyasirid and nuculanid bivalves (e.g. Włodarska-Kowalczuk and Pearson 2004; Sejr et al. 2010; Włodarska-Kowalczuk et al. 2012), which seem to be underrepresented in the samples analyzed in this study. Finally, the more oceanic part of the fjord should contain a wide variety of benthic habitats characterized by a high and original biodiversity, including biogenic habitats (e.g. maerl/rhodolith or horse mussel beds could be typical in such an environment).

 These data will be crucial for future predictions on the evolution of benthic communities in response to future global changes, and in particular increased freshwater inputs to coastal areas in response to the accelerated melting of the Greenland Ice Sheet (Mouginot et al. 2019; Höning et al. 2023). Although the structure of benthic communities from sill fjords is generally considered to be stable on multi-decadal scales (Renaud et al. 2007), the long-term dynamics of benthic communities in Artic fjords are still poorly understood. To date, few long-term monitoring studies have been carried out in the Arctic, and to our knowledge none in East Greenland. It would therefore be important to carry out such monitoring to ensure community stability over longer time scales, especially when warmer water masses of Atlantic origin are increasing along the Northeast Greenland Shelf (Gjelstrup et al. 2022) and could drive an expansion of boreal species into the region. Complementary sampling methods, including scientific SCUBA diving, have been used in other Arctic regions (e.g., Svalbard) to study shallow benthic communities (i.e., < 30 m, e.g., Jørgensen and Gulliksen 2001). For example, image analysis based on photographs taken by SCUBA divers has been used to study small- scale spatial variability in benthic communities (Jørgensen and Gulliksen 2001) or long-term successional dynamics following episodes of disturbance or climate change (Beuchel et al. 2006; Beuchel and Gulliksen 2008; Al-Habahbeh et al. 2020). Applying these methods to future studies in Young Sound could be an interesting way to obtain complementary data on benthic communities and would also allow comparative studies with Svalbard fjords, which have been studied for several decades.

Author contributions

 FO, LC, MKS conducted the field sampling campaign. GB, LP, JJ, TB, MLD and JG conducted species identifications. GB carried out the statistical analyses, interpreted the results and wrote the original draft of the manuscript. FO, LP, JJ, LC, MKS, TB, MLD, and JG helped to interpret the results and write the manuscript. FO, LC, MKS, and JG obtained the necessary funding for the study. FO, LC, and JG developed the sampling design. All authors read and approved the manuscript.

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Statements and Declarations

 Competing Interests: The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

 Data availability: All data supporting the results and analyses in this article are available in the tables included in Online Resource 1.

References

 Al-Habahbeh AK, Kortsch S, Bluhm BA, Beuchel F, Gulliksen B, Ballantine C, Cristini D, Primicerio R (2020) Arctic coastal benthos long-term responses to perturbations under climate warming. Phil Trans R Soc A 378: 20190355. http://dx.doi.org/10.1098/rsta.2019.0355

- AMAP (Arctic Monitoring and Assessment Programme) (2017) Snow, water, ice and permafrost in the Arctic (SWIPA). Arctic Monitoring and Assessment Programme (AMAP), Oslo. http://hdl.handle.net/11374/1931
- Ambrose WG, Renaud PE (1997) Does a pulsed food supply to the benthos affect polychaete recruitment patterns in the Northeast Water Polynya? J Mar Syst 10:483-495. https://doi.org/10.1016/S0924-7963(96)00053-X
- Anderson MJ, Gorley RN, Clarke KR (2008) PERMANOVA+ for PRIMER: Guide to Software and Statistical Methods. Primer-E, Plymouth
- Anderson MJ, Walsh, DC (2013) PERMANOVA, ANOSIM, and the Mantel test in the face of heterogeneous dispersions: what null hypothesis are you testing? Ecol Monogr 83:557- 574. https://doi.org/10.1890/12-2010.1
- Attard KM, Hancke K, Sejr MK, Glud RN (2016) Benthic primary production and mineralization in a High Arctic fjord: in situ assessments by aquatic eddy covariance. Mar Ecol Prog Ser 554:35-50. https://doi.org/10.3354/meps11780
- Bendtsen J, Gustafsson KE, Rysgaard S, Vang T (2007) Physical conditions, dynamics and model simulations during the ice-free period of the Young Sound/Tyrolerfjord system, p. 46-59. In: Rysgaard S, Glud RN (eds.) Carbon cycling in Arctic marine ecosystems: Case study Young Sound, v. 58. Meddr. Grønland. Bioscience.
- Bendtsen J, Mortensen J, Rysgaard S (2014) Seasonal surface layer dynamics and sensitivity to runoff in a high Arctic fjord (Young Sound/Tyrolerfjord, 74°N). J Geophys Res Oceans 119:6461-6478. https://doi.org/10.1002/2014JC010077
- Berge J, Johnsen G, Nilsen F, Gulliksen B, Slagstad D (2005) Ocean temperature oscillations enable reappearance of blue mussels *Mytilus edulis* in Svalbard after a 1000 year absence. Mar Ecol Prog Ser 303:167-175. https://doi.org/10.3354/meps303167
- Beuchel F, Gulliksen B (2008) Temporal patterns of benthic community development in an Arctic fjord (Kongsfjorden, Svalbard): results of a 24-year manipulation study. Polar Biol 31:913-924. https://doi.org/10.1007/s00300-008-0429-9

 Beuchel F, Gulliksen B, Carroll ML (2006) Long-term patterns of rocky bottom macrobenthic community structure in an Arctic fjord (Kongsfjorden, Svalbard) in relation to climate variability (1980–2003). J Mar Syst 63:35-48. https://doi.org/10.1016/j.jmarsys.2006.05.002

 Bluhm BA, Gebruk AV, Gradinger R, Hopcroft RR, Huettmann F, Kosobokova KN, Sirenko BI, Węsławski JM (2011) Arctic marine biodiversity: An update of species richness and examples of biodiversity change. Oceanography 24:232-248. https://doi.org/10.5670/oceanog.2011.75

 Brandt A, Schnack K (1999) Macrofaunal abundance at 79°N off East Greenland: opposing data from epibenthic-sledge and box-corer samples. Polar Biol 22:75-81. https://doi.org/10.1007/s003000050392

 Bridier G, Meziane T, Grall J, Chauvaud L, Sejr MK, Menneteau S, Olivier F (2019) Coastal waters freshening and extreme seasonality affect organic matter sources, quality, and transfers in a High Arctic fjord (Young Sound, Greenland). Mar Ecol Prog Ser 610:15- 31. https://doi.org/10.3354/meps12857

 Bridier G, Olivier F, Chauvaud L, Sejr MK, Grall J (2021) Food source diversity, trophic plasticity, and omnivory enhance the stability of a shallow benthic food web from a high-Arctic fjord exposed to freshwater inputs. Limnol Oceanogr 66:S259-S272. https://doi.org/10.1002/lno.11688

 Chao A, Chiu CH (2016) Species richness: estimation and comparison. Wiley StatsRef: 594 Statistics Reference Online 1-26. https://doi.org/10.1002/9781118445112.stat03432.pub2

 Cloern JE, Foster SQ, Kleckner AE (2014) Phytoplankton primary production in the world's estuarine-coastal ecosystems. Biogeosciences 11:2477-2501. https://doi.org/10.5194/bg-11-2477-2014

 Colwell RK, Mao CX, Chang J (2004) Interpolating, extrapolating, and comparing incidence- based species accumulation curves. Ecology 85:2717-2727. https://doi.org/10.1890/03- 0557

- Cottier-Cook EJ, Bentley-Abbot J, Cottier FR, Minchin D, Olenin S, Renaud PE (2023) Horizon scanning of potential threats to high-Arctic biodiversity, human health and the economy from marine invasive alien species: A Svalbard case study. Glob Change Biol 30:e17009. https://doi.org/10.1111/gcb.17009
- De Cesare S, Meziane T, Chauvaud L, Richard J, Sejr MK, Thébault J, Winkler G, Olivier F (2017) Dietary plasticity in the bivalve *Astarte moerchi* revealed by a multimarker study in two Arctic fjords. Mar Ecol Prog Ser 567:157-172. https://doi.org/10.3354/meps12035
- Filbee-Dexter K, Wernberg T, Fredriksen S, Norderhaug KM, Pedersen MF (2019) Arctic kelp forests: diversity, resilience and future. Global Planet Change 172:1-14. https://doi.org/10.1016/j.gloplacha.2018.09.005
- Fredriksen R, Christiansen JS, Bonsdorff E, Larsen LH, Nordström MC, Zhulay I, Bluhm BA (2020) Epibenthic megafauna communities in Northeast Greenland vary across coastal, continental shelf and slope habitats. Polar Biol 43:1623-1642. https://doi.org/10.1007/s00300-020-02733-z
- Gaillard B, Meziane T, Tremblay R, Archambault P, Blicher ME, Chauvaud L, Rysgaard S, Olivier F (2017) Food resources of the bivalve *Astarte elliptica* in a sub-Arctic fjord: a multi-biomarker approach. Mar Ecol Prog Ser 567:139-156. https://doi.org/10.3354/meps12036
- Gjelstrup CVB, Sejr MK, de Steur L, Christiansen JS, Granskog MA, Koch BP, Møller EF, Winding MHS, Stedmon CA (2022). Vertical redistribution of principle water masses on the Northeast Greenland Shelf. Nat Commun 13:7660. https://doi.org/10.1038/s41467-022-35413-z
- Glud RN, Rysgaard S, Kühl M, Hansen JW (2007) The sea ice in Young Sound: Implications for carbon cycling, p. 62-85. In: Rysgaard S, Glud RN (eds.) Carbon cycling in Arctic marine ecosystems: Case study Young Sound, v. 58. Meddr. Grønland. Bioscience.
- Gotelli NJ, Colwell RK (2001) Quantifying biodiversity: procedures and pitfalls in the measurement and comparison of species richness. Ecol Lett 4:379-391. https://doi.org/10.1046/j.1461-0248.2001.00230.x
- Gradinger R, Bluhm BA, Hopcroft RR, Gebruk AV, Kosobokova KN, Sirenko BI, Węsławski JM (2010) Chapter 10. Marine life in the Arctic, p. 183-202. In: McIntyre AD (ed) Life in the World's Ocean: Diversity, Distribution, and Abundance. Wiley-Blackwell, Oxford.
- Grall J, Le Loc'h F, Guyonnet B, Riera P (2006) Community structure and food web based on 636 stable isotopes ($\delta^{15}N$ and $\delta^{13}C$) analysis of a North Eastern Atlantic maerl bed. J Exp Mar Biol Ecol 338:1-15. https://doi.org/10.1016/j.jembe.2006.06.013
- Gulliksen B, Holte B, Jakola K-J (1985) The soft bottom fauna in Van Mijenfjord and Raudfjorden, Svalbard, p. 199-215. In: Gray JS, Christiansen ME (eds). Marine Biology of Polar Regions and Effects on Marine Organisms. Wiley, Oslo.
- Hobson KA, Ambrose WG, Renaud PE (1995) Sources of primary production, benthic-pelagic coupling, and trophic relationships within the Northeast Water Polynya: insights from 643 δ^{13} C and δ^{15} N. Mar Ecol Prog Ser 128:1-10. https://doi.org/10.3354/meps128001
- Holding JM, Markager S, Juul-Pedersen T, Paulsen ML, Møller EF, Meire L, Sejr MK (2019) Seasonal and spatial patterns of primary production in a high-latitude fjord affected by Greenland Ice Sheet run-off. Biogeosciences 16:3777-3792. https://doi.org/10.5194/bg-16-3777-2019
- Höning D, Willeit M, Calov R, Klemann V, Bagge M, Ganopolski A (2023) Multistability and Transient Response of the Greenland Ice Sheet to Anthropogenic CO² Emissions. Geophys Res Lett 50:e2022GL101827. https://doi.org/10.1029/2022GL101827
- Jørgensen LL, Gulliksen B (2001) Rocky bottom fauna in arctic Kongsfjord (Svalbard) studied by means of suction sampling and photography. Polar Biol 24:113-121. https://doi.org/10.1007/s003000000182
- Josefson AB, Mokievsky V, Bergmann M, Blicher ME, Bluhm B, Cochrane S, Denisenko NV, Hasemann C, Jørgensen LL, Klages M, Schewe I, Sejr MK, Soltwedel T, Węsławski JM, Włodarska-Kowalczuk M (2013) Marine Invertebrates, p. 225-257. In: Meltofte H (ed.) Arctic Biodiversity Assessment. Status and trends in Arctic biodiversity. Conservation of Arctic Flora and Fauna, Akureyri.

 Kędra M, Gromisz S, Jaskła R, Legeżyńska J, Maciejewska B, Malec E, Opanowski A, Ostrowska K, Włodarska-Kowalczuk M, Węsławski JM (2010) Soft bottom macrofauna of an All Taxa Biodiversity Site: Hornsund (77°N, Svalbard). Polar Res 31:309-326. https://doi.org/10.2478/v10183-010-0008-y

- Kędra M, Pabis K, Gromisz S, Węsławski JM (2013) Distribution patterns of polychaete fauna in an Arctic fjord (Hornsund, Spitsbergen). Polar Biol 36:1463-1472. https://doi.org/10.1007/s00300-013-1366-9
- Knox GA, Lowry JK (1977) A comparison between the benthos of the Southern Ocean and the North Polar Ocean, with special reference to the Amphipoda and the Polychaeta. In: Dunbar MJ (ed) Polar Oceans. Arctic Institute of North America, Calgary, p 423–462
- Kortsch S, Primicerio R, Beuchel F, Renaud PE, Rodrigues J, Lønne OJ, Gulliksen B (2012) Climate-driven regime shifts in Arctic marine benthos. Proc Natl Acad Sci USA 109:14052-14057. https://doi.org/10.1073/pnas.1207509109
- Kotwicki L, Węsławski JM, Włodarska-Kowalczuk M, Mazurkiewicz M, Wenne R, Zbawicka M, Minchin D, Olenin S (2021) The re-appearance of the *Mytilus* spp. complex in Svalbard, Arctic, during the Holocene: The case for an arrival by anthropogenic flotsam. Glob Planet Change 202:103502. https://doi.org/10.1016/j.gloplacha.2021.103502
- Krause-Jensen D, Duarte CM (2014) Expansion of vegetated coastal ecosystems in the future Arctic. Front Mar Sci 1:77. https://doi.org/10.3389/fmars.2014.00077
- Legendre P, Gallagher ED (2001) Ecologically meaningful transformations for ordination of species data. Oecologia. 129:271-280. https://doi.org/10.1007/s004420100716

 Mackie SY (1987) A review of species currently assigned to the genus *Leitoscoloplos* Day, 1977 (Polychaeta: Orbiniidae), with description of species newly referred to *Scoloplos* Blainville, 1982. Sarsia 72:1-28. https://doi.org/10.1080/00364827.1987.10419701

 Mathiesen SS, Thyrring J, Hemmer-Hansen J, Berge J, Sukhotin A, Leopold P, Bekaert M, Sejr MK, Nielsen EE (2017) Genetic diversity and connectivity within *Mytilus* spp. in the subarctic and Arctic. Evol Appl 10:39-55. https://doi.org/10.1111/eva.12415

 McGovern M, Poste AE, Oug E, Renaud PE, Trannum HC (2020) Riverine impacts on benthic biodiversity and functional traits: A comparison of two sub-Arctic fjords. Estuar Coast Shelf Sci 240:106774. https://doi.org/10.1016/j.ecss.2020.106774

- McMeans BC, McCann KS, Humphries M, Rooney N, Fisk AT (2015) Food web structure in temporally-forced ecosystems. Trends Ecol Evol 30:662-672. https://doi.org/10.1016/j.tree.2015.09.001
- Meire L, Mortensen J, Meire P, Juul‐Pedersen T, Sejr MK, Rysgaard S, Nygaard R, Huybrechts P, Meysman FJR (2017) Marine‐terminating glaciers sustain high productivity in Greenland fjords. Glob Change Biol 23:5344-5357. https://doi.org/10.1111/gcb.13801
- Merrill AS, Turner RD (1963) Nest building in the bivalve genera *Musculus* and *Lima*. Veliger 6:55-59.
- Middelbo AB, Sejr MK, Arendt KE, Møller EF (2018) Impact of glacial meltwater on spatiotemporal distribution of copepods and their grazing impact in Young Sound NE, Greenland. Limnol Oceanogr 63:322-336. https://doi.org/10.1002/lno.10633
- Mincks SL, Smith CR, Jeffreys RM, Sumida PYG (2008) Trophic structure on the West 702 Antarctic Peninsula shelf: Detritivory and benthic inertia revealed by $\delta^{13}C$ and $\delta^{15}N$ analysis. Deep Sea Res II 55:2502-2514. https://doi.org/10.1016/j.dsr2.2008.06.009
- Mouginot J, Rignot E, Bjørk AA, van den Broeke M, Millan R, Morlighem M, Noël B, Scheuchl B, Wood M (2019) Forty-six years of Greenland Ice Sheet mass balance from 1972 to 2018. Proc Natl Acad Sci USA, 116:9239-9244. https://doi.org/10.1073/pnas.1904242116
- Ockelmann WK (1958) The zoology of East Greenland: marine Lamellibranchiata. Medd Groenl 122:1-256.
- Oksanen J, Blanchet FG, Friendly M, Kindt R, Legendre P, McGlinn D, Minchin PR, O'Hara RB, Simpson GL, Solymos P, Stevens MHH, Szoecs E, Wagner H (2019) Vegan: community ecology package. R package version 2.5-6. Available at https://CRAN.R-project.org/package=vegan

 Ørberg SB, Duarte CM, Geraldi NR, Sejr MK, Wegeberg S, Hansen JLS, Krause-Jensen D (2022) Prevalent fingerprint of marine macroalgae in arctic surface sediments. Sci Total Environ 898:165507. https://doi.org/10.1016/j.scitotenv.2023.165507

- Pabis K, Sicinski J (2010) Polychaete fauna associated with holdfasts of the large brown alga *Himantothallus grandifolius* in Admiralty Bay, King George Island, Antarctic. Polar Biol 33:1277-1288. https://doi.org/10.1007/s00300-010-0816-x
- Piepenburg D, Schmid MK (1996) Distribution, abundance, biomass, and mineralization potential of the epibenthic megafauna of the Northeast Greenland shelf. Mar Biol 125:321-332. https://doi.org/10.1007/BF00346313
- Piepenburg D, Ambrose WG, Brandt A, Renaud PE, Ahrens MJ, Jensen P (1997) Benthic community patterns reflect water column processes in the Northeast Water polynya (Greenland). J Mar Syst 10:467-482. https://doi.org/10.1016/S0924-7963(96)00050-4
- Piepenburg D (2005) Recent research on Arctic benthos: common notions need to be revised. Polar Biol 28:733-755. https://doi.org/10.1007/s00300-005-0013-5
- Piepenburg D, Archambault P, Ambrose WG, Blanchard AL, Bluhm BA, Carroll ML, Conlan KE, Cusson M, Feder HM, Grebmeier JM, Jewett SC, Lévesque M, Petryashev VV, Sejr MK, Sirenko BI, Włodarska-Kowalczuk M (2011) Towards a pan-Arctic inventory of the species diversity of the macro-and megabenthic fauna of the Arctic shelf seas. Mar Biodiv 41:51-70. https://doi.org/10.1007/s12526-010-0059-7
- Renaud PE, Włodarska-Kowalczuk M, Trannum H, Holte B, Węsławski JM, Cochrane S, Dahle S, Gulliksen B (2007) Multidecadal stability of benthic community structure in a high- Arctic glacial fjord (van Mijenfjord, Spitsbergen). Polar Biol 30:295-305. https://doi.org/10.1007/s00300-006-0183-9
- Renaud PE, Sejr MK, Bluhm BA, Sirenko B, Ellingsen IH (2015) The future of Arctic benthos: Expansion, invasion, and biodiversity. Prog Oceanogr 139:244-257. https://doi.org/10.1016/j.pocean.2015.07.007
- Rysgaard S, Nielsen TG, Hansen BW (1999) Seasonal variations in nutrients, pelagic primary production and grazing in a high-Arctic coastal marine ecosystem, Young Sound,
- Northeast Greenland. Mar Ecol Prog Ser 179:13-25. https://doi.org/10.3354/meps179013
- Rysgaard S, Vang T, Stjernholm M, Rasmussen B, Windelin A, Kiilsholm S (2003) Physical conditions, carbon transport, and climate change impacts in a northeast Greenland fjord. Arct Antarct Alp Res 35:301-312. https://doi.org/10.1657/1523- 0430(2003)035[0301:PCCTAC]2.0.CO;2
- Sejr MK, Jensen KT, Rysgaard S (2000) Macrozoobenthic structure in a high-Arctic East Greenland fjord. Polar Biol 23:792−801. https://doi.org/10.1007/s003000000154
- Sejr MK, Włodarska-Kowalczuk M, Legeżyńska J, Blicher ME (2010) Macrobenthic species composition and diversity in the Godthaabsfjord system, SW Greenland. Polar Biol 33:421-431. https://doi.org/10.1007/s00300-009-0717-z
- Sejr MK, Bruhn A, Dalsgaard T, Juul-Pedersen T, Stedmon CA, Blicher M, Meire L, Mankoff KD, Thyrring J (2022) Glacial meltwater determines the balance between autotrophic and heterotrophic processes in a Greenland fjord. Proc Natl Acad Sci USA 119:e2207024119. https://doi.org/10.1073/pnas.2207024119
- Sirenko BI (2001) List of species of free-living invertebrates of Eurasian Arctic seas and adjacent deep waters. Explorations of the Fauna of the Seas 51:1-129
- Starmans A, Gutt J, Arntz WE (1999) Mega-epibenthic communities in Arctic and Antarctic shelf areas. Mar Biol 135:269-280. https://doi.org/10.1007/s002270050624
- Udalov A, Chikina M, Chava A, Vedenin A, Shchuka S, Mokievsky V (2021) Patterns of Benthic Communities in Arctic Fjords (Novaya Zemlya Archipelago, Kara Sea): Resilience vs. Fragility. Front Ecol Evol 9:777006. https://doi.org/10.3389/fevo.2021.777006
- Voronkov A, Hop H, Gulliksen B (2013) Diversity of hard-bottom fauna relative to environmental gradients in Kongsfjorden, Svalbard. Pol Res 32:11208. https://doi.org/10.3402/polar.v32i0.11208
- Węsławski JM, Kendall MA, Włodarska-Kowalczuk M, Iken K, Kędra M, Legezynska J, Sejr MK (2011) Climate change effects on Arctic fjord and coastal macrobenthic diversity—

 observations and predictions. Marine Biodiversity 41:71-85. https://doi.org/10.1007/s12526-010-0073-9

- Włodarska-Kowalczuk M, Pearson TH (2004) Soft-bottom macrobenthic faunal associations and factors affecting species distributions in an Arctic glacial fjord (Kongsfjord, Spitsbergen). Polar Biol 27:155-167. https://doi.org/10.1007/s00300-003-0568-y
- Włodarska-Kowalczuk M (2007) Molluscs in Kongsfjorden (Spitsbergen, Svalbard): a species list and patterns of distribution and diversity. Polar Res 26:48-63. https://doi.org/10.1111/j.1751-8369.2007.00003.x
- Włodarska-Kowalczuk M, Renaud PE, Węsławski JM, Cochrane SKJ, Denisenko SG (2012) Species diversity, functional complexity and rarity in Arctic fjordic versus open shelf benthic systems. Mar Ecol Prog Ser 463:73-87. https://doi.org/10.3354/meps09858
- Yakovis EL, Artemieva AV, Fokin MV, Varfolomeeva MA, Shunatova NN (2007) Effect of habitat architecture on mobile benthic macrofauna associated with patches of barnacles and ascidians. Mar Ecol Prog Ser 348:117-124.<https://doi.org/10.3354/meps07060>

Figures

Figure 1

- **Figure 1** Map showing the four sampling stations (white dots) and the three main rivers (white
- squares) in Young Sound. River catchment areas are from Bendtsen et al. (2014)

 Figure 2 Principal component analyses based on Hellinger-transformed fauna abundances. Upper plots (a, b) show the first two principal components while lower plots show the first and third principal components. Left plots (a, c) represent the position of each replicate while right plots (b, d) represent the position of the first twenty species that explain at least 60 % of the spatial variation

798 **Figure 3**

Figure 3 Relative abundance (%, a) and total abundance (ind. m⁻², b) of major taxonomic groups

801 from Pass Hytten (from PH 1 to PH 3), Basalt Island (from BI 1 to BI 5), Daneborg (from DA

802 1 to DA 2) and Kap Breusing (KB 1 & KB 2) stations

 Figure 4 Submarine landscapes of Pass Hytten (a, depth = 18 m), Basalt Island (b, depth = 21.5 807 m), Daneborg (c, depth = 20 m) and Kap Breusing stations (d, depth = 20 m). Photo credits: Erwan Amice

Figure 5

 Figure 5 Accumulation curves for each replicate from Pass Hytten, Basalt Island, Daneborg and Kap Breusing stations and the entire Young Sound site (combining 15 replicates). The vertical line represents the number of individuals selected for comparisons between replicates (35 individuals)

 Figure 6 Chao2 (a) and Jackknife1 (b) accumulation curves calculated for Pass Hytten, Basalt Island, Daneborg, Kap Breusing and the four stations combined.

822 **Tables**

823 **Table 1** Geographic coordinates, depth, total area sampled, and material collected (with number

Station		Latitude Longitude	Depth (m)	Total sampled area (m^2)	Collected material
Pass Hytten	N 74.41'	W 20.33'	18.0	0.2700	Macrofauna ($n = 3$) and sediment ($n = 1$)
Basalt Island	N 74.33'	W 20.36'	21.5	0.4500	Macrofauna ($n = 5$)
Daneborg	N 74.31'	W 20.24'	20.0	0.4500	Macrofauna $(n=5)$
Kap Breusing N 74.21'		W 20.11'	20.0	0.1525	Macrofauna ($n = 2$) and sediment ($n = 3$)

824 of replicates in brackets) for each station

- 826 **Table 2** List of the ten most abundant taxa for each station, with their relative abundance (%),
- 827 their cumulative relative abundance (% Cum.), their abundance (Abund., number of individuals
- 828 per square meter [ind. m-²]) and the Pielou index (mean \pm standard error of the mean) of the
- 829 corresponding station (calculated at the scale of the whole community)

