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2 **Title:** The challenges and promises of genetic approaches for ballast water management

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9 **Abstract:**

10 Ballast water is a main vector of introduction of Harmful Aquatic Organisms and Pathogens, which
11 includes Non-Indigenous Species. Numerous and diversified organisms are transferred daily from a
12 donor to a recipient port. Developed to prevent these introduction events, the International
13 Convention for the Control and Management of Ships' Ballast Water and Sediments will enter into
14 force in 2017. This international convention is asking for the monitoring of Harmful Aquatic
15 Organisms and Pathogens. In this review, we highlight the urgent need to develop cost-effective
16 methods to: (1) perform the biological analyses required by the convention; and (2) assess the
17 effectiveness of two main ballast water management strategies, *i.e.* the ballast water exchange and
18 the use of ballast water treatment systems. We have compiled the biological analyses required by the
19 convention, and performed a comprehensive evaluation of the potential and challenges of the use of
20 genetic tools in this context. Following an overview of the studies applying genetic tools to ballast
21 water related research, we present metabarcoding as a relevant approach for early detection of
22 Harmful Aquatic Organisms and Pathogens in general and for ballast water monitoring and port risk
23 assessment in particular. Nonetheless, before implementation of genetic tools in the context of the
24 Ballast Water Management Convention, benchmarked tests against traditional methods should be
25 performed, and standard, reproducible and easy to apply protocols should be developed.

26 **Keywords:**

27 ballast water management convention; non-indigenous species; early detection; genetic methods;
28 metabarcoding

29 **1 Introduction**

30 Ballast water discharges are recognized as critical sources of pathogens, harmful algae blooms and
31 Non-Indigenous Species (NIS) introduction (Aguirre-Macedo et al., 2008; Drake and Lodge, 2004;
32 Hallegraeff, 2007; Molnar et al., 2008). To prevent potential environmental, human health and
33 socioeconomic impacts of these introductions, the International Convention for the Control and
34 Management of Ships' Ballast Water and Sediments (referred after as "BWM Convention") was
35 adopted in February 2004 by the International Maritime Organization (IMO, 2004). The BWM
36 Convention will enter into force in September 2017 as the required ratification by at least 30 States
37 representing 35% of world merchant shipping tonnage has finally been reached (IMO, 2016). One of
38 the many challenges of the BWM Convention is ballast water monitoring in commercial ports, *i.e.*
39 screening the whole biodiversity discharged from ballast water to guarantee the prevention and
40 control of Harmful Aquatic Organisms and Pathogens (HAOP) which includes NIS, in recipient ports,
41 which are also monitored for presence of HAOP. Traditionally, ballast water biological inventories,
42 which should preferably be done to the lowest taxonomic level, have relied upon morphological
43 identification, which is costly, time-consuming (Ji et al., 2013) and requires a high level of taxonomy
44 expertise, skill that is becoming rare (Agnarsson and Kuntner, 2007). Besides, one of the most crucial
45 issues associated with this traditional approach is the difficulty to identify early developmental stages
46 (*e.g.* larvae and eggs), broken organisms or morphologically indistinguishable species; all are
47 common in ballast water (Gollasch et al., 2002). Thus, alternative fast, cost-effective, accurate and
48 broadly applicable methods need to be developed in order to improve ballast water monitoring
49 (Lehtiniemi et al., 2015). Genetic methods overcome some of the main limitations of morphological
50 identification and have been described as a challenging revolution in the assessment and management
51 of species diversity (Fonseca et al., 2010; Wood et al., 2013). Specially, numerous reviews highlight
52 the promises of these tools for studying marine biological invasions processes (Rius et al., 2015;
53 Viard et al., 2016), including early detection (Bott, 2015; Comtet et al., 2015) and provide
54 recommendations regarding the associated technical challenges and possible solutions of using High-
55 Throughput Sequencing technologies in such context (Xiong et al., 2016). As a result, a great number
56 of ballast water related studies already use the advantages of genetic tools to describe the in-tank
57 biodiversity of ballast water (*e.g.* eukaryotes, Zaiko et al., 2015b and viruses, Kim et al., 2015) and
58 sediments (*e.g.* diapausing eggs of invertebrates, Briski et al., 2011) as well as to provide helpful
59 biological data for assessing the efficiency of ballast water managements (Briski et al., 2015; Hess-
60 Erga et al., 2010). The rapid development of genetic tools applied to ballast water management have
61 prompted the need for detailed assessments of their relevance and synthetic views of the available
62 tools, as emphasized by Darling and Frederick (2017). While their work provides a comprehensive

63 analysis of the genetic tools for ballast water monitoring in a general context, here, we give a thorough
 64 assessment and analyze the suitability of genetic tools to provide data for the BWM Convention and
 65 their feasibility in real conditions. For that aim, we meticulously studied the BWM Convention to
 66 detect opportunities where genetic tools could provide biological data while evaluating and discussing
 67 their great promise as valuable alternatives or additions by early detecting Harmful Aquatic
 68 Organisms and Pathogens introduced via ballast water.

69 **2 The preventing actions of the BWM Convention and their required biological**
 70 **analyses**

71 The BWM Convention is based on regulations defining the specific legal preventive actions to
 72 manage the introduction of Harmful Aquatic Organisms and Pathogens (HAOP), and on Guidelines
 73 providing technical guidance for all stakeholders to help the implementation of these regulations
 74 (Table 1).

75 Table 1. Regulations and associated Guidelines of the BWM Convention; shaded grey indicates
 76 Guidelines where genetic tools could provide relevant biological data

Regulations (R)	Guidelines (G)
Section A: General provisions	
R. A1: Definitions	
R. A2: General applicability	
R. A3: Exceptions	
R. A4: Exemptions	G7: Guidelines for Ballast Water Risk Assessment
R. A5: Equivalent Compliance	G3: Guidelines for Ballast Water Management Equivalent Compliance
Section B: Ballast Water Management and control requirements for ships	
R. B1: Ballast Water Management Plan	G4: Guidelines for Ballast Water Management and the Development of Ballast Water Management Plans
R. B2: Ballast Water Record Book	
R. B3: Ballast Water Management for ships	G5: Guidelines for Ballast Water Reception Facilities
R. B4: Ballast Water Exchange	G14: Guidelines on Designation of Areas for Ballast Water Exchange G6: Guidelines for Ballast Water Exchange
R. B5: Sediments Management for Ships	G1: Guidelines for Sediment Reception Facilities G12: Guidelines on Design and Construction to Facilitate Sediment Control on Ships
R. B6: Duties of officers and crew	
Section C: Special requirements in certain areas	
R. C1: Additional measures	
R. C2: Warning concerning Ballast Water uptake in certain areas and related flag state measures	G13: Guidelines for Additional Measures regarding Ballast Water Management including Emergency Situations
R. C3: Communication of information	

Section D: Standard for Ballast Water Management

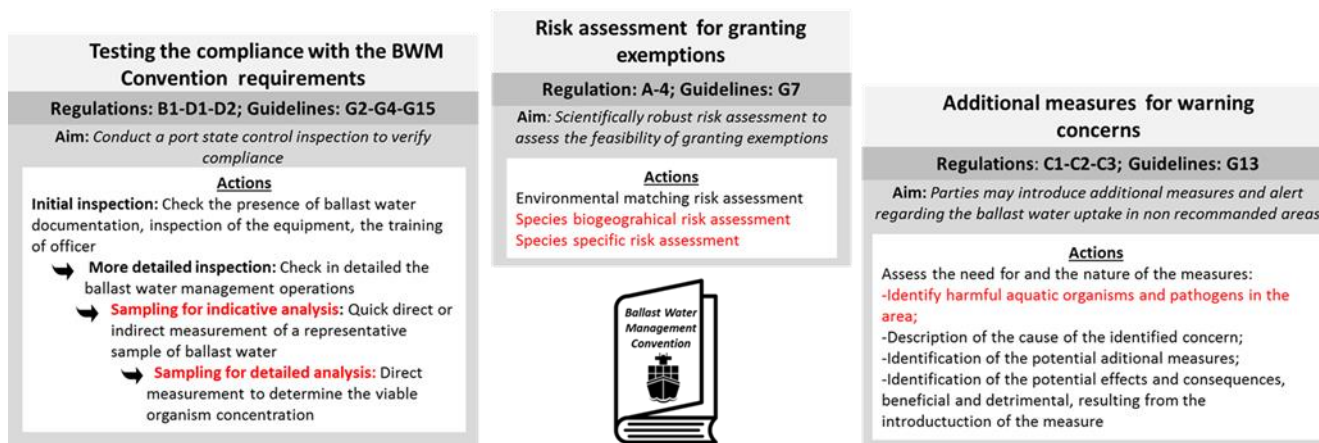
R. D1: Ballast Water Exchange standard	G2: Guidelines on Ballast Water Sampling G6: Guidelines for Ballast Water Exchange G11: Guidelines for Ballast Water Exchange Design and Construction Standards
R. D2: Ballast Water performance standard	G2: Guidelines on Ballast Water Sampling
R. D3: Approval requirements for Ballast Water Management Systems	G8: Guidelines for Approval of Ballast Water Management Systems G9: Procedure for Approval of Ballast Water Management Systems that Make Use of Active Substance
R. D4: Prototype of Ballast Water Treatment Technologies	G10: Guidelines for Approval and Oversight of Prototype Ballast Water Treatment Technology Programs
R. D5: Review of standards by the organization	

Section E: Survey and certification requirements for Ballast Water Management

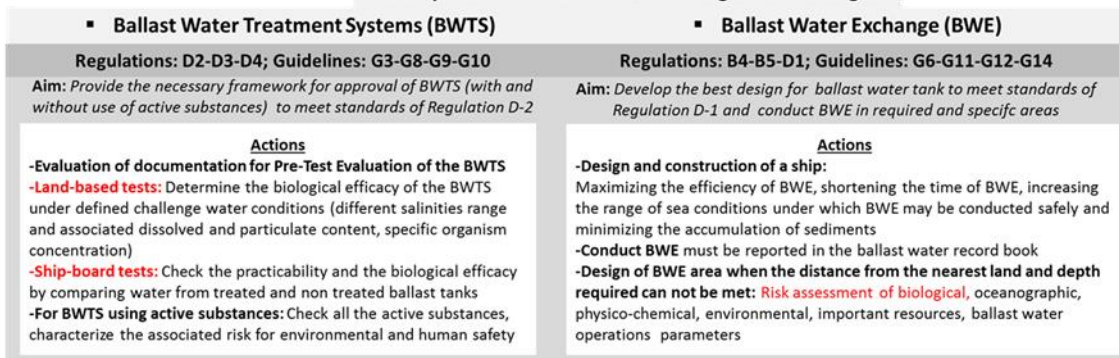
R. E1: Surveys
R. E2: Issuance or Endorsement of a certificate
R. E3: Issuance or Endorsement of a certificate by another party
R. E4: Form of the certificate
R. E5: Duration and validity of the certificate

Article	Guidelines
Article 9: Inspection of ships	G15: Guidelines for Port State Control under the BWM Convention

77



Development of Ballast Water Management strategies:



78

79 Figure 1: Highlights of the main actions with associated Regulations and Guidelines proposed by the
80 BWM Convention grouped in four categories to prevent Harmful Aquatic Organisms and Pathogens
81 (HAOP) introduction. The actions requiring biological analyses are shown in red.

82 To understand the common requirements for ballast water monitoring and facilitate the detection of
83 opportunities where genetic tools could provide relevant biological data for the BWM convention, all
84 preventive actions are grouped in four main categories (Fig. 1) and further described below. To
85 prevent new introduction of HAOP, biological analyses (in red in Fig. 1) are required and the nature
86 of the analyses and method used will be different depending on the targeted action.

87 • Actions related to “Testing the compliance with the BWM Convention requirements”: Ships
88 will be first inspected for compliance with the BWM Convention by the port state control
89 through an administrative control of the documentations required on board (*i.e.* ballast water
90 record book) (King and Tamburri, 2010; Wright and Welschmeyer, 2015); then, if evidences
91 of non-appropriate management are found, ballast water may be sampled for a first and quick
92 indicative biological analysis, and for a more detailed one if additional support for non-
93 compliance is found. The biological methods to assess the compliance must meet certain
94 criteria to be applicable (David and Gollasch, 2015; IMO, 2015); for example, they shall be
95 designed to take into account organismal minimum size, abundance and viability but also to
96 be fast, applicable onboard, and usable by a non-specialist (David and Gollasch, 2015; Wright
97 and Welschmeyer, 2015). Regarding the type of analysis and organisms, the method can be
98 qualitative, semi-quantitative, and quantitative. For now, the recommended methods for
99 detailed analysis are visual counting including mobility test for zooplankton, counting
100 chamber with epifluorescence microscopy as well as machine counts coupled with viability
101 stains for phytoplankton, and grow of bacterial colonies for indicator microbes (David and
102 Gollasch, 2015).

103 • Actions related to “Development of Ballast Water Management strategies”: A major concern
104 of the BWM Convention is to test and approve the efficacy of the two mains ballast water
105 management actions, which are the interim Ballast Water Exchange (BWE) in open sea, at
106 least from 200 nautical miles from the nearest land and at 200 meter depth, and the installation
107 of on-board Ballast Water Treatment Systems (BWTS) such as filtration combined with
108 chemical (*e.g.* chlorination, use of biocides) or physical treatments (*e.g.* UV radiation,
109 deoxygenation) (David and Gollasch, 2015; Stehouwer et al., 2015; Tsolaki and
110 Diamadopoulos, 2010). When the Ballast Water Exchange is not possible because the
111 distance and depth requirements cannot be met, alternative area to do the exchange needs to

112 be designate and biological assessment of the area should be performed to check the presence
113 of Harmful Aquatic Organisms and Pathogens (HAOP). To test the performance of these
114 measures, Regulations D-1 and D-2 standards (IMO, 2004) were set; the former by
115 exchanging ballast water in a rate enough that guarantees “almost” clean waters, and the later
116 by limiting the amount of viable individuals (zooplankton and phytoplankton) and
117 concentration of indicator microbes that can be discharged (Albert et al., 2013). To assess the
118 compliance with Regulation D-2, Ballast Water Treatment Systems must be rigorously tested
119 before approval with similar biological methods as the one used “in real” for testing the
120 compliance of a vessel. During the land-based and ship-board tests for approval, biological
121 methods are used to detect, enumerate, identify viable organisms, and these methods need to
122 take into account the organisms’ rarity in treated water.

- 123 • Actions related to “Risk assessment for granting exemptions”: Shipping companies will
124 likely seek exemptions from applying the BWM Convention to avoid the extra time and
125 investment required for BWE and BWTS. Regulation A-4 states that granting exemptions for
126 5 years is possible if the process follows Guidelines G7, which provides advice regarding
127 scientifically robust risk assessments (IMO, 2007). Three risk assessment approaches have
128 been outlined based on comparison of environmental factors such as temperature and salinity
129 or distribution of HAOP, to assess the likelihood of survival of a transferred species between
130 recipient and donor regions: the environmental matching, the species biogeographical, and
131 the species-specific approaches (IMO, 2007). Biological methods are required for doing such
132 risk assessments and are based on comprehensive port baseline surveys and identification of
133 target species (HELCOM/OSPAR, 2013).
- 134 • Actions related to “Additional measures for warning concerns”: The BWM Convention
135 encourages monitoring ballast water uptake zones, especially in areas known to contain
136 populations of HAOP such as harmful algae bloom species near sewage outfalls (IMO, 2004).
137 Also, parties may develop a higher level of protection against species introduction if they
138 prove the nature of their concern (*i.e.* the potential consequences of the introduction of
139 harmful organisms in the concerned area) and detail the additional measures required. If
140 identification of species is needed, it must be done following a scientific risk assessment at
141 least to the same level of rigor as in Guidelines G7.

142 **3 Overview of the genetic tools and their application in ballast water management**

143 Genetic tools have been increasingly developed to screen ballast water biodiversity and, tightly
144 related, also for monitoring commercial ports (Table 2). In general, genetic tools provide more

145 accurate and cost-effective taxonomic identification compared to a visual taxonomy approach at all
146 life stages such as eggs, spores, larvae, resting stages, juveniles and broken/incomplete adults; this is
147 particularly relevant for Non-Indigenous Species (NIS) planktonic life-history stages (Darling et al.,
148 2008; Harvey et al., 2009; Mountfort et al., 2012), benthic invertebrates and resting stages from ballast
149 tanks sediments (Briski et al., 2011, 2010). Besides, genetic tools proved to be highly sensitive to
150 early detect organisms in very low abundance (Pochon et al., 2013). The workflow of the genetic-
151 based studies includes (1) sampling for target species, (2) extraction of the molecule of interest (DNA
152 or RNA), (3) amplification of a particular region of the genome or transcriptome and analysis of the
153 amplified product(s).

154 (1) Genetic analyses have been performed on all types of samples retrieved from ballast tanks or
155 ports such as sorted individual specimens (Miralles et al., 2016), mixed specimens (*e.g.* zooplankton
156 Brown et al., 2016), filtered water (Zaiko et al., 2015b) or sediment (Briski et al., 2011). From filtered
157 water or sediment, it is possible to extract environmental DNA (eDNA) which refers to the genetic
158 material released in the surrounding environment in form of feces, saliva, tissue, or simply free DNA
159 (Lodge et al., 2012). Analyzed eDNA isolated from several liters of water seems very promising as
160 it allows early identification of taxa present in low abundances such as recently introduced NIS, which
161 are not detectable by traditional means (Lodge et al., 2012; Valentini et al., 2015).

162 (2) Genetic tools target mostly DNA as a molecule for species identification because it is stable,
163 long-lived and can be easily preserved (Pereira et al., 2008). Yet, DNA is limited when detecting
164 living organisms is required, so RNA is usually targeted instead (Cenciarini-Borde et al., 2009;
165 Doblin et al., 2007).

166 (3) Most genetic tools described in Table 2 are based on the Polymerase Chain Reaction (PCR).
167 This technique utilizes primers (short DNA fragments that bind to a specific region of the genome)
168 to amplify a region of interest. Primers can be species-specific (Patil et al., 2005), target a given
169 taxonomic group (Stehouwer et al., 2013a) or even the whole diversity (Zaiko et al., 2015b). In the
170 conventional PCR, the amplified product is monitored at the end of the reaction, and can be either
171 analyzed by electrophoresis to check presence/absence or to differentiate among closely related
172 species (*e.g.* DGGE; see Darling and Blum, 2007), or sequenced. In the Quantitative PCR (qPCR),
173 the amplified product is measured in real time during the PCR allowing to quantify the copy number
174 of the targeted gene (Nathan et al., 2014; Wood et al., 2013). Amplicons from single species are
175 generally obtained using non-specific primers and directly sequenced by Sanger in a process named
176 barcoding (Hebert et al., 2003a). Mix of species can be amplified using species-specific primers to
177 look for a given species and sequenced by Sanger or amplified using more generic primers to amplify
178 several species at the same time and sequenced on High-Throughput Sequencing (HTS) platforms

179 (Shokralla et al., 2012) in a process called metabarcoding (Taberlet et al., 2012), which has been
 180 described as a powerful tool to detect invasive species (Comtet et al., 2015). At the end of the process,
 181 comparison of the sequences to a reference database such as BOLD (www.barcodinglife.org) to
 182 assign barcodes to species or taxonomic groups is required. The choice of the DNA barcode depends
 183 on the research question (e.g. species detection level or broad biodiversity assessment) and the
 184 taxonomic group of interest. Traditionally, the mitochondrial cytochrome c oxidase subunit I gene
 185 (COI) is used for metazoans (Hebert et al., 2003b), the 18S RNA gene for protists (Forster et al.,
 186 2016) and the 16S RNA for bacteria (Muyzer et al., 1993). The metabarcoding approach has been
 187 widely used in ballast water biodiversity monitoring to screen zooplankton (Ghabooli et al., 2016)
 188 and phytoplankton (Steichen and Quigg, 2015). Also, interestingly, the development of HTS is
 189 opening opportunities to assess the biodiversity of usually under looked microorganisms such as
 190 bacteria, virus (Kim et al., 2015; Ma et al., 2009; Xu et al., 2011) and eukaryotic microorganisms also
 191 called protists (Pagenkopp Lohan et al., 2015; Pagenkopp Lohan et al. 2017; Steichen et al., 2014).
 192 Many protists are toxic and pathogen species and are considered to be very abundant in ballast water,
 193 which underlines the importance of detecting them with a passive surveillance genetic approach such
 194 as developed in Pagenkopp Lohan and collaborators (2015).

195

196 Table 2: Examples of studies applying genetic tools to ballast water management

Molecule	Techniques	Target gene(s)	Target organisms	Objectives	References
DNA	qPCR	groEL and tcpA	<i>Vibrio cholerae</i>	To develop molecular methods for rapid monitoring of <i>V. cholerae</i> in ballast water according to the detection levels set by the IMO (1CFU/100 ml)	Fykse et al., 2012
RNA	NASBA				
DNA	qPCR	18S and ctxA	<i>Vibrio cholerae</i> and <i>Potamocorbula amurensis</i>	To test the efficiency of qPCR for detecting <i>V. cholerae</i> and <i>P.amurensis</i> in sediments, biofilm, benthic assemblage and seawater	Mountfort et al., 2012
DNA	qPCR Metabarcoding (Cloning)	18S, 16S, LSU, 23S, myc, hly, himA, lpaH, hsp, ctxA, vvhA	Harmful microalgae, cyanobacteria, Bacteria	To determine the tank variability resulting from ballast age, season, region, vessel characteristics, and exchange status	Burkholder et al., 2007
DNA RNA	qPCR Species-specific sequencing	16S and mcyE	Toxic cyanobacteria: <i>Microcystis sp</i> and <i>Anabaena sp</i>	To investigate harmful microalgae dynamics (viability and toxin production) in ballast water and to provide data for port to-port risk assessment	Doblin et al., 2007
DNA	Species-specific sequencing	SSU and LSU rDNA	Harmful algae: <i>Gymnodinium catenatum</i>	To specifically detect <i>G. catenatum</i> in cultures, in heterogeneous ballast water and environmental samples	Patil et al., 2005
DNA	Species-specific sequencing Metabarcoding (Cloning)	18S,ITS,5.8S, 28S	Nonnative bivalves, crustaceans and algae	To assess the presence of targeted species in ballast water and ocean environmental samples	Harvey et al., 2009
DNA	Metabarcoding (Cloning)	16S	Cultivable bacteria	To explore for the presence of novel micro-organisms species in ballast water	Xu et al., 2011
DNA	RFLP	COI	<i>Carcinus. maenas</i> and <i>C. aestuarii</i>	To detect <i>C. maenas</i> and <i>C. aestuarii</i> larvae in environmental samples	Darling et al., 2008

DNA	RFLP Metabarcoding (Cloning)	16S	Planktonic bacterial community	To compare bacterial community structure of ballast water to the local seawater; and to use this study as background information to risk assessment	Ma et al., 2009
DNA	DGGE Species-specific sequencing	COI and 18S	Asterias larvae	To specifically detect <i>A. amurensis</i> larvae in ballast water samples	Deagle et al., 2003
DNA	DGGE Barcoding	16S	Heterotrophic bacterial community	To study survival and succession of heterotrophic bacteria after disinfection by ultraviolet irradiation or ozonation of seawater	Hess-Erga et al., 2010
DNA	DGGE Barcoding	16S	Phytoplankton	To compare different analytical techniques for phytoplankton counting and identification	Stehouwer et al., 2013a
DNA	DGGE Barcoding	16S	Microbial community	To study the environmental impact of acetate compounds of Peracleans@Ocean Ballast Water Treatment System on microbial dynamic	Stehouwer et al., 2013b
DNA	DGGE Barcoding	18S	Eukaryotic microorganisms and dinoflagellates and diatoms	To examine the eukaryotic microorganism diversity being discharged into a port	Steichen et al., 2014
DNA	DGGE Barcoding	18S	Dinoflagellates and diatoms	To test the viability of phytoplankton transiting via ballast water	Steichen and Quigg, 2015
DNA	Barcoding	COI and 16S	Invertebrate dormant stages	To assess the effect of the saltwater flushing regulations on the density and diversity of invertebrate dormant stages in ballast sediment	Briski et al., 2010
DNA	Barcoding	COI and 16S	Diapausing eggs	To test the accuracy of DNA barcoding as a tool for species-level identification of diapausing eggs of invertebrates found in ballast sediment	Briski et al., 2011
DNA	Barcoding	COI and 16S	Adult macroinvertebrates	To explore the presence and species diversity of adult macroinvertebrates transported by transoceanic and coastal vessels	Briski et al., 2012
DNA	Barcoding	16S	<i>Pseudomonas sp</i> and <i>Vibrio sp</i>	To evaluate a rapid and cost-effective method for monitoring bacteria in ballast water	Emami et al., 2012
DNA	Barcoding	COI and 16S	Zooplankton	To compare the efficiency of ballast water exchange plus ballast water treatment versus ballast water treatment alone	Briski et al., 2015
DNA	Barcoding	COI and 16S	Mollusks	To investigate exotic mollusks in 3 ports of different intensities of maritime traffic in the Cantabrian Sea	Pejovic et al., 2015
DNA	Barcoding	COI	Zooplankton	To examine the genetic information of the known NIS <i>Acartia tonsa</i> in estuaries including the commercial port of Bilbao and to discover any novel invasive species	Albaina et al., 2016
DNA	Barcoding	18 and COI	Invertebrates	To inventory the fouling invertebrate communities from 8 ports in Northern Spain	Miralles et al., 2016
DNA	Barcoding	COI, 12S and 16S	Zooplankton	To identify coastal NIS present in the non-exchanged ballast water: assessment of the effectiveness of ballast water exchange	Ware et al., 2016
DNA	Metabarcoding	16S	Microbial community	To determine the effectiveness of NaOH treatment to reduce microbial community structure in ballast water	Fujimoto et al., 2014
DNA	Metabarcoding	COI	Mollusk <i>Peringia ulvae</i>	To detect mollusk species that could have survived in ballast water	Ardura et al., 2015a
DNA	Metabarcoding	16S	Viral community	To investigate the composition and taxonomic diversity of viruses in ballast water	Kim et al., 2015
DNA	Metabarcoding	V4 and V9 SSU	Protists	To identify microbial eukaryotes present in ballast water	Pagenkopp Lohan et al., 2015
DNA	Metabarcoding	COI	All organisms	To assess the performance of metabarcoding for species detection in ballast water and assess consistency of	Zaiko et al., 2015a

				different sequencing platforms; and possible biases due to DNA decay and/or degradation in ballast water	
DNA	Metabarcoding	RuBisCo and COI	All organisms	To assess the applicability of metabarcoding for detection of organisms in ballast waters during en-route survey	Zaiko et al., 2015b
DNA	Metabarcoding	18S	Zooplankton	To detect NIS with the application of metabarcoding in 16 Canadian ports	Brown et al., 2016; Chain et al., 2016
DNA	Metabarcoding	18S	Zooplankton	To assess community changes in zooplankton transiting in ballast water	Ghabooli et al., 2016
DNA	Metabarcoding	18S	Eukaryotic from early biofouling communities	To assess the potential of metabarcoding to detect the community structure of biofouling eukaryotic assemblages and the patterns of initial succession in ports	Zaiko et al., 2016
DNA	Metabarcoding	V4 SSU	Protists	To compare diversity and community composition of protists in ballast water among and between ports	Pagenkopp Lohan et al., 2017
DNA	Metabarcoding	16S	Bacteria	To analyze diversity and community composition of bacterial assemblages in ballast water of commercial ships following voyages in the North Atlantic Ocean.	Lymperopolou and Dobbs, 2017
DNA	Metabarcoding	16S	Bacteria	To characterize changes in bacterial assemblages of ballast water and assess the effects of ballast water exchange, duration time of the voyage as well as season sampling	Johansson et al., 2017

197 **4 Genetic data for helping the implementation of the BWM Convention**

198 The potential of the most promising genetic tools for performing biological analyses required by the
 199 Guidelines of the Convention are exposed in Table 3, and are further described in sections 4.1 to 4.2.
 200 This analysis comes from crossing the biological analyses required for each Guidelines and the review
 201 of the studies presented in Table 2, where a genetic approach has been used to supply biological data
 202 relevant for BWM Convention.

203 Table 3: Main genetic techniques with their related characteristics for ballast and port water
 204 monitoring and BWM Convention’s Guidelines where they could be integrated (see also Table 1).

Monitoring application (and associated guidelines)

	<u>For ports and specific areas</u>		<u>For ballast water discharges or in tank</u>	
	Biological baseline survey (G7 and G14)	Target species identification (G7, G13 and G14)	Assessing viable community (G2, G8, G10 and G15)	Target species identification (G2, G8, G10 and G15)
Genetic method proposed	Metabarcoding	qPCR	Metabarcoding	qPCR
<i>Key requirements of the BWM Convention:</i>				
<i>Estimation of viability</i>		Not required		RNA or DNA coupled with propidium monoazide: need experimental studies (e.g. mock communities)
<i>Estimation of abundance</i>	Low: relative abundance, still limited	Yes	Low: relative abundance, still limited	Yes

<i>Species level detection</i>	Moderate to high	High	Moderate to high	High
<i>Low abundance detection</i>	Moderate to high	High	Moderate to high	High
<i>Portability</i>	Not required	Not required	Portable machines: PCR, sequencer	Portable qPCR: need further development
<i>Time-Consumption per analysis</i>	Weeks with traditional sequencing platform but could be reduced to hours with portable devices (<i>e.g.</i> MinION)	Hours	Weeks with traditional sequencing platform but could be reduced to hours with portable devices (<i>e.g.</i> MinION)	Hours
<i>Cost: operating costs, consumable, personnel</i>	Low for high number of samples processed	Low	Low for high number of samples processed	Low
<i>Standardization</i>	High potential: need developed protocol	High: need developed species specific primers for all targeted species	High potential: need developed protocol	High: need developed species specific primers for all targeted species

205

206 **4.1 Compliance inspection: Guidelines G2, G8, G10 and G15**

207 Testing the compliance with the discharge standards of Regulation D-2 does not require species
 208 identification (except for indicator microbes) and only relies on the number of viable organisms.
 209 Nonetheless, the Article 6 of the BWM Convention asks to each Party to promote relevant information
 210 on the effectiveness of ballast water management obtained from any monitoring program (IMO,
 211 2004). In this context, identification of organisms discharged is beneficial to assess the efficiency of
 212 the ballast water management measures and thus reduce the events of introduction of Harmful Aquatic
 213 Organisms and Pathogens, at the same time, early detect introduction of a potential NIS.

214 **4.1.1 Determine viability of organisms**

215 Viability is the key criteria to assess if a vessel will comply with Regulation D-2 and the universal
 216 method to detect only viable organisms is yet to be developed. Nonetheless, genetic tools hold great
 217 potential if their targets are linked with viability such as using RNA or if the technique used prevent
 218 the amplification of DNA from dead cells such as with the application of propidium monoazide
 219 (PMA) (Fujimoto et al., 2014). Doblin et al (2007) shows the potential of RNA analyses to reveal that
 220 harmful algae species could synthesis active toxin after 11 days in a ballast water tank. To assess the
 221 living community discharged with ballast water targeting RNA instead of DNA for metabarcoding
 222 has been proposed as an appealing solution (Zaiko et al., 2015a, 2015b) since studies already
 223 successfully developed protocols based on RNA metabarcoding to screen living protists (Pawlowski
 224 et al., 2014; Visco et al., 2015). Yet, RNA is very difficult to preserve, and thus working with RNA
 225 is more challenging when compared to DNA. Nonetheless, several studies have proposed optimized
 226 protocols to extract high quality RNA (Asai et al., 2015; Zhang et al., 2013). Concerning the use of

227 propidium monoazide, it seems well defined for microbes as Fujimoto et al (2014) used it to show
228 alkali ballast water treatment efficacy to reduce alive microbial diversity. Nonetheless, its potential
229 remains unclear for multicellular organisms such as for zooplankton tested in compliance (Lance and
230 Carr, 2012; Zetsche and Meysman, 2012) and further research is needed.

231 **4.1.2 Test the efficacy of ballast water management**

232 Genetic methods have been successfully used to provide biological data on the effects of ballast water
233 management by identifying the community found in the tanks or at the discharge line after the Ballast
234 Water Exchange (Briski et al., 2015, 2010; Ware et al., 2016), by assessing the efficacy of Ballast
235 Water Treatment Systems to reduce bacteria and phytoplankton diversity (Fujimoto et al., 2014;
236 Stehouwer et al., 2013b, 2013b) and by identifying the composition of “recolonizer” organisms after
237 the application of a treatment (Hess-Erga et al., 2010). A future interesting direction will be to develop
238 RNA metabarcoding (see 4.1.1) to provide a better depiction of “the living organisms” after the
239 application of the Ballast Water Treatment System.

240 **4.1.3 Assess the representativeness of sampling to inspect compliance**

241 Sampling for compliance testing requires that samples must be representative of the whole ballast
242 water discharged (IMO, 2008). The large volumes of water, tank characteristics (shape, size and
243 number) and heterogeneous distribution along with low density of organisms within tanks make very
244 challenging to develop the best representative approach for compliance sampling and research is still
245 going on (Basurko and Mesbahi, 2011; Carney et al., 2013; Hernandez et al., 2017). To our
246 knowledge, genetic tools have never been applied to such use yet. Metabarcoding could be very useful
247 to assess if there is a significant difference in the communities inferred using alternative sampling
248 approaches. Indeed, the volume, number and frequency of samples, and the availability to sample
249 from the discharged lines or directly from the tank are factors that need to be tested and optimized to
250 get an accurate representation of the discharged community.

251 **4.2 Target species detection and biological monitoring: Guidelines G7, G13 and G14**

252 **4.2.1 Risk assessment for granting exemptions**

253 HELCOM/OSPAR Guidelines have been developed in line with the Guidelines G7 to provide advice
254 on risk assessment procedures for granting exemptions from the port survey sampling to the decision
255 support and administrative decisions (HELCOM/OSPAR, 2013). The biological methods proposed
256 are based on considerable sampling effort of all types of organisms to detect Non-Indigenous Species
257 (NIS) and native harmful species and on checking the presence of potential pathogens. The analysis
258 of NIS is done with a visual taxonomic identification approach. The port of Rotterdam, for example,

259 has conducted such survey and 32 NIS were detected over 225 species identified to the species level
260 recorded in more than 250 samples taken from 118 different locations (Gittenberger et al., 2014). The
261 species accumulation curves indicated that more sampling would have resulted in scoring more
262 species for several habitats of the port. So, despite such important sampling effort, the entire diversity
263 of the habitats was not recovered. Genetic methods, by enhancing detection sensitivity, increasing
264 specificity of target identification and reducing monitoring time and costs, hold great promise over
265 traditional methods (Darling and Mahon, 2011) and has been proposed as a cost-effective method in
266 this context (Comtet et al., 2015; Lehtiniemi et al., 2015).. Genetic-based background information to
267 assess the risk associated with Harmful Aquatic Organisms and Pathogens in ports are numerous.
268 Examples are Ma et al., 2009 for pathogens, Miralles et al., 2016; Pejovic et al., 2015 for invasive
269 mollusks, Doblin et al., 2007 for harmful algae species, Brown et al., 2016 for zooplankton, Hirst and
270 Bott, 2016 for fouling organisms and zooplankton, Pochon et al., 2015; Zaiko et al., 2016 for early
271 fouling organisms. One very interesting study in Canadian ports shows that metabarcoding detected
272 24 NIS species of zooplankton of which 11 were firstly reported in this regions (Brown et al., 2016).
273 Also, the combination of sampling artificial settlement plates with DNA metabarcoding based species
274 detection seems very promising to early detect potential invasive species in ports where species are
275 at very low densities and morphologically undistinguishable early life stages occur (Pochon et al.,
276 2015).

277 **4.2.2 Specific ballast water uptake zone**

278 As described in section 2. The preventing actions of the BWM Convention and their required
279 biological analyses in Actions related to “Development of Ballast Water Management Strategies” and
280 “Additional measures for warning concerns”, biological monitoring is required when there is a need
281 to design alternative ballast water exchange areas and when a Party shall warn about the uptake of
282 ballast water in certain zones. In both cases, metabarcoding appear as sensitive methods to perform
283 such biological monitoring as they allow to detect sewage contamination environment by identifying
284 sewage-related bio-indicator microbial species (Tan et al., 2015). Also, qPCR has been demonstrated
285 to play a major role in identifying harmful algae bloom species, understanding their ecology as well
286 as facilitating the management of their outbreaks (Antonella and Luca, 2013).

287 **5 Future investments for the use of genetic tools in ballast water management**

288 **5.1 Standardization and further development of genetic methods**

289 Implementation of genetic methods for regular application in ballast water monitoring requires
290 benchmarking against traditional methods and standardization of procedures (Aylagas et al., 2016;

291 Aylagas and Rodríguez-Ezpeleta, 2016). Each step, from sampling to taxonomic assignment, can be
292 performed using alternative protocols and the choice of the most appropriate will depend on i) type
293 of sample, *e.g.* water, sediment, fouling, ii) taxonomic group tackled, *e.g.* bacteria, protists, metazoa,
294 iii) pursued aim, *e.g.* to obtain a comprehensive species inventory or look for a specific species, iv)
295 and other specific conditions such as the need to detect only living organisms. It is thus crucial that
296 reproducible protocols are defined, tested and broadly communicated so that results obtained can be
297 comparable and usable for regular monitoring purposes (Cristescu, 2014). Also, to validate protocols,
298 one key element is the use of artificial communities to i) develop rigorous laboratory workflows (Port
299 et al., 2016) and standard bioinformatic pipelines (Brown et al., 2015; Flynn et al., 2015) and to ii)
300 assess the sensitivity of genetic tools such as metabarcoding to detect invasive species in complex
301 environmental samples (Pochon et al., 2013). Also, one important technical shortcoming of DNA
302 metabarcoding is the PCR bias (amplification success is not equal for all taxa present in a sample)
303 which can lead to false negatives and impede estimating taxa abundances based on sequences
304 (Elbrecht and Leese, 2015). To overcome PCR bias, community genome sequencing (metagenomic)
305 could be applied (Taberlet et al., 2012). Although this approach shows promising results regarding
306 their use to species identification and estimate relative abundance (Gómez-Rodríguez et al., 2015;
307 Srivathsan et al., 2015; Tang et al., 2015, 2014), it is much more expensive and requires more
308 complicated data analyses that rely on the availability of large numbers of complete genomes in
309 databases, which is currently only existing for bacteria.

310 **5.2 Improvement of reference databases**

311 Both DNA barcoding and metabarcoding rely on the availability of a well populated and curated
312 reference database that associates DNA sequences with species names (*e.g.* BOLD
313 www.barcodinglife.org, Silva www.arb-silva.de). Unfortunately, such databases are far from being
314 complete (Aylagas et al., 2014; Comtet et al., 2015) and are usually biased towards certain organismal
315 groups (Briski et al., 2016). Recently, significant effort has been devoted to increase reference
316 databases, and the number of sequences, especially from eukaryotes, has notably increased during the
317 past few years (Briski et al., 2016). Additionally, the AquaNIS database
318 (www.corpi.ku.lt/databases/aquanis) which constitutes the reference information system on aquatic
319 non-indigenous and cryptogenic species, is also gathering molecular information among all the
320 information recorded for a species and has undertaken a promising step forward action to integrate
321 molecular data into the management of NIS. Also, the metabarcoding research is moving forward to
322 the use of a multiple marker approach (*e.g.* COI and 18S) instead of a single marker which help
323 providing a better catalogue of biodiversity (Cowart et al., 2015; Marcelino and Verbruggen, 2016;

324 Zaiko et al., 2015b). This approach is thus, strongly encouraged for future studies assessing the ballast
325 water discharged community.

326 **5.3 Increasing practicability of genetic tools in the field**

327 The BWM Convention requires fast, portable and user-friendly sampling and analysis protocols
328 which appears difficult to achieve with genetic tools because most require specialized equipment and
329 training. Nonetheless, several alternatives have been designed and could be considered for detecting
330 species in ballast water samples such as combining “minutes”-DNA/RNA extraction techniques with
331 portable DNA/RNA amplification devices such as portable PCR or Loop-mediated isothermal
332 amplification (LAMP) machine (Agrawal et al., 2007; Lee, 2017). LAMP is commonly used for food
333 safety testing for pathogens and fungi as it has a very short reaction time, only requires a heating
334 block to perform the amplification and holds high sensitivity for species detection (Niessen et al.,
335 2013). Also, in the context of passive surveillance (*i.e.* screening the biodiversity) of ballast water,
336 the development of the Oxford Nanopore Technologies® MinION device, a miniaturized and
337 portable real-time high-throughput sequencer, shows great promises to reduce the time of samples
338 processing and facilitate *in situ* monitoring for environmental research (Edwards et al., 2016;
339 Mitsuhashi et al., 2017; Ramgren et al., 2015). Mitsuhashi et al (2017) shows the possibility of rapid
340 sequencing and bacterial composition identification within 2 hours. Nonetheless, the in-field
341 preparation of samples (homogenization, DNA/RNA extraction and library preparation) as well as
342 the bioinformatic analyses still needs optimized protocols and further adjustments to be used routinely
343 in ballast water monitoring.

344 **5.4 Exploring alternative sampling strategies**

345 eDNA appears very promising source from which to infer biodiversity that reduces sampling effort
346 especially in a context of port baseline survey or targeted species identification where exhaustive
347 sampling is usually required. The field of sampling eDNA for detecting invasive species is of growing
348 interest as numerous studies show its great potential (Ardura et al., 2015b; Jerde et al., 2011; Simmons
349 et al., 2015), and eDNA should be a focal point of research efforts to apply its use in ports monitoring
350 in a near future. Yet, eDNA degradation appears to be influenced by environmental conditions and
351 the approximate time range can go from less than one day to a couple of weeks (Barnes et al., 2014;
352 Thomsen et al., 2012) challenging its use as a proxy for the living organisms present in an environment
353 such as a ballast water tank. Future work is needed to disentangle the complex relationship between
354 eDNA persistence in ballast tank and the actual presence of organisms, especially in a context of
355 ballast water sampling for compliance. But, interestingly, monitoring eDNA in ballast water can
356 detect species that survive the harsh conditions of a ballast water tank by showing an increase in

357 proportion of DNA signal during the journey (Ardura et al., 2015a; Zaiko et al., 2015b). These studies
358 show the potential of eDNA sample to monitor ballast water biodiversity to potentially detect trend
359 in species that could need a special attention to be monitored in case of discharge of ballast water into
360 a port.

361 **5.5 Efforts to integrate of genetic methods into monitoring programs**

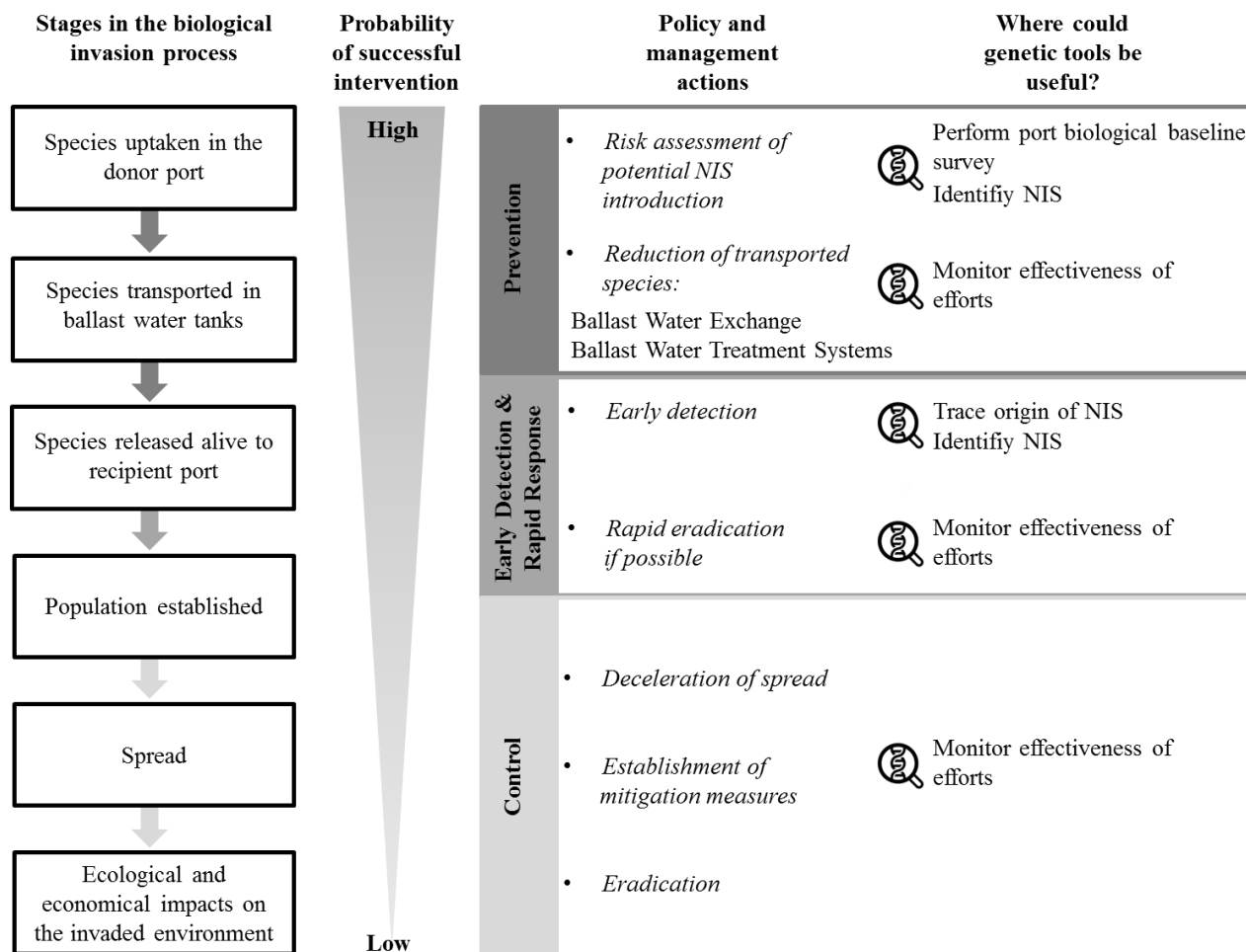
362 Despite the potential of genetic tools for ballast water monitoring and management, the application
363 of these tools by stakeholders remains still experimental, as it has associated challenges (Bucklin et
364 al., 2016; Darling, 2015; Darling and Mahon, 2011; Kelly et al., 2014; Wood et al., 2013). Only very
365 few countries such as New Zealand or Australia started to integrate genetic tools to early detect
366 invasive species and harmful algal blooms species (Hirst and Bott, 2016; Wood et al., 2013). One
367 major issue is the difficult communication between researchers and stakeholders, which makes the
368 former unaware of the real problems and the latter unknowledgeable of the potential of genetic tools
369 (Darling, 2015). To provide an effective integration of genetic methods into monitoring programs,
370 geneticists, environmental monitoring agencies and ballast water policy stakeholders (*e.g.* Port State
371 Control) should together define standardized processes so that genetic methods can provide reliable
372 data for decision making (*e.g.* interdisciplinary theoretical and hands-on workshops are encouraged).

373 **6 Conclusions and future directions**

374 The analyze of the research studies using genetic tools for ballast water monitoring exposed the great
375 potentials of such tools to be applied from the risk assessment to the early detection and monitoring
376 of Non-Indigenous Species and so, provide valuable information at all stages of the invasion process
377 (Fig. 2).

378 Given their potential in terms of accuracy and cost-effectiveness along with their fast-growing
379 evolution, genetic tools, with an emphasize on metabarcoding, should definitely continue to be
380 developed for ballast water monitoring. DNA metabarcoding and qPCR standardized protocols need
381 to be created for performing respectively, port baseline surveys and target species detection required
382 in the BWM Convention in Guidelines G7, G13 and G14. Nonetheless, genetic tools are facing
383 technical challenges associated to the techniques itself (*e.g.* lack of completed reference database,
384 standardized protocols) or specific to ballast water monitoring. Indeed today, genetic tools do not
385 fulfill all the key criteria of the BWM Convention such as portability and rapidity of results or
386 viability assessment. Yet, extensive experimental research is needed to assess the potential of RNA
387 to detect the viable community discharged via ballast water (*e.g.* use of artificial communities), as
388 well as testing the feasibility for on-board genetic analyses. Also, the evaluation of the number and

389 minimum dimension of organisms discharged via ballast water remains very limited with genetic
 390 tools alone and needs to be combined with other appropriate tools. But we are confident that the
 391 worldwide burning development of DNA/RNA-based environmental monitoring is leading to better
 392 refinement of these tools for qualitative assessment of biodiversity discharged via ballast water.



393 Figure 2: Stages of the ballast water induced invasion process with the corresponding management
 394 actions (adapted from Lodge et al., 2006) and usefulness of potential genetic tools

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