



## 22 **Abstract**

23 Prenatal life is the most sensitive stage of human development to environmental pollutants. Early  
24 exposure to Persistent Organic Pollutants (POPs) may increase the risk of adverse health effects  
25 during childhood. The mechanisms of transference of POPs during pregnancy are still not well  
26 understood. The present study is aimed to investigate the transfer of POPs between mother and  
27 fetus. The concentrations of 14 organochlorine pesticides, 7 polychlorinated biphenyls (PCBs)  
28 and 14 polybromodiphenyl ethers (PBDEs) congeners have been measured in 308 maternal  
29 serum samples, their respective umbilical cords and 50 placental tissues from a mother-infant  
30 cohort representative of Spanish general population. In general, the adjusted lipid-basis  
31 concentrations were higher in maternal serum than in cord serum and placenta. The  
32 concentrations of most pollutants between maternal serum and cord serum and between maternal  
33 serum and placenta were significantly correlated. These distributions were consistent with a  
34 predominant maternal source that transfers the pollutants into the placenta and the fetus.  
35 However, this distribution did not correspond to passive diffusion of these compounds between  
36 these tissues according to lipid content. The compounds more readily metabolized were higher in  
37 newborns, which suggest that differences in metabolic capabilities may be responsible of the  
38 observed variations in POP distributions between mother and newborns. Prenatal exposure to  
39 4,4'-DDT and some PBDEs such as BDE 99 and BDE 209 is much higher than it could be  
40 anticipated from the composition of venous maternal blood. POP exposure assessment studies of  
41 newborns may overlook the effects of some of these pollutants if they only consider maternal  
42 determinations.

43

44 **Highlights**

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- 46 1. The concentrations of most POPs in maternal and cord serum are correlated
- 47 2. A predominant mother-to-fetus pollutant transfer is observed
- 48 3. Pollutant transfer does not respond to passive lipid-associated diffusion
- 49 4. Immature fetal metabolism leads to higher POP accumulation
- 50 5. The degradable PBDEs are in higher concentrations in fetal than maternal serum

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53 **Keywords:** Prenatal exposure, transplacental transfer, maternal serum, umbilical cord serum,  
54 placenta, DDT, DDE, hexachlorobenzene, hexachlorocyclohexane, PCBs, PBDEs.

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59 **Abbreviations:**

60 BDE: bromodiphenyl ether

61 BMI: Body mass Index

62 CB: Chlorobiphenyl

63 HCB: Hexachlorobenzene

64 HCH: Hexachlorocyclohexane

65 GC-ECD: Gas chromatography coupled to electron capture detection

66 GC-MS: Gas chromatography coupled to mass spectrometry

67 INMA: Spanish Children's Health and Environment

68 IQR: Interquartile range

69 LOD: Limit of detection

70 ng/g: Nanogram per gram

71 ng/mL: Nanogram per milliliter

72 NICI: Negative Ionization Chemical Ionization

73 OC: Organochlorine Compound

74 OCP: Organochlorine Pesticide

75 PBDE: Polybromodiphenyl ether

76 PCB: Polychlorobiphenyl

77 POP: Persistent Organic Pollutant

78 P : Percentile

79 SD: Standard deviation

80

## 81 1. Introduction

82 Human exposure to Persistent Organic Pollutants (POPs) begins in the uterine life period by  
83 transplacental transfer (Rogan et al. 1986). Placenta may prevent transfer of some pollutants but  
84 there is evidence that POPs, even those of high molecular weight can reach the fetuses (Vizcaino  
85 et al. 2011). Transfer of contaminants during pregnancy may have implications for fetus health.  
86 Fetuses are more vulnerable than adults to chemical exposure as their immune system and  
87 detoxification mechanisms are not fully developed. In-utero exposure may lead to severe  
88 repercussions for newborns and may predispose to late adult deleterious effects (Boekelheide et  
89 al. 2012). Thus, in utero exposure to POPs, including polybromodiphenyl ethers (PBDEs),  
90 polychlorobiphenyls (PCBs) and organochlorine pesticides (OCPs), has shown to increase the  
91 risk of adverse development outcomes in children (Gascon et al. 2012; Herbstman et al. 2010;  
92 Lopez-Espinosa et al. 2011; Park et al. 2008a; Ribas-Fito et al. 2007; Valvi et al. 2012). These  
93 results have increased notably the interest of the scientific community on exposure to these  
94 compounds during gestation.

95         Consequently, the number of studies reporting prenatal concentrations of POPs has  
96 increased in the recent years. Examination of this previous work evidences difficulties for  
97 comparison since there is a lack of standardization regarding subject selection, timing of  
98 sampling and reported levels (Jakobsson et al. 2012). Placenta, maternal and cord serum are the  
99 most common matrices to assess prenatal exposure to POPs, notwithstanding the processes of  
100 transfer of these pollutants from mother to fetus during pregnancy are still not clear (Barr et al.  
101 2005). Previous studies stated that the distribution in body compartments of chemicals with log  
102  $Kow > 4$  is driven solely by lipid fraction in tissues and blood (Haddad et al. 2000). Accordingly,  
103 partition ratios between matrices of POPs should be close to 1 when adjusted for lipid content.

104 However, there is small experimental evidence from human studies to evaluate this statement.  
105 Some exposure assessment studies have shown good correlations between mother, placenta and  
106 cord serum (Bergonzi et al. 2009) but studies describing the distributions and partition ratios of  
107 POPs between placenta, cord and maternal serum in humans are very scarce and limited to a  
108 reduced number of subjects (Needham et al. 2011) .

109         The present study is aimed to give insight into the transfer of POPs through placenta in a  
110 population exposed to baseline levels by examination of maternal and fetal distribution of POPs  
111 in mother-child pairs and quantification of the partition ratios between placenta, maternal and  
112 cord serum samples. Hexachlorobenzene (HCB),  $\beta$ -hexachlorocyclohexane ( $\beta$ -HCH), 4,4'-DDT  
113 and their principal metabolite 4,4'-DDE, PCBs (PCB 118, 138, 153 and 180) and PBDEs (BDE  
114 28, 47, 99, 153, 154 and 209) have been studied. A mother-infant cohort from Asturias (Spain)  
115 that is representative of Spanish general population has been selected as test case. Except BDE  
116 209 all these compounds are listed in the Stockholm Convention on POPs as priority chemicals.  
117 BDE 209 has been restricted in Europe but as its production and use is still ongoing in most of  
118 the world (EBFRIP, 2009) continued environmental exposure to this compound is expected over  
119 next years.

120

## 121 **2. Material and Methods**

### 122 **2.1 Study Population**

123 The cohort of study was established in Asturias by the Preventive Department of the University  
124 of Oviedo, as part of the INMA –Infancia y Medio Ambiente (Environment and Childhood)  
125 Project. This project encompasses seven Spanish areas and analyzes the influence of prenatal  
126 environmental exposures on growth, development, and health of infants from early fetal life until

127 childhood (Guxens et al. 2012). 494 pregnant women were recruited between May 2004 and  
128 June 2007. Deliveries took place between October 2004 and February 2008 at the reference  
129 hospital San Agustín, in Avilés (Asturias, Spain). 326 cord blood samples were successfully  
130 collected from assistance to 485 childbirths within the cohort population. 308 mother-umbilical  
131 cord blood paired samples were finally available as consequence of this project. Placental tissues  
132 were collected in a subset of 50 women. We present data of POP concentration for the 308 paired  
133 samples available and 50 placenta samples. The characteristics of the mothers from this group of  
134 308 paired samples did not show significant differences from the whole recruited group (data not  
135 shown). The study protocol was approved by the Ethics Committee of the reference hospital, and  
136 informed consent was obtained for every participant.

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## 138 **2.2 Data and sample collection**

139 Maternal blood samples were collected during the first trimester of gestation (median = 12  
140 weeks; range = 10-13 wks). Whole cord blood samples were collected using venipuncture of cord  
141 vessels before the placenta was delivered. Maternal and cord serum were collected after  
142 centrifugation for 10 minutes, separated into aliquotes of 1 ml and stored at  $-80^{\circ}\text{C}$  until analyses.  
143 The whole placenta was collected immediately after delivery. Half of the placenta, including  
144 maternal and fetal sides and central and peripheral parts, was placed in a glass container of a  
145 mixer (Büchi Mixer B-400 Büchi Laboratories AG, Flawil, Switzerland) for its homogenization.  
146 Once homogenized, aliquots of 25 g were stored and frozen at  $-80^{\circ}\text{C}$ . Pregnant women  
147 completed two detailed in-person questionnaires (weeks 10–13 and 28–32) on anthropometric  
148 and sociodemographic characteristics and lifestyle variables.

149

### 150 2.3 Laboratory analyses

151 The laboratory analytical methods and quality control procedures for the analysis of POPs have  
152 been described elsewhere (Grimalt et al. 2010; Vizcaino et al. 2009). Concentrations of 7 PCB  
153 congeners (PCB28, PCB52, PCB101, PCB118, PCB153, PCB138 and PCB180),  $\alpha$ -HCH,  $\beta$ -  
154 HCH,  $\gamma$ -HCH,  $\delta$ -HCH, HCB, PeCB, 2,4'-DDT, 4,4'-DDT, 2,4'-DDE, 4,4'-DDE, 2,4'-DDD,  
155 4,4'-DDD and 14 PBDE congeners (BDE17, BDE28, BDE47, BDE66, BDE71, BDE85,  
156 BDE99, BDE 100, BDE153, BDE154 BDE138 BDE 183 and BDE 190 and BDE 209) were  
157 analyzed in placental, maternal and cord serum samples.

158 Briefly, 1 mL of serum or 1 gr of placental tissue were spiked with the surrogate  
159 standards tetrabromobenzene (TBB) and decachlorobiphenyl (CB 209) and vortex stirred for 30 s  
160 at 2,000 rpm. n-Hexane (3 mL) was added, followed by concentrated sulfuric acid (2 mL). After  
161 reaction, the mixture was stirred for 30 s and the supernatant n-hexane phase was separated by  
162 centrifugation. The remaining sulfuric acid solution was re-extracted twice with 2 mL of n-  
163 hexane (each by 30 s stirring and centrifugation). The combined n-hexane extracts (7 mL) were  
164 additionally cleaned with sulfuric acid (2 mL, stirring 30 s). Then, the n-hexane phase was  
165 separated by centrifugation and reduced to a small volume under a gentle nitrogen stream. The  
166 extract was transferred to gas chromatography (GC) vials using four 25  $\mu$ l rinses of isooctane.  
167 CB 142, BDE 118 (20  $\mu$ l) and [ $^{13}$ C]-BDE 209 (10  $\mu$ l) were added as internal standards before  
168 injection. Organochlorine compounds (OCs) were determined by GC with electron capture  
169 detection (GC-ECD). BDE congeners were analyzed by GC coupled to mass spectrometry in  
170 chemical ionization mode and negative ion recording.

171 Total cholesterol and triglycerides were determined in maternal and cord serum samples using  
172 colorimetric enzymatic methods in the General Biochemistry Laboratory of Hospital San



173 Agustin. The samples were processed using a Roche Diagnostics COBAS C711. Total serum  
174 lipids concentrations were calculated as described by Phillips et al (1989) using the following  
175 formula:

$$176 \quad \text{TL} = (2.27 * \text{TC}) + \text{TG} + 62.3 \text{ mg} \cdot \text{dL}^{-1}$$

177 Placental lipids were determined gravimetrically. 1 gr of placental tissue was  
178 homogenized in 5 ml of chloroform:methanol:hydrochloric acid (20:10:0.1) (v/v/v). After  
179 repeating the process, 10 ml of 0.1 N HCl were added and centrifuged at 3000 rpm for 10 min.  
180 The organic phase was then collected; the non-organic phase was re-extracted and added to the  
181 first extraction product. Total lipid content was determined after drying the organic extracts  
182 under a nitrogen stream to constant weight and total lipid were expressed in grams of lipid per  
183 gram of placenta (Lopez-Espinosa et al. 2007).

184 Validation of analytical results (including POPs and total lipid concentrations) was made by  
185 analysis of reference material obtained from the Arctic Monitoring and Assessment Program  
186 (AMAP). We participate regularly in the AMAP Ring Test Proficiency Program for POPs in  
187 human serum (Centre de Toxicologie Institut National de Santé Publique du Québec, Québec,  
188 Canada) and the laboratory results usually are within 20% of the consensus values, including  
189 lipid concentrations.

190

#### 191 **2.4. Data analysis**

192 POP levels were expressed in ng/mL. They were also adjusted to total serum lipid concentrations  
193 (ng/g lipid). Values of half detection limit were assigned when measurable analyte concentrations  
194 were not found. Spearman correlation and scatter plots were used to examine associations

195 between POP levels in placenta, maternal and cord serum. Placental transfer was evaluated by  
196 calculation of the concentrations ratios between paired samples for each compound on ng/mL  
197 and ng/g lipid:

198 
$$R_{cm} = \frac{C_{uc}}{C_m}; R_{pm} = \frac{C_{uc}}{C_p}$$

199 where  $C_{uc}$  is the umbilical cord concentration,  $C_m$  is the maternal concentration and  $C_p$  is the  
200 placental concentration. Correlations and concentration ratios were calculated excluding non  
201 detected values. Values exceeding three times the standard deviation of the mean were  
202 considered outliers and consequently excluded from the ratio calculations. The ratios of each  
203 compound were only calculated if there were at least 10 paired samples above the detection limit  
204 (Needham, et al. 2011). Regression analyses with forcing regression to 0 were also calculated.  
205 No major differences were found between these options (data non shown). Therefore, only  
206 median concentration ratios are reported.

207

### 208 3. Results

209 Mean maternal age and standard deviation at delivery was 31.5 (4.3) (Table 1). 97.4% of the  
210 mothers were originally born in Spain, 63.5 % were primiparous and around 54% belonged to  
211 the lowest socioeconomic status. More than 42% had finished secondary education and about  
212 76% were employed during pregnancy. 21.5% of the mothers were overweight and 6.8% were  
213 obese according to WHO body mass index (BMI) standards. On average, gestational weight gain  
214 was  $14.1 \pm 5.2$  kg (Table 1).

215 The mean lipid content in placenta was 1.2% (range 0.43-1.7%) and total lipids in cord  
216 and maternal serum were 2.6 g/L (1.7-16 g/L) and 5.3 g/L (3.3-11 g/L), respectively.

217 PCB congeners 52 and 101, PBDE congeners 17, 66, 71, 85, 100, 138, 183 and 190 and  
218 PeCB,  $\gamma$ -HCH,  $\delta$ -HCH,  $\alpha$ -HCH, 2,4'-DDT, 2,4'-DDE, 4,4'-DDD and 2,4'-DDD were usually  
219 below limit of detection. The concentrations of POPs quantifiable in more than 50% of the  
220 samples in at least one of the studied matrices are shown in Table 2. 4,4'-DDE was the pesticide  
221 found in highest abundance and was found above limit of quantification in 100% of all samples  
222 analyzed. HCB was also above limit of quantification in 100% of the maternal serum and  
223 placenta samples and in 98% of the cord serum samples.

224 On lipid basis, the concentrations of organochlorine pesticides (OCPs) were higher in  
225 mothers than in newborns and placental samples (Table 2). 4,4'-DDT was the only exception to  
226 this trend and was higher in newborns. PCBs were found above limit of detection in all maternal  
227 samples. The maternal concentrations of these compounds (median=164 ng/g lipid and range  
228 between 47-1353 ng/g lipid) were much higher than in newborns (median=118 ng/g lipid and  
229 range between 24-967 ng/g lipid) and placenta (median=40 ng/g lipid and range between 10-230  
230 ng/g lipid). In all matrices, the PCB distributions were dominated by PCB 153, followed in  
231 abundance by PCB 180 and PCB 138 in newborns and mothers. In placenta, PCB 153 was  
232 followed by PCB 138 and PCB 28. The median concentrations of total PBDEs were higher in  
233 maternal serum (11 ng/g lipid) than in newborns (5.4 ng/g lipid) and placenta (2.3 ng/g lipid)  
234 (Table 2).

235 Different PBDE congener profiles were found in each of these three types of matrices.  
236 BDE 153 and 154 were most abundant congeners in maternal serum. In placenta the dominant  
237 congener was BDE 209. The median values of all congeners in cord blood serum were below  
238 detection limit but BDE47 was the compound found at higher concentrations in some samples  
239 (Table 2).

240 Spearman correlations between paired samples ranged from weakly negative to strongly  
241 positive (Spearman rho= -0.04 to 0.9; Table 3). HCB and  $\beta$ -HCH showed significant correlations  
242 between maternal and cord serum and placenta ( $p < 0.001$ ). 4,4'-DDE also showed significant  
243 correlations between these matrices. The concentrations of 4,4'-DDT in maternal and cord serum  
244 were significantly correlated but those between maternal serum and placenta were not (Table 4).  
245 The more chlorinated PCB congeners, PCBs 138, 153 and 180, were again showing significant  
246 correlations between maternal serum, cord serum and placenta. The concentrations of the less  
247 chlorinated congeners did not exhibit significant correlations. No correlation was observed for  
248 the concentrations of PBDEs between maternal serum and placenta. However, statistically  
249 significant correlations were observed for BDEs 47, 153 and 154 between maternal and cord  
250 serum (Table 3). No substantial correlation differences were observed when considering the  
251 concentrations on wet weight or lipid weight basis (data not shown).

252 Median  $C_{uc}/C_m$  varied between 0.28 and 0.91 when the concentrations were considered in  
253 ng/mL and between 0.57 and 1.8 when calculating the concentrations in ng/g lipid (Table 4). The  
254 PCB congeners showed different ratios following a trend that was consistent with the number of  
255 chlorine substituents. CB 118 (5 chlorine substituents) had the highest ratio, 0.45, CB138 and  
256 CB153 (6 chlorine substituents) had ratios of 0.39 and 0.37, respectively, and CB180 (7 chlorine  
257 atoms) had a median ratio of 0.28. 4,4'-DDT showed the highest ratio, 0.91, of all studied  
258 compounds. Among PBDEs, BDE 209 showed the highest ratio, 0.8, followed by BDE 99, 0.66,  
259 and BDE 47, 0.58.

260 Median  $C_p/C_m$  varied between 0.36 and 1.2 in ng/mL and between 0.17 and 0.61 in ng/g  
261 lipid.  $\beta$ -HCH presented the highest accumulation in placenta ( $C_p/C_m$  of 1.2 in ng/mL and 0.61 in  
262 ng/ g lipid) and 4,4'-DDE the lowest ( $C_p/C_m$  of 0.36 in ng/mL and 0.17 in ng/ g lipid).

263 Another way to represent the distribution of these pollutants among the three matrices  
264 may be obtained by calculation of the relative percent distribution of the concentrations between  
265 maternal and cord samples (Fig. 1) or maternal, placental and cord samples (Fig. 2). These  
266 relative distributions must be calculated for each individual and the resulting proportions  
267 averaged. Obviously, only the individuals having concentration above quantification limits for all  
268 matrices considered were included in the compound averages. The distributions can also be  
269 calculated using either concentrations in  $\text{ng L}^{-1}$  or  $\text{ng/g}$  lipid. As observed for the ratios in the  
270 second case, the proportions of pollutants in cord serum or placenta increased (Figs. 1 and 2)  
271 because maternal serum has higher lipid content than cord serum and placenta, 5.3, 2.6 and 1.2 g  
272  $\text{L}^{-1}$ , respectively.

273

#### 274 **4. Discussion**

##### 275 *Correspondences in the composition of POPs in maternal and fetal serum and placenta*

276 The maternal and newborn concentrations of the examined pollutants and their relative  
277 distributions differed slightly but the observed concentrations were highly correlated for most  
278 compounds. Thus, HCB,  $\beta$ -HCH, 4,4'-DDT, 4,4'-DDE, PCB 138, PCB 153, PCB 180, BDE 47,  
279 BDE 153 and BDE 154 showed significant Spearman coefficients ( $p < 0.001$ ; Table 3). These  
280 results are in agreement with previous observations in which strong maternal-faetal correlations  
281 for the concentrations of these pollutants were found (Bergonzi, et al. 2011; Eik Anda et al. 2007;  
282 Fukata, et al. 2005; Mazdai, et al. 2003; Meironyté Guvenius, et al. 2003; Waliszewski et al.  
283 2000a) and others involving significant correlations with lower Spearman coefficients (Covaci,  
284 et al. 2002; Jarrell, et al. 2005; Kawashiro et al. 2008; Koopman-Esseboom, et al. 1994; Park, et  
285 al. 2008b). Studies in Slovakia (Park et al. 2008b), Catalonia (Sala et al. 2001) or Belgium

286 (Covaci, et al. 2002) found similar maternal-newborn rates of PCBs, about 20-30% of maternal  
287 concentrations in cord serum on ng/mL as in the present study. In Sweden (Meironyté Guvenius  
288 et al. 2003), Poland (Jaraczewska, et al. 2006), Faroe Islands (Needham, et al. 2011) and Canada  
289 (Muckle, et al. 2001) lipid adjusted concentrations of cord serum ranged from 50 to 90% of  
290 maternal concentrations which is again in agreement with the present results.

291 Conversely, these correlations were not found in other studies (Antignac, et al. 2009;  
292 Gómara, et al. 2007; Kanja et al. 1992; Nair et al. 1996; Sala, et al. 2001; Soechitram, et al.  
293 2004) and in some cases, e.g. Japan (Fukata et al. 2005) or Netherlands (Soechitram et al. 2004),  
294 lipid adjusted concentrations of PCBs were almost equal or slightly higher in cord than in  
295 maternal serum.

296 The correlations observed in the present study were found between maternal venous  
297 blood collected at 12 weeks of pregnancy and cord blood collected at delivery. In general, no  
298 changes in maternal POP concentrations have been observed during gestation (Glynn et al. 2011;  
299 Jarrell et al. 2005; Longnecker et al. 1999; Meijer et al. 2008) although some studies show  
300 discrepant results (Bloom et al. 2009; Bloom et al. 2007; Hansen et al. 2010).

### 301 *Mother- to-fetus POPs transfer*

302 The correlations observed in the present study involve a direct correspondence between  
303 higher maternal and newborn concentrations which is consistent with the transfer of pollutants  
304 from mother to fetus. Accordingly, all compounds exhibiting significant correlation coefficients  
305 between these two matrices have  $C_{uc}/C_m$  ratios  $< 1$  when calculated from ng/mL units or, with the  
306 only exception of 4,4'-DDT, when calculated from ng/g lipid units (Table 4). Representation of  
307 the averaged relative proportion distributions of these pollutants in the maternal newborn serum  
308 pairs also show higher proportion of all correlated pollutants in the former than in the latter when

309 calculated over the ng/mL data as well as over ng/g lipid, with the only exception of 4,4'-DDT in  
310 this last case (Fig. 1).

311 Similarly, a high number of significant Spearman coefficients were observed between  
312 maternal and placental concentrations, e.g. HCB,  $\beta$ -HCH, 4,4'-DDE, PCB 138, PCB 153 and  
313 PCB 180 ( $p < 0.001$ ). Again, these coefficients document a significant association between  
314 higher concentrations of pollutants in maternal serum and placenta which is again consistent with  
315 a transfer from the former to the second. Accordingly, the  $C_p/C_m$  ratios of the compounds  
316 exhibiting significant correlations between these two matrices are  $< 1$  when calculated over  
317 ng/mL (with the only exception of  $\beta$ -HCH) or over ng/g lipid (Table 4) and the averaged  
318 distributions of these compounds between maternal and newborn serum and placenta show  
319 higher relative proportions of these correlated pollutants in maternal serum than in placenta in all  
320 cases except  $\beta$ -HCH when calculated over ng/mL. However, in these maternal-placental  
321 correlations 4,4'-DDT and none of the PBDEs show significant coefficients which constitutes a  
322 distinct feature from the results of the maternal-fetal concentrations.

323 Passive diffusion might control the transport of POPs across membranes (Myllynen et al.  
324 2005) if the concentrations of these compounds tend to distribute uniformly among lipid-rich  
325 tissues (Russell et al. 1999; Waliszewski et al. 2001). In the present study, the concentration  
326 ratios between cord serum and maternal serum or placenta and maternal serum are not close to 1  
327 when lipid adjusted (Table 4) which indicates that other processes also influence the distribution  
328 of these compounds among the different tissues.

329 Pollutant properties such as molecular weight, lipid solubility and protein binding  
330 (Myllynen et al. 2009) could also determine the transfer of pollutants from mother to fetus to a  
331 great extent (Needham et al. 2011). However, statistical analysis did not show any correlation

332 between these concentration ratios and chemical properties of these pollutants such as molecular  
333 weight, molar volume, number of halogen substituents or log octanol water partition coefficient  
334 ( $K_{ow}$ ) (data non shown).

335 These results suggest that other processes besides transfer related to physical-chemical  
336 equilibrium are significant for the distribution of these pollutants between mother, placenta and  
337 fetus.

#### 338 *Selective accumulation in cord blood serum or placenta*

339  $\beta$ -HCH was the only compound which displayed higher concentrations in placenta than in cord  
340 serum (Fig. 2) suggesting that this membrane may act as a partial barrier for this contaminant.

341 The PCB distributions in maternal and cord blood serum and in placenta were quite similar.

342 Some previous studies reported decreases of the relative concentrations of the high chlorinated  
343 PCBs in cord serum (Koopman-Esseboom et al. 1994; Soechitram, et al. 2004) which are not  
344 observed in this study nor in other cohort studies (Carrizo et al., 2006; Vizcaino et al., 2010).

345 Similarly, in some previous studies PCB distributions dominated by the less chlorinated  
346 congeners were reported in placenta (Fernandez et al. 2012; Gómara et al. 2012; Ma et al. 2012;  
347 Needham, et al. 2011) and this is not observed in the present study.

348 In this study, the significant correlations observed between PCBs in maternal serum and  
349 placenta and cord serum showed the highest Spearman coefficients among the congeners of  
350 higher chlorination (Table 3). These strong correlations require a uniform distribution of PCBs  
351 between the three types of matrices as observed in Table 2.

352 Conversely, the concentrations of the PCB congeners of lower degree of chlorination did  
353 not exhibit significant correlations between the three types of samples. These congeners are less  
354 hydrophobic and more difficult to accumulate in human tissues. The observed distributions show



355 concentrations above limit of quantification for all congeners in maternal serum and below limit  
356 of detection for PCB 118 in cord serum and PCB 28 in cord serum and placenta. The significant  
357 correlations of the concentrations of the higher chlorinated PCBs in all three matrices and their  
358 higher abundance in the maternal serum are consistent with the above mentioned distribution  
359 involving a maternal source that transfers these pollutants to the placenta and the fetal cord  
360 blood. However, this transfer is not a passive process related to diffusion into the lipid materials  
361 present in these tissues. Normalization to lipid content does not reflect similar values between  
362 these three sample matrices. Some active mechanisms such as the transport of enzymes through  
363 the membranes are likely associated to this transplacental transfer. These mechanisms may  
364 explain that compounds such as BDE 209 accumulate in cord blood.

365         However, some compounds exhibit specific trends. The  $C_{uc}/C_m$  ratios of 4,4'-DDT are  
366 higher than those of the other OCs. The higher 4,4'-DDT concentrations in newborns may result  
367 from a more rapid transformation to more stable metabolites such as 4,4'-DDE in mothers than  
368 fetus. The latter do not have efficient elimination mechanisms of toxicants (Alcorn and  
369 McNamara 2003), once pollutants cross the placenta they do not have the same capacity for  
370 metabolizing these compounds than their mothers. High concentrations of 4,4'-DDT in cord  
371 blood serum in relation to other OCs have also been observed in other studies (Al-Saleh et al.  
372 2012; Muckle et al. 2001; Pathak et al. 2008; Waliszewski et al. 2000b)

373         Decreasing trends between maternal and fetal PBDE concentration ratios at higher degree  
374 of bromination have been reported in some studies (Frederiksen et al. 2009; Jakobsson, et al.  
375 2012; Meironyté Guvenius, et al. 2003) in which it was concluded that the higher brominated  
376 congeners of these mixtures had more difficulties to cross the placenta. However, none of these  
377 studies analyzed BDE 209 or had sufficiently low detection limits to detect this compound in

378 cord blood serum. In the present study, we have observed higher concentration of BDE 209 in  
379 newborns than in their mothers which is in accordance to other studies (Antignac et al. 2009;  
380 Gómara et al. 2007) which reported an enrichment of BDE209 in cord serum. The presence of  
381 BDE209 in cord serum and placenta indicates its bioavailability and transport across placenta  
382 despite its size. In principle, small molecules penetrate membranes more easily than large ones  
383 (Arnot et al. 2010). However, once large molecules penetrate placenta and reach the fetus it  
384 might be more difficult to eliminate them due to the lower biotransformation capabilities of early  
385 life metabolism (Alcorn and McNamara 2003). BDE 209 was not very frequent in newborns, but  
386 whenever it was found it was the dominant PBDE congeners.

387 Similarly, higher  $C_{uc}/C_m$  ratios have been found for BDE 99 than for BDE 47. These  
388 results can be explained by differences in metabolic transformation between congeners. BDE 99  
389 is usually metabolized to a greater extent than BDE 47 in adults (Stapleton et al. 2009). Previous  
390 studies have found high concentrations of BDE 99 in cord blood (Antignac, et al. 2009; Gómara,  
391 et al. 2007; Kim et al. 2012; Mazdai et al. 2003). Similarly, the high prevalence of BDE 154 in  
392 mothers compared to newborns indicates the difference of metabolic capacities of mothers and  
393 fetus. Higher presence of BDE 154 reflects biotransformation of more brominated congeners  
394 such as BDE 183 (Roberts et al. 2011).

395

## 396 **Conclusions**

397 The distributions of most OCs between maternal serum and cord serum and maternal serum and  
398 placenta are significantly correlated. In general, the highest relative concentrations are found in  
399 maternal serum and the lowest in cord serum. These distributions are consistent with a  
400 predominant maternal source that transfers the pollutants to the placenta and the fetus. However,

401 these distributions do not correspond to pollutant passive diffusion among the three types of  
402 tissues according to their lipid content. Conversely, they require an active transplacental transfer  
403 of the compounds possibly in association to the transport of enzymes through the membranes.  
404 The compounds that can be metabolically transformed, namely 4,4'-DDT and several PBDEs,  
405 have been observed to accumulate selectively in cord blood. Once these are able to reach the  
406 fetus they are better preserved than in the maternal tissues. This difference evidences a low  
407 capacity of fetal metabolism for the degradation of organic pollutants which may lead to the  
408 accumulation of pollutants that usually are found in minor concentrations in adults or in mothers.  
409 POP exposure assessment studies of newborns may overlook the effects of some of these  
410 pollutants if they only consider maternal determinations.

411

412

413 **Acknowledgments and grant support:** The authors are grateful to the mothers who participated  
414 in the study, to the medical board and the gynaecology and paediatric departments of Hospital  
415 San Agustín de Avilés and to the health centre of Las Vegas in Corvera de Asturias. This study  
416 was funded by the Spanish Ministry of Health and Ministry of Science and Innovation (INMA  
417 G03/176, Consolider Ingenio GRACCIE, CSD2007-00067, FISS-PI042018, FISS-PI09/02311),  
418 Obra Social Cajastur, Universidad de Oviedo and ArcRisk (FP7-ENV-2008-1-226534), CROME  
419 (LIFE 12 ENV/GR/001040) and HEALS (FP7-ENV-2013-603946) EU Projects.

420

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629 **Figure captions**

630 Fig. 1. Average percentage distribution of Persistent Organic Pollutants (POPs) between  
631 maternal and fetal serum expressed as ng/ml (a) and lipid adjusted concentrations (b)  
632 Interval bars correspond to 95% confidence interval.

633 Fig. 2. Average percentage distribution of Persistent Organic Pollutants (POPs) between maternal  
634 serum, placenta and fetal serum expressed as ng/ml (a) and lipid adjusted concentrations  
635 (b). Interval bars correspond to 95% confidence interval.

636