- 1 Title: Mineral control of organic carbon storage in acid temperate forest soils in the
- 2 Basque Country
- 3
- 4 Author names and affiliations: Nahia Gartzia Bengoetxea^a, Iñigo Virto^b, Ander Arias-
- 5 González^a, Alberto Enrique^b, Oihane Fernández-Ugalde^{c1}, Pierre Barré^d
- ⁶ ^a NEIKER-Tecnalia, Department of Forest Science, 812 Bizkaia Science and
- 7 Technology Park, 48160 Derio, Spain
- ⁸ ^b Departamento Ciencias, ETSIA. IS-FOOD, Institute for Innovation & Sustainable
- 9 Development in Food Chain, Universidad Pública de Navarra, Campus Arrosadia,
- 10 31006 Pamplona, Spain
- ^c Department of Mineralogy and Petrology, Science and Technology Faculty, University
- 12 of the Basque Country (UPV/EHU), Apartado 644, 48080 Bilbao, Spain
- ¹³ ^d Laboratoire de Géologie de l'ENS, PSL Research University, UMR8538 du CNRS, 24
- 14 rue Lhomond, 75231 Paris cedex 05, France
- 15
- 16 Corresponding author: Dr Nahia Gartzia-Bengoetxea
- 17 ngartzia@neiker.eus
- 18 Tel: (34) 94 403 4388
- 19 Fax: (34) 94 403 4310
- 20
- 21 ¹ Present address:
- 22 European Commission, Joint Research Centre (JRC), Directorate for Sustainable
- 23 Resources, Land Resources Unit, Via E. Fermi 2749, I-21027 Ispra (VA), Italy

24 Abstract

25 Soil minerals are known to play a major role in soil organic carbon (SOC) stabilization. 26 However, the exact nature of the role is still poorly understood as comparative studies 27 are complicated by the fact that mineralogy usually varies with other important factors 28 such as climate, land-use, texture and acidity. This study investigated naturally-29 occurring minerals and C storage in nine mature, second-rotation *Pinus radiata* D. Don 30 forest plantations at a similar development stage and growing under similar climatic 31 conditions, but established on soils derived from three different parent materials: 32 sandstone, basalt and trachyte. We quantified the SOC stocks in the 0-20 cm soil layer, 33 clay-size minerals and metal oxides. The distribution of SOC in labile and mineral-34 bound fractions and the chemical structure of the SOC were also determined, by 35 CPMAS NMR spectroscopy. For similar tree biomass, mean SOC storage was 2 times 36 higher in the soils derived from basalt and 2.5 times higher in the soils derived from 37 trachyte which were relatively rich in chlorite/vermiculite mixed layers and short-range 38 order aluminium and iron forms, than in the sandstone-derived soils dominated by illite 39 and with no short-range order aluminium. The parent material also affected the major C 40 components of the SOC. Thus, the proportion of aromatic-C forms in the total SOC was 41 higher in the sandstone-derived soil than in the soils of volcanic origin. Furthermore, the 42 proportion of SOC in the mineral-bound fraction was around 30% in the sandstone-43 derived soil and around 50% in the volcanic soils. We conclude that estimates of 44 ecosystem SOC stocks in acid temperate forest soils should include proxies for clay 45 mineralogy, such as phyllosilicate minerals, contents of Fe and Al (hydr-) oxides and 46 non-crystalline compounds. In addition, we recommend the inclusion of soil parent 47 material as a proxy in models aimed at estimating regional SOC stocks.

48

49	Keywords: Forest soils, soil mineralogy, soil organic C, oxides, phyllosilicates
50	
51	Highlights:
52	• SOC differed in forest soils with similar pH/clay content but different bedrock
53	• Clay mineralogy differed in sandstone-derived and volcanic soils
54	• The amounts and quality of SOC may be related to mineralogical differences
55	• Soil parent material could be included as a proxy in SOC model development
56	
57	

59 1. Introduction

60 Soil organic matter (SOM) is considered a continuum of progressively decomposing 61 organic compounds, ranging from large plant and animal residues to more oxidized 62 compounds of smaller molecular size, including compounds of microbial origin 63 (Lehmann and Kleber, 2015). Stabilization of SOM depends on the multiple interactions 64 between non-organic soil components and the continuum of progressively decomposing 65 organic compounds. Recent research stresses the need for global, regional and local 66 studies of SOM stabilization to take into account the nature of these soil components 67 and also the physical-chemical characteristics of the matrix in which the components 68 interact (Rasmussen et al., 2018; Rowley et al., 2018). In this context, information about 69 SOM stabilization mechanisms, as affected by the composition of the soil mineral 70 fraction, is needed to improve soil carbon management and to enable climate-related 71 targets and the United Nations' sustainable development goals to be met. 72 Regional studies have demonstrated a relationship between the soil parent material and 73 organic C storage in topsoil (e.g. Gray et al., 2015; Silva et al., 2016; Barré et al., 2017). 74 The mechanisms underlying such interactions are, however, not straightforward and are 75 probably dependent on local conditions (Barré et al., 2017). Recent research findings 76 indicate that, among the different soil properties associated with soil genesis, the 77 mineralogical characteristics play an important role in SOC stabilization (Singh, et al., 78 2018; Barré et al., 2014). Interactions between minerals and organic matter are known 79 to be affected by the properties of the minerals, such as specific surface area and the 80 capacity to bind to organic moieties in soil through electrostatic interactions (Six et al. 81 2002a; Kleber et al., 2007; Feng et al. 2013). Chemical recalcitrance is not always 82 associated with mature, stable organic matter in soils (Kleber et al., 2011; Derrien et al., 83 2016), and it is the interactions with soil minerals at different levels that protect organic

84 components from further degradation (Lehmann and Kleber, 2015). Thus, the ability of 85 soil minerals to form *primary* and *secondary* organo-mineral complexes (as defined by 86 Christensen, 1996) is a determining factor in long-term SOC stabilization. 87 The properties of soil minerals are primarily defined by the parent material, but are also 88 affected by weathering processes. Regarding the reactive minerals present in the clay 89 fraction, parent materials such as granite and most common sedimentary rocks favour 90 the formation of illite-like minerals, while ferromagnesian igneous rocks favour the 91 formation of chlorite-vermiculite mixed layer minerals (Wilson et al., 1984). In soils 92 derived from parent materials rich in Al and Fe, the secondary minerals formed via 93 weathering of primary minerals generally also include Fe-oxides (e.g. ferrihydrite and 94 goethite), Al-oxyhydroxides (e.g. gibbsite), short-range order (SRO) or non-crystalline 95 minerals (e.g. allophane and imogolite) and phyllosilicates, including halloysite and 96 kaolinite (Lawrence et al., 2015). The SRO minerals are known to contribute to SOC 97 stabilization in soils derived from igneous rocks (e.g. Torn et al., 1997). Coarse-textured soils have also received some attention in this regard (Grand and Lavkulich, 2011; 98 99 Grand and Lavkulich, 2015). In such soils, dissolved weathering products, including Al 100 and Fe, react with organic ligands to form organo-metal complexes (e.g., Matus et al., 101 2006; Sollins et al., 1996). These stable complexes determine Al/Fe solubility in soils 102 (Berggren and Mulder, 1995; Dijkstra and Fitzhugh, 2003), and they have been shown 103 to influence the precipitation of secondary minerals (Lawrence et al., 2014). Several 104 studies have also demonstrated a strong significant correlation between sodium 105 pyrophosphate-extractable Al/Fe and total soil C (e.g. Masiello et al., 2004; Wagai et 106 al., 2011), implying a widespread association between organo-metal complexation and 107 long-term storage and stability of SOC in soils in which these elements are abundant.

108 The nature of organic components of soil and their ability to be incorporated into 109 organo-mineral structures at different levels are also important aspects of organomineral complexation. In addition to the nature of the soil organic components, and in 110 111 relation to direct complexation (primary complexes as defined by Christensen, 1996), a 112 number of determining factors -such as the physico-chemical characteristics of the soil 113 solution- determine the various possible models of binding of organic molecular 114 fragments in these complexes (Kleber et al., 2007). 115 Selective preservation due to the specific chemical structure of SOM compounds and

116 fractions found in association with minerals has been proposed as a mechanism for

117 protecting SOC (Marschner et al., 2008). Some particular types of SOM compounds are

118 preferentially stabilized by association with soil minerals (Gu et al., 1994; Kaiser et al.,

119 1997; Spielvogel et al., 2008). SOM functional fractions such as particulate organic

120 matter (POM) and the chemical structure of SOM, determined by ¹³C cross polarization

121 magic angle spinning nuclear magnetic resonance (CPMAS NMR) spectroscopy, have

also been found to be involved in preferential stabilization in soils with different

123 mineralogy (Schöning et al., 2005; Spielvogel et al., 2008). All of the above evidence

124 indicates that studies investigating the influence of soil parent material and soil

125 mineralogy on the preservation of soil organic matter should consider the possible bias

126 induced by the presence of organic compounds of different origin (e.g. Plante et al.,

127 2010). At a practical level, studies comparing different soils should consider the nature

128 of the SOM and should ideally compare soils with similar organic inputs.

129 Studying the relationship between lithology and SOC stabilization is also of interest

130 because geological and lithological data is generally more accessible and accurate than

131 soil-related data (Araujo et al., 2017).

132 However, studies of the role of soil mineralogy on SOM protection under temperate

133 climate remain scarce. This particularly applies to clay minerals, which may have

134 different characteristics (e.g. cation exchange capacity and specific surface area) and

thus different capacities to protect SOC (Barré et al., 2014; Fernández-Ugalde et al.,

136 2016).

137 The following main difficulties in interpreting the results of these studies have been 138 reported (Barré et al., 2014): i) data on soils rich in 1:1 clay minerals mainly correspond 139 to tropical zones (Balbino et al., 2004), whereas data on soils dominated by 2:1 clay 140 mainly correspond to temperate zones (thus climate, pH and oxide content covary with 141 phyllosilicate mineralogy; Chenu et al., 2009); ii) phyllosilicate mineralogy is seldom 142 determined in soil studies; and iii) the presence of oxides associated with phyllosilicates 143 can also minimize or even conceal the relationship between phyllosilicate mineralogy 144 and SOC protection (Saidy et al., 2013).

145 On the basis of existing knowledge about the role of minerals on SOC storage, we 146 studied SOC storage in relation to mineralogy in three acidic temperate Atlantic forest 147 soils, which were developed on volcanic rock (basalt and trachyte) and sandstone. The 148 three soils are located in the same geographical region, under similar climatic conditions 149 and with similar vegetation, thus ensuring valid comparisons by eliminating drawback 150 (i) described by Barré et al. (2014). In addition, we determined the mineral composition, 151 including the possible interactions between clay minerals, SRO and Fe and Al oxides, to 152 address drawbacks (ii) and (iii) described by the same authors. The total amounts of 153 organic C and of various functional OC groups were also investigated at two depths (0-154 5 and 5-20 cm) to elucidate how they are related to soil parent material, soil mineralogy 155 and SOM stabilization.

156 Our study aimed to address the following: (1) the possible influence of parent material

157 on SOC stocks and concentration in temperate forest in the Basque Country; (2) the

158 involvement of mineralogical differences (primary and secondary organo-mineral

associations) in any such relationship; (3) the possible influence of mineralogy on the

160 chemistry of SOC in temperate forest in the Basque Country.

161 Our hypotheses were as follows: (1) soils developed on volcanic rocks are richer in

162 minerals that favour organo-mineral interactions (swelling clays, SRO minerals,

163 metallic oxy-hydroxides); (2) this will result in higher SOC concentrations and stocks in

164 soils developed on volcanic rocks than in soils developed on sandstone; and (3) this will

165 influence the chemistry of SOC in the different soils.

166 **2. Materials and methods**

167 2.1. Study area and collection of soil samples

168 An observational study was conducted in mature, second rotation *Pinus radiata* D. Don

169 forest plantations in two areas of the Basque Country (northern Spain): (i) Mount Oiz, a

170 sandstone massif (30T, 534075, 4783284), and (ii) Mount Karakate, a volcanic massif

171 (with basalt and trachyte as dominant rocks) (30T, 548126, 4781437), separated by a

172 distance of less than 25 km (Figure 1). The climate in both areas is temperate and

173 humid, with annual mean temperature and annual mean precipitation of around 9 °C and

174 1200 mm, respectively, and 190 rainy days. Precipitation is distributed rather evenly

throughout the year, with maximum levels in autumn and spring. The sites are thus

176 representative of Atlantic temperate conditions.

177 Three forest stands developed over each type of parent material (sandstone, basalt and

trachyte) (total, 9 stands) were selected for study (Figure 1). The mean tree height

179 (based on LIDAR data) was 26.9 m \pm 0.4 standard error (SE) in the stands developed on

180 sandstone, 24.7 m \pm 2.6 SE in the stands over basaltic rock and 27.8 m \pm 2.5 SE in the

181 stands on trachytic rock. Tree height was used as a proxy for forest development stage

182 (Moreno et al., 2016) and did not differ in relation to bedrock. Five representative

183 sampling points were selected in each stand. Soil samples were collected from the upper

184 20 cm of the soil profile obtained by digging a 20x20x20 cm pit (after removal of forest

185 floor material). The samples were then divided into 0-5 cm and 5-20 cm layers and

186 combined to produce one composite sample per stand and layer. The soil samples were

187 air-dried and sieved to < 2 mm. An additional soil core was obtained in each pit for

188 determination of soil bulk density.

189 2.2. Soil chemical, physical and mineralogical data

190 Soil pH values were measured in deionised water and in 1 M KCl at a soil/ solution 191 ratio of 1:2.5. The pH_{NaF} was measured in a suspension of 1g soil mixed with 50 ml of 192 1M NaF solution after stirring for 2 min. The exchangeable cations and cation exchange 193 capacity (CEC) were determined in unbuffered solutions. The exchangeable aluminium 194 and cations were extracted with 0.1 M BaCl₂ (Gillman, 1979), and the exchangeable 195 aluminium was determined by titration with 0.1 M NaOH to pH 7.2. The exchangeable 196 Ca²⁺, Mg²⁺, Na⁺ and K⁺ in the extracts were determined by atomic absorption 197 spectroscopy. The effective cation exchange capacity (eCEC) was calculated as the sum 198 of base cations and Al. Air-dried soil samples were ground to pass a 250 µm sieve for 199 total carbon and nitrogen analysis in a LECO TruSpec autoanalyzer (LECO 200 Corporation, MI, USA).

201 After removal of the organic matter by oxidation with H₂O₂ and dispersion of the

sample with Na-hexametaphosphate, the soil texture was determined by laser

203 diffractometry (Mastersizer 2000 diffractometer, Malvern Instruments, Worcestershire,

204 UK).

205	Mineralogical analysis of air-dried sieved soil samples (< 2 mm) (using Cu radiation
206	from 2° to 80° 2 θ) and of the clay fraction (< 2 μ m) on oriented slides was conducted,
207	by powder X-ray diffraction (XRD), to identify primary and secondary crystalline
208	minerals in each soil sample.
209	Air-dried sieved soil samples (< 2 mm) were washed with demineralized H_2O and
210	disaggregated by attrition and ultrasound for 4 cycles each of 30 min. Clay-size particles
211	were isolated by repeated mixing with weakly basified demineralized water until $pH = 8$
212	(NH4OH) and centrifuging at 3500 rpm (in Kubota K8410 centrifuge). After each
213	centrifugation cycle, the supernatant liquid (devoid of particles) was removed until
214	slight turbidity was observed. The suspensions were distributed in 100 x 30 mm glass
215	tubes and centrifuged at 1000 rpm for 100 sec. The upper 4 cm of the supernatant
216	liquid, which only contains particles <2 microns (Stokes law), was then extracted. Clays
217	were oriented and mounted on glass slides with the following standard treatments: Mg
218	saturation, Mg saturation/glycerol solvation, K saturation, and heat treatment of K-
219	saturated samples at 300°C. X-ray analysis was conducted using a diffractometer
220	(Philips PW1710) with Bragg-Brentano geometry (θ ,2 θ) and Ni-filtered CuK α
221	radiation. Diffraction data on powdered whole samples were analysed using
222	PANalytical X'Pert Highscore Plus software, coupled to the PDF2 database of the
223	International Centre for Diffraction Data (ICDD). Clay minerals were identified
224	according to the position of the (001) series of basal reflections on XRD patterns
225	collected under indicated treatments of oriented aggregates, in the range 1 to $30^{\circ} 2\theta$
226	(Moore and Reynolds, 1997).
227	2.3 Selective extraction of soil for Fe and Al
228	The Al, Fe and Si compounds were selectively dissolved in three chemical reactants to
229	yield the non-crystalline components. "Active" Al, Fe and Si compounds (Alox, Feox,

230 Si_{ox}) were extracted in darkness after shaking the soil with an ammonium oxalate 231 solution buffered at pH 3 (Schwertmann, 1964). For extraction of Fe and Al complexed 232 with organic matter compounds (Fe_p, Al_p), the soil was shaken with a sodium 233 pyrophosphate solution (Blakemore et al., 1987). "Free" Fe and Al compounds (Fed, 234 Al_d) were extracted using a citrate dithionite solution (Mehra and Jackson, 1960). 235 Concentrations of Fe, Al and Si in supernatants were determined by absorption 236 spectrophotometry (Perkin Elmer 2380, Norwalk, CT, USA). Aluminium and iron 237 associated with inorganic short-range order (SRO) material was calculated as oxalate-238 extractable Al and Fe minus pyrophosphate extractable Al and Fe (Grand and 239 Lavkulich, 2011). 240 2.4. SOM fractionation 241 2.4.1. Particulate organic matter (POM) and mineral-bound organic matter (MOM) 242 The C fractions in POM (>53 μ m) and MOM (<53 μ m) were determined by applying

243 ultrasound energy to a soil-water suspension. A suspension of 5 g soil (oven-dry 244 equivalent) and 25 mL of distilled water was held in a beaker in an ice bath and 245 sonicated (total input 450 J mL⁻¹) with a digital sonifier (VCX130 ultrasonic processor, 246 max. output, 58.5W) equipped with a flat tipped probe of 13 mm diameter, operated at 247 20 kHz.

The POM was retained on a 53 μ m sieve and the mixture of silt+clay fraction (< 53 μ m) was centrifuged at 6000 xg for 30 min and filtered (cellulose acetate filter, 0.45 μ m) in a Millipore Sterifil filtration system. Fractions were dried at 50°C for 48 hours and weighed to determine the recovery. The total C in both fractions was determined using the same elemental analyzer.

253 2.4.2. Solid-state NMR spectroscopy

Cross-polarization magic-angle spinning (CPMAS) ¹³C NMR spectra of all soil samples 254 255 were obtained. Samples were previously treated with HF to remove paramagnetic 256 compounds (Gonçalves et al., 2003), as the high concentration of paramagnetic 257 compounds in these samples does not allow acquisition of NMR spectra without prior treatment with HF.¹³C solid state NMR spectra were acquired in a spectrometer (Bruker 258 Avance DSX 300) equipped with a CP/MAS probe (Bruker BL 7mm). The ¹³C 259 260 frequency was 75.49 MHz and the 1H frequency, 300.13 MHz. CP/MAS spectra were 261 recorded with high power decoupling. A contact time of 1 ms, a 90 degree pulse (^{13}C) 262 of 7.5 microseconds, and a scan delay of 2 seconds proved optimal for acquisition of 263 maximum signal and relaxed spectra. An FID of 4004 points with a spectral width of 50 264 KHz was measured, and exponential apodization of 30 Hz was applied prior to Fourier 265 transformation.

The spectra were subdivided into four regions: (i) alkyl-C (0–45 ppm); (ii) O-alkyl-C (45–110 ppm); (iii) aromatic-C (110–162 ppm), and (iv) carbonyl-C (162–190 ppm). Integration of the peaks within each of the chemical shift regions enabled estimation of the relative C contents expressed as percentages of the total area.

270 2.6. Data analysis

As this study was observational, no randomization could be carried out; n=1 (including three pseudoreplicates) per parent material. Thus, mean values and 95% confidence intervals (CI) for the mean were used to reflect the degree of uncertainty in the measurements (Crawley, 2012). Depths (0–5 and 5–20 cm) were analyzed separately. Confidence intervals (95%) for the weighted (0-20 cm) means are also shown as supplementary data.

277

278 **3. Results**

279 3.1. Soil chemical, physical and mineralogical characterization

280 3.1.1. Chemical and physical properties

All soil samples were acidic, and the values of $pH_{H_{2}O}$ (< 5) and pH_{KCl} (< 4.5) did not differ in relation to parent material (Table 1). However, in the 0-5 cm soil layer, the pH_{NaF} was between 0.4 and 2.7 higher in soils developed on trachyte than in those

- 284 developed on sandstone (Table 1). The pH_{NaF} was similar in all soils in the 5-20 cm
- 285 layer (Supplementary Data 1). The three types of soil showed similar values in either
- the eCEC or exchangeable bases (Table 1). However, in the 0-5 cm layer the mean
- eCEC at pH7 was between 14.2 and 36.9 cmol+kg⁻¹ lower in sandstone-derived soils
- than trachyte-derived soils and between 9.9 and 26.0 cmol+kg⁻¹ lower in sandstone-
- derived soils than basalt-derived soils (Table 1). Soil acidity decreased with depth, as
- 290 did eCEC and exchangeable bases.
- Soil bulk density in the upper 20 cm was 1.07 g cm⁻³ \pm 0.04 CI in the sandstone-derived
- soil, 0.62 g cm⁻³ \pm 0.03 CI in basalt-derived soil and 0.64 g cm⁻³ \pm 0.02 CI in the
- trachytic-derived soil. The mean soil bulk density in sandstone-derived soils was
- between 0.38 and 0.52 g cm⁻³ higher than in basalt derived-soils and between 0.36 and
- $295 \quad 0.48 \text{ g cm}^{-3}$ than in trachyte-derived soils. In the sandstone-derived soil, the sand
- 296 particle content was higher and silt particles less abundant than in the soils derived from
- 297 basalt and trachyte. There was no difference between the soils in the clay particle
- 298 content, determined by laser diffractometry, neither in the upper 0-5 cm layer nor in the
- 299 5-20 cm layer (Table 1 and Supplementary Data 1).

300 3.1.2. Mineralogy

- 301 Quartz was the dominant primary mineral in the sandstone-derived soil (almost 75 % of
- 302 the mineral fraction) and the mean was 50-60% higher than in the soils derived from
- 303 basalt and trachyte (Table 2 and Supplementary Data 1). By contrast, highly

304 weatherable minerals such as plagioclases were detected in the volcanic soils. Soils

305 developed on basalt differed from those developed on trachyte with respect to the

306 haematite content (Table 2 and Supplementary Data 1).

307 The proportion of phyllosilicates was 2 times lower in the sandstone-derived soil than in

308 the soils derived from basalt and trachyte (Table 2 and Supplementary Data 1). In the

309 sandstone-derived soil, the clay-size fraction showed a mixed mineralogy of

310 phyllosilicates, dominated by illite, with lower proportions of hydroxyvermiculite,

311 hydrobiotite and kaolinite, and a total absence of interstratified chlorite/vermiculite

312 (Table 2 and Supplementary Data 2). By contrast, in the soils developed on volcanic

313 rocks, the major crystalline mineral identified in the clay-size fraction was interstratified

314 chlorite/vermiculite. No major differences were observed between basalt and trachyte-

315 derived soils in the phyllosilicate mineralogy of this fraction (Table 2 and

316 Supplementary Data 2).

324

317 The selective dissolution of aluminium showed that Al_{ox}, Al_d and Al_p contents were

318 always lower in the soil developed on sandstone than in soils developed on basalt and

319 trachyte (Table 3 and Supplementary Data 1), but with no differences between the two

320 soil depths (Table 3). SRO inorganic phases of aluminium (Al_{SRO}) were not detected in

321 the soil developed on sandstone (Table 3).

322 Selective extraction of Fe showed that the concentration of oxalate-extractable iron

323 (Fe_{ox}) was between 0.7 and 10.04 g kg⁻¹ lower in the sandstone-derived soil than in the

volcanic soils (Supplementary Data 1); within these, the concentration was higher in the

325 basalt-derived soil than in the trachyte-derived soil. The concentration of Fe_d was lowest

in the sandstone-derived soil, and there was no difference between basalt- and trachyte-

327 derived soils. The three types of soil were similar in pyrophosphate extractable iron

328 (Fe_p). The amount of oxalate-extractable silicon was higher in the basalt-derived soil

than in the sandstone- and trachyte-derived soils at both depths considered (Table 3 andSupplementary Data 1).

331 3.2 SOC and SOC fractions

332 3.2.1. Total SOC storage

333 The concentrations of SOC differed between soils (Table 1 and Supplementary Data 1).

In the upper horizon (0-5 cm), the SOC concentration was respectively 2 and 2.7 times

higher in the soils developed on basalt and from trachyte than in the soil developed on

336 sandstone. In the subsoil horizon (5-20 cm), the SOC was 4 times and 5 times higher in

the basalt and trachyte-derived soils than in the sandstone-derived soil. SOC

338 concentrations in the basalt- and trachyte-derived soils were similar (Table 1 and

339 Supplementary Data 1). As a consequence, and despite the differences in bulk density,

340 the SOC stocks were lower in the sandstone-derived soil than in the trachyte- derived

341 soils (Table 1 and Supplementary Data 1). The SOC stocks in the upper 20 cm

342 amounted to 61.5 Mg C ha⁻¹ \pm 4.2 CI in the sandstone-derived soil, 119.5 Mg C ha⁻¹ \pm

343 34.7 CI in the basalt-derived soil and 152.8 Mg C ha⁻¹ \pm 54.8 CI in the trachyte-derived

soil, and the differences in SOC stocks were greatest in the deeper soil layers (5-20 cm).

345 3.2.2. Particulate organic matter (POM) and mineral-bound organic matter (MOM)

346 The mean concentrations of POM-associated C were 29.9 ± 3.9 CI (mg C-POM g⁻¹ soil)

in the sandstone-derived soil, and 53.6 ± 19.3 CI and 92.0 ± 30.4 CI (mg C-POM g⁻¹

soil) in respectively the basalt- and trachyte-derived soils, in the upper 0-5 cm of the

soil profile (Table 1). The proportion of SOC in the POM fraction was similar in all

350 soils in the upper layer; however, in the lower layer (5-20 cm), the proportion of SOC in

351 the POM fraction was higher in the sandstone-derived soil than in either of the volcanic

352 soils.

- 353 Conversely, in the upper 0-5 cm of the soil profile, the concentration of MOM-
- associated C was 19.5 ± 3.1 CI (mg C-MOM g⁻¹ soil) in the sandstone-derived soil and

 55.8 ± 21.4 CI and 49.4 ± 5.6 CI (mg C-MOM g⁻¹ soil) in respectively the basalt- and

356 trachyte-derived soils (Table 1) The proportion of SOC in the MOM fraction for the

- 357 soils developed on basalt was higher compared to the soils developed on sandstone or
- trachyte in the 0-5 cm layer. In the lower part of the studied layer (5-20 cm), both
- 359 volcanic soils contained higher proportions of SOC bound to minerals: around 50% of
- 360 the SOC was present in the MOM fraction in the volcanic soils, while 30% of SOC was
- 361 present in MOM in soils derived from sandstone.
- 362 3.2.3. Chemical structure of organic matter
- 363 Integration of the major regions of ${}^{13}C$ NMR spectra indicated that the major C
- 364 components in all soils were dominated by O-alkyl C (37.6-40.3 % of total intensity)
- and alkyl C (34.3-42.3 % of total intensity) (Fig 2a). This indicates high proportions of

366 lipids, amino acids and polysaccharides as well as cellulose and hemicellulose. The

- 367 proportion of alkyl-C in the total C increased with depth, but no differences were
- 368 observed in the proportion of O-alkyl C in the total C. The trachyte-derived soil
- 369 contained lower proportions of aromatic-C in total C than the sandstone-derived soil,
- 370 while the values were intermediate in the basalt-derived soil (Fig 2a and Supplementary
- 371 Data 1). The proportion of aromatic-C in total C decreased with depth (Fig 2b and 2c).
- 372 **4. Discussion**

4.1. Influence of parent material on soil characteristics and SOC in temperate

374 forest in the Basque Country

375 In this study, we hypothesized that parent material would determine to some extent the

- amount and type of organic matter stored in the upper layer of forest soils in the Basque
- 377 Country. We based this hypothesis on previous observations that organo-mineral

interactions are dependent on soil mineralogy (Six et al., 2002b; Kleber et al., 2007;

379 Rasmussen et al., 2018), which in turn was expected to be different in the soils derived

380 from volcanic rocks or a clastic sedimentary rock such as sandstone.

381 Our observations confirmed that the underlying parent material strongly influenced the 382 soil matrix in the upper soil layer, which differed in texture and mineralogy at the three

383 study sites (Tables 1, 2 and 3).

384 Establishing an interaction between parent material, soil mineralogy and organic matter

385 stabilisation is, however, not straightforward (Wiesmeier et al., 2019). In fact, most

temperate soils contain a mixture of layer silicates, crystalline oxides and poorly

387 crystalline materials, and there is evidence showing that all of these materials play direct

388 or indirect roles in stabilizing organic matter (Sollins et al. 2009). However, although

389 soil clay mineralogy is known to influence the composition, turnover and amount of soil

390 organic matter (Barré et al., 2014; Singh et al., 2018; Wiesmeier et al., 2019), it is

391 difficult to attribute differences in SOM in different soil types to phyllosilicate

392 mineralogy alone, as many parameters can co-vary with clay mineralogy. In addition,

393 stabilization mechanisms are also influenced by land management, climate and

394 vegetation (Matus et al., 2014; Wiesmeier et al., 2019).

395 In our study, the nine stands considered were located in the same geographical area, and

396 were thus affected by similar climatic conditions, and on soils with similar pH

397 conditions and exchangeable cations. All were radiata pine stands established at least 50

398 years ago and they were at a similar stage of forest development, thus enabling us to

- assume similarities in the amount and type of OM inputs (litter, exudates...). Site-
- 400 related differences in SOC were therefore assumed to be the result of the interactions
- 401 between SOC and the soil matrix. In this context, total SOC concentration was greater
- 402 in the soils developed from volcanic materials (basalt and trachyte) than in the soil

403 developed on sandstone, although the values were intermediate in the soil derived from 404 basalt. Total C stocks in the upper 20 cm of the soil profile were also higher in the 405 volcanic soils, despite the lower bulk densities. In another recent study on forest soils, 406 Barré et al. (2017) also observed that parent material affected SOM stocks, which were 407 much lower in soils developed on loess deposits than in soils developed on chalk, 408 mudstone or grainstone. However, the parent material only marginally explained the 409 observed effect on topsoil organic C stocks due to the effect on soil clay abundance. In 410 the present study, the clay content of the soils was similar, and the observed differences 411 are probably explained by the nature of the minerals. 412 As summarized in the recent review by Wiesmeier et al. (2019), current knowledge of 413 the impact of clay mineralogy on SOC stabilization is based on limited and conflicting 414 data; however, it is usually accepted that the SOC stabilizing capacity is higher in soils 415 dominated by 2:1 clay minerals (mainly smectite) than in soils dominated by 1:1 clay 416 minerals (kaolinite) (Bruun et al., 2010). This is due to the higher reactivity of 2:1 clay 417 minerals, which would promote organo-mineral complexation (Kaiser and 418 Guggenberger, 2003; Kahle et al., 2004; Kleber et al., 2007; Feng et al., 2013), and/or to 419 the capacity of these minerals to promote aggregation (Fernández-Ugalde et al., 2013). 420 Clay mineralogy in the soils under study was dominated by 2:1 clay minerals (Table 2). 421 However, the soils showed contrasting clay mineral composition and concentrations of 422 Al and Fe forms (Table 3). Within the 2:1 clay minerals, those with a high cation 423 exchange capacity (CEC) and larger specific surface, such as montmorillonite and 424 vermiculite, have a higher SOM binding potential than clay minerals with a lower CEC 425 and smaller specific surface area, such as illite (Greenland, 1965). Mixed layer chlorite-426 vermiculite minerals, considered an intermediate step in the vermiculitization process, 427 are formed in weathered rock microsystems (Velde and Meunier, 2008), such as the

428 basalts and trachytes considered here. Illite can be formed during weathering cycles 429 (water/rock interactions) in temperate to cold climates on rocks containing abundant 430 aluminium and potassium. The most common sedimentary rocks (e.g. sandstone) favour 431 the formation of illite (Velde and Meunier, 2008). These phyllosilicate clay minerals 432 occur in association with oxides and organic matter forming microaggregates, which 433 ultimately contribute to stabilising SOC (Chenu and Plante, 2006); however, further 434 research is required in order to clarify the associations between phyllosilicates, oxides 435 and organic matter within microaggregates (Churchman, 2018). 436 Regarding Al and Fe forms, previous studies (Kaiser et al., 2002; Eusterhues et al., 437 2005) have suggested that iron oxides and SRO Al and Fe phases are the most important 438 surfaces for the formation of organo-mineral associations in acid forest Podzols. Grand 439 and Lakuvlich (2015) also reported that organically-bound Fe and Al were the major 440 factors determining the SOC in podzols of forests in coastal British Columbia. Without 441 considering clay mineralogy, Garrido and Matus (2012) observed that the Al_p, soil pH 442 and allophane content explained most of the variation in SOC. Bruun et al. (2010) also 443 concluded that the content of active Fe (hydr-) oxides determined SOC stabilization in 444 kaolinitic soils under moist semi-deciduous primary or secondary forests located in 445 lowland humid tropical areas. More recently, Rowley et al. (2018) reviewed the relative 446 role of polyvalent cations in SOC stabilization. These authors proposed a conceptual 447 model in which, in acidic environments, complexation of organic ligands by free Al^{3+} and Fe³⁺, as well as their mineral forms and the cementation of aggregates by Fe oxides 448 449 are probably determining factors in SOC stabilization. This model provides a good 450 explanation for the observed results, in which within three acid environments with 451 similar amounts and types of organic inputs, the amount and nature of organic matter

452 differed clearly in the two soils containing the highest concentrations of Fe and Al453 (Table 3).

Overall, our results (a difference between 32 and 150 Mg C ha⁻¹ between trachyte- and

sandstone-derived soils, and a difference between 19 and 97 Mg C ha⁻¹ between basalt-

and sandstone-derived soils) are generally consistent with those reported by Peña

457 Ramirez et al. (2009), who observed a strong relationship between Al content in

458 organo-mineral complexes (extractable with pyrophosphate) and the accumulation of

459 organic C in volcanic soils less than 8000 years old. However, as phyllosilicate

460 mineralogy covaried with metallic oxy-hydroxides, the influence of these factors was

461 not able to be determined in the present study.

462 **4.2. SOC chemistry differed slightly in the three soils under study**

463 The second hypothesis of this study was that the soil parent material would also

464 influence the SOC chemistry in the different soils, as a result of the various interactions

465 of the soil matrix and organic C inputs from the vegetation.

466 The first factor regulating such interactions would be the nature of these inputs. In this

467 respect, differences in the nature of SOC between the two depths considered in the three

468 soil types were slight, as alkyl-C and O-alkyl-C were the dominant forms of C in topsoil

469 (0-5 cm) and subsoil (5-20 cm) in the three soils, followed by aromatic-C and carbonyl-

470 C. This can be attributed to the type of vegetation in the study stands, dominated by

471 *Pinus radiata*. The forest floor of mature pine plantation of radiata pine in the same

472 climatic area showed a major signal in the O-alkyl region of carbons, which is

473 characteristic of the carbons in cellulose and hemicellulose (Gartzia-Bengoetxea et al.,

474 2009).

454

455

475 However, comparison of SOC composition between sites revealed an important

476 difference in the proportional contribution of aromatic-C forms to total SOC, which was

477 higher in the sandstone-derived soil than in both soils of volcanic origin. Organic matter 478 in direct contact with mineral surfaces is often depleted in aromatic structures, such as 479 lignin and phenolic components (Guggenberger et al., 1994; Kiem and Koegel-Knabner, 480 2003), and enriched in carboxyl and O-alkyl carbon, which are considered highly labile 481 structures (Mahieu et al., 1999; Spielvogel et al., 2008). This has been confirmed by e.g. 482 Kinyangi et al. (2006), who found that aromatic-C decreased significantly when organic 483 matter was bound to minerals, suggesting a transformation that oxidized C=C to COOH 484 forms. These results are consistent with the higher proportion of POM over SOC in the 485 sandstone-derived soil in the 5-20 cm layer, and they therefore support our hypothesis 486 of the soil matrix exerting an influence on the typology or organic forms in the studied 487 soils, in line with Kleber et al. (2010), who also observed the absence of highly 488 aromatic humic-like organic compounds from mineral-associated soil organic matter 489 pools. 490 In summary, the fact that the concentration of phyllosilicates was lower in the 491 sandstone-derived soil, together with a lower concentration of Fe and Al forms 492 potentially contributing to aggregate stabilization (Rasmussen et al., 2005), may explain 493 why SOM in this soil was richer in relatively fresh and only slightly decomposed plant 494 material than the SOM in the trachyte-derived soil (Schoning et al., 2005).

495

496 **5.** Conclusions

497 Although it seems that the consistency of the importance of parent material to SOC may 498 be weak at the global scale (Wiesmeier et al., 2019), the present findings support the 499 observation reported in a recent review study (Rasmussen et al., 2018). Thus, the 500 mineralogy of the air dried (<2mm) soil and clay fraction may be good predictors of 501 SOC content in acid temperate forest soils in the Basque Country, as there were no 502 differences in the commonly used proxies for SOC (i.e. pH and clay content) between 503 the three soils under study, even though SOC storage differed. Soils derived from basalt 504 and from trachyte, which were relatively rich in chlorite/vermiculite mixed layers and 505 short-range order aluminium and iron forms, contained between 35-102 g C kg⁻¹ and 506 between 42-140 g C kg⁻¹ more SOC than the sandstone-derived soils dominated by illite 507 and with no short-range order aluminium.

508 Dynamic models of SOC turnover in acid temperate forest soils could therefore be 509 improved by including proxies for clay mineralogy and Fe and Al (hydr-) oxide 510 contents. However, further research is required to determine the particular influence of 511 phyllosilicate mineralogy and metallic oxy-hydroxides.

512

513 At a larger scale, our findings support the view that soil parent material could be 514 included as a proxy in models aimed at estimating regional SOC stocks, as geological 515 maps are more widely available and often more accurate than soil maps.

516

517 Acknowledgements

518 This study was conducted as part of the CARBINIA project (SUM2006-00013-00-00) 519 funded by Spanish Institute for Agronomic Research (INIA) and the Department of 520 Agriculture of the Basque Government. We also acknowledge the input of the 521 CHARFOR project (RTA2012-00048-00-00). The authors are grateful for technical and 522 collaborative support provided by SGIker, X-Ray Service, Rocks and Minerals Unit 523 (UPV/EHU/ ERDF, EU). We are also grateful to Dr Javier Arostegi and Dr Javier 524 Sangüesa for assistance with soil mineralogical analysis and to Dr Juan Mari Alberdi for 525 assistance with the CPMAS NMR analysis.

527 **References**

- Araujo, M.A., Zinn, Y.L., Lal, R., 2017. Soil parent material, texture and oxide contents
 have little effect on soil organic carbon retention in tropical highlands. Geoderma
 300, 1-10.
- Balbino, L.C.,Bruand, A.,Cousin, I.,Brossard,M.,Quétin, P.,Grimaldi,M., 2004. Change
 in the hydraulic properties of a Brazilian clay Ferralsol on clearing for pasture.
 Geoderma 120, 297–307.
- Barré, P., Durand, H., Chenu, C., Meunier, P., Montagne, D., Castel, G., Billiou, D.,
 Soucémarianadin, L., Cécillon, L. 2017. Geological control of soil organic carbon
 and nitrogen stocks at the landscape scale. Geoderma 285: 50-56.
- Barré, P., Fernandez-Ugalde, O., Virto, I., Velde, B., Chenu, C., 2014. Impact of
 phyllosilicate mineralogy on carbon stabilization in soils: incomplete knowledge
 and exciting prospects. Geoderma 235-236, 382-395.
- 540 Berggren, D., Mulder, J., 1995. The role of organic matter in controlling aluminium
 541 solubility in acidic mineral soil horizons. Geochim. Cosmochim. Acta 59, 4167–
 542 4180.
- 543 Blakemore, L.C. 1978. Exchange complex dominated by amorphous material
 544 (ECDAM). In: The Andisol Proposal (ed. G.D. Smith), pp. 21 22. New
 545 Zealand Soil Bureau, Lower Hutt.
- 546 Bruun, T.B., Elberling, B.,Christensen, B.T., 2010. Lability of soil organic carbon in 547 tropical soils with different clay minerals. Soil Biol. Biochem. 42, 888–895.
- 548 Chenu, C., Plante, A.F. 2006. Clay-sized organo-mineral complexes in a cultivation
- 549 chronosequence: revisiting the concept of the "primary organo-mineral complex".
- 550 Eur. J. Soil Sci. 57, 596-607.
- 551 Chenu, C., Virto, I., Plante, A.F., Elsass, F., 2009. Clay-size organo-mineral complexes in

- 552 temperate soils: relative contributions of sorptive and physical protection. In: Laird,
- 553 D.A., Cervini-Silva, J. (Eds.), Carbon Stabilization by Clays in the Environment:
- 554 Process and Characterization Methods. Clay Mineral Society.
- 555 Christensen, B.T., 1996. Carbon in primary and secondary organomineral complexes.
- In: Carter, M.R., Stewart, B.A. (Eds.), Structure and Organic Matter Storage in
 Agricultural Soils. CRC Press, Boca Raton, Florida, pp. 97–165.
- 558 Churchman, G.J., 2018. Game changer in soil science functional role of clay minerals in
 559 soil. J. Plant Nutr. Soil Sci. 181, 99-103.
- 560 Crawley, M.J., 2013. The R book. 2nd edition. John Wiley & Sons Ltd. Chichester,
 561 West Sussex. United Kingdom.
- 562 Derrien, D. et al. 2016. Stocker du C dans les sols : Quels mécanismes, quelles
- 563 pratiques agricoles, quels indicateurs?. Étude et Gestion des Sols, 23: 193 -223.
- 564 Dijkstra, F., Fitzhugh, R., 2003. Aluminium solubility and mobility in relation to
 565 organic carbon in surface soils affected by six tree species of the northeastern
 566 United States. Geoderma 114, 33–47.
- Eusterhues, K., Rumpel, C., Kögel-Knabner, I., 2005. Organo-mineral associations in
 sandy acid forest soils: importance of specific surface area, iron oxides and
 micropores. Eur. J. Soil Sci. 56, 753-763.
- Feng, W., Plante, A.F., Six, J., 2013. Improving estimates of maximal organic carbon
 stabilization by fine soil particles. Biogeochemistry 112, 81-93.
- 572 Fernández-Ugalde, O., Barré, P., Hubert, F., Virto, I., Girardin, C., Ferrage, E., Caner,
- 573 L., Chenu, C., 2013. Clay mineralogy differs qualitatively in aggregate-size 574 classes: clay-mineral-based evidence for aggregate hierarchy in temperate soils.
- 575 European Journal of Soil Science, 2013 doi: 10.1111/ejss.12046

576	Fernández-Ugalde, O., Barré, P., Virto, I., Hubert, F., Billiou, D., Chenu, C. 2016. Does
577	phyllosilicate mineralogy explain organic matter stabilization in different particle-
578	size fractions in a 19-year C3/C4 chronosequence in a temperate Cambisol?
579	Geoderma 264: 171 – 178
580	Garrido, E., Matus, F., 2012. Are organo-mineral complexes and allophane content

- determinant factors for the carbon level in Chilean volcanic soils? Catena 92, 106–
 112.
- Gartzia-Bengoetxea, N., González-Arias, A., Martínez de Arano, I., 2009. Effects of
 tree species and clear-cut forestry on forest floor characteristics in adjacent
 temperate forests in northern Spain. Can. J. For. Res. 39: 1302-1312.
- Gillman, G.P., 1979. A proposed method for the measurement of exchange properties of
 highly weathered soils. Aust. J. Soil Res. 17, 129-139.
- 588 Gonçalves, C.N., Dalmolin, R. S.D., Dick, D.P., Knicker, H., Klamt, E., Kögel-
- 589 Knabner, I., 2003. The effect of 10% HF treatment on the resolution of CPMAS
- 590 13C NMR spectra and on the quality of organic matter in Ferralsols. Geoderma591 116, 373-392.
- Grand, S., Lavkulich, L.M., 2011. Depth distribution and predictors of soil organic
 carbon in podzols of a forested watershed in southwestern Canada. Soil Sci. 176,
 164-174.
- Grand, S., Lavkulich, L.M., 2015. Short-range order mineral phases control the
 distribution of important macronutrients in coarse-textured forest soils of coastal
 British Columbia, Canada. Plant and Soil 390, 77-93.
- Gray, J.M., Bishop, T.F.A., Wilson, B.R., 2015. Factors Controlling Soil Organic
 Carbon Stocks with Depth in Eastern Australia. Soil Sci. Soc. Am. J. 79, 1741.

- Greenland, D.J., 1965. Interactions between clays and organic compounds in soils. Part
 I. Mechanisms of interaction between clays and defined organic compounds. Soils
 and Fertilizers 28, 415–532.
- Gu, B., Schmitt, J., Chen, Z., Liang, L., McCarthy, J.F., 1994. Adsorption and
 desorption of natural organic matter on iron oxide: mechanisms and models. Env.
 Sci. Technol. 28, 38-46.
- Kahle, M., Kleber, M., Jahn, R. 2004. Retention of dissolved organic matter by
 phyllosilicate and soil clay fractions in relation to mineral properties. Organic
 Geochemistry, 35, 269–276.
- Kaiser, K., Guggenberger, G. 2003. Mineral surfaces and soil organic matter. European
 Journal of Soil Science, 54, 219–236.
- Kaiser, K., Eusterhues, K., Rumpel, C., Guggenberger, G., Kögel-Knabner, I., 2002.
 Stabilization of organic matter by soil minerals: investigations of density and
 particle-size fractions from two acid forest soils. J. Plant Nutr. Soil Sci. 165, 451459.
- 615 Kaiser, K., Guggenberger, G., Haumaier, L., Zech, W., 1997. Dissolved organic matter
- 616 sorption on subsoils and minerals studied by 13C-NMR and DRIFT spectroscopy.
- 617 Eur. J. Soil Sci. 48, 301-310.
- Kiem, R., Kögel-Knabner, I., 2003. Contribution of lignin and polysaccharides to the
 refractory carbon pool as studied in C-depleted arable soils. Soil Biol. Biochem.
 35, 101-118.
- Kinyangi, J., Solomon, D., Liang, B., Lerotic, M., Wirick, S., Lehmann, J., 2006.
 Nanoscale Biogeocomplexity of the Organomineral Assemblage in Soil:
 Application of STXM Microscopy and C 1s-NEXAFS Spectroscopy. Soil Sci.
 Soc. Am. J. 70, 1708–1718.

625	Kleber, M., Nico, P.S., Plante, A., Filley, T., Kramer, M., Swanston, C., Sollins, P.,
626	2010. Old and stable soil organic matter is not necessarily chemically recalcitrant:
627	implications for modelling concepts and temperature sensitivity. Glob. Change
628	Biol. 17, 1097-1107.

629 Kleber, M., Nico, P.S., Plante, A., Filley, T., Kramer, M., Swanston, C., Sollins, P.

2011. Old and stable soil organic matter is not necessarily chemically recalcitrant:

- 631 implications for modeling concepts and temperature sensitivity. Global Change632 Biology: 17, 1097–1107.
- Kleber, M., Sollins, P., Sutton, R. 2007. A conceptual model of organo-mineral
 interactions in soils: self-assembly of organic molecular fragments into zonal
 structures on mineral surfaces. Biogeochemistry 85:9–24.
- Lawrence, C., Harden, J.W., Maher, K., 2014. Modelling the influence of organic acids
 on soil weathering. Geochim. Cosmochim. Acta 139, 487–507.
- Lawrence, C.R., Harden, J.W., Xu, X., Schulz, M.S., Trumbore, S.E., 2015. Long-term
 controls on soil organic carbon with depth and time: A case study from the
 Cowlitz River Chronosequence, WA USA. Geoderma 247-248, 73-87.
- 641 Lehmann, J. and Kleber, M. 2015. The contentious nature of soil organic matter.
 642 Nature 528: 60–68.
- Mahieu, N., Powlson, D.S., Randall, E.W., 1999. Statistical analysis of published
 carbon-13 CPMAS NMR spectra of soil organic matter. Soil Sci. Soc. Am. J. 63,
 307–319.
- Marschner, B., Brodowski, S., Dreves, A., Gleixner, G., Gude, A., Grootes, P.M.,
 Hamer, U., Heim, A., Jandl, G., Ji, R., Kaiser, K., Kalbitz, K., Kramer, C.,
 Leinweber, P., Rethemeyer, J., Schäffer, A., Schmidt, M.W.I., Schwark,

- 649 L.,Wiesenberg, G.L.B., 2008. How relevant is recalcitrance for the stabilization of
 650 organic matter in soils? J. Plant Nutr. Soil Sci. 171, 91–110.
- Masiello, C.A., Chadwick, O.A., Southon, J., Torn, M., Harden, J.W., 2004. Weathering
 controls on mechanisms of carbon storage in grassland soils. Glob. Biogeochem.
 Cycles 18, 1-9.
- Matus, F., Amigo, X., Kristiansen, S.M., 2006. Aluminium stabilization controls
 organic carbon levels in Chilean volcanic soils. Geoderma 132, 158–168.
- 656 Matus, F., Rumpel, C., Neculman, R., Panichini, M., Mora, M.L., 2014. Soil carbon

storage and stabilisation in andic soils: A review. Catena 120, 102-110.

- Mehra, O.P., Jackson, M.L., 1960. Iron oxide removal from soils and clays by a
 dithionite-citrate system buffered with sodium bicarbonate. Clays Clay Miner., 7,
 317-327.
- Moore, D.M.,Reynolds, R.C., 1997. X-Ray Diffraction and the Identification and
 Analysis of Clay Minerals. Oxford University Press, Oxford, 378.
- Moreno, A., Neumann, M., Hasenauer, H., 2016. Forest structures across Europe.
 Geosci. Data J 4: 17-28.
- Peña-Ramírez, V.M., Vázquez-Selem, L., Siebe, C., 2009. Soil organic carbon stocks
 and forest productivity in volcanic ash soils of different age (1835-30,500 years
 B.P.) in Mexico. Geoderma 149, 224-234.
- Plante, A.F.; Virto, I.; Mahli, S.S. 2010. Pedogenic, mineralogical and land-use controls
 on organic carbon stabilization in two contrasting soils. Canadian Journal of Soil
 Science. 90 1: 15 26.
- Rasmussen, C., Torn, M.S., Southard, R.J., 2005. Mineral assemblage and aggregates
 control carbon dynamics in a California conifer forest. Soil Sci. Soc. Am. J. 69
 (6), 1711–1721.

- 674 Rasmussen, C., Heckman, K., Wieder, W.R., Keiluweit, M., Lawrence, C.R., Asefaw
- 675 Berhe, A., Blankinship, J.C., Crow, S.E., Druhan, J.L., Hicks Pries, C.E., Marin-
- 676 Spiotta, E., Plante, A.F., Schädel, C., Schimel, J.P., Sierra, C.A., Thompson, A.,
- Wagai, R., 2018. Beyond clay: towards an improved set of variables for predicting
 soil organic matter content. Biogeochemistry 137: 297-306.
- Rowley, M.C., Grand, S., Verrecchia, E.P., 2018. Calcium-mediated stabilisation of soil
 organic carbon. Biogeochemistry 137: 27-49.
- Saidy, A.R., Smernik, R.J., Baldock, J.A., Kaiser, K, Sanderman, J. 2013. The sorption
 of organic carbon onto differing clay minerals in the presence and absence of
 hydrous iron oxide. Geoderma 209-210, 15-21
- Schöning, I., Morgenroth, G., Kögel-Knabner, I., 2005. O/N-alkyl and alkyl C are
 stabilised in fine particle size fractions of forest soils. Biogeochemistry 73, 475497.
- 687 Schwertmann, U., 1964. Differenzierung der Eisenoxide des Bodens durch Extraktion
 688 mit Ammoniumoxalat-Lösung. Z. Pflanzenernähr. Düng. Bodenkd. 105, 194–
 689 202.
- Silva, Y.J.A.B., Nascimento, C.W.A., Biondi, C.M., Straaten, P., Souza Jr., V.S.,
 Ferreira, T.O., 2016. Weathering rates and carbon storage along a climosequence
 of soils developed from contrasting granites in northeast Brazil. Geoderma 284: 1–
 12.
- Singh, M., Sarkar, B., Sarkar, S., Churchman, J., Bolan, N., Mandal, S., Menon, M.,
 Purakayastha, T.J., Beerling, D.J. 2018. Stabilization of soil organic carbon as
 influenced by clay mineralogy. Advances in Agronomy, 148, 33-84.
- 697 Six, J., Conant, R.T., Paul, E.A., Paustian, K., 2002a. Stabilization mechanisms of soil
 698 organic matter: implications for C-saturation of soils. Plant Soil 241,155–176.

- 699 Six, J., Feller, C., Denef, K., Ogle, S.M., de Moraes, J. C., Albertcht, A. 2002b. Soil
 700 organic matter, biota and aggregation in temperate and tropical soils Effects of
 701 no-tillage. Agronomie 22: 755–775.
- Sollins, P., Homann, P., Caldwell, B.A., 1996. Stabilization and destabilization of soil
 organic matter: mechanisms and controls. Geoderma 74, 65–105.
- 704 Sollins, P., Kramer, M.G., Swanston, C., Lajtha, K., Filley, T., Aufdenkampe, A.K.,
- Wagai, R., Bowden, R.D., 2009. Sequential density fractionation across soils of
 contrasting mineralogy: evidence for both microbial- and mineral-controlled soil
 organic matter stabilization. Biogeochemistry 96, 209–231.
- Spielvogel, S., Prietzel, J., Kögel-Knabner, I., 2008. Soil organic matter stabilization in
 acidic forest soils is preferential and soil-type specific. Eur. J. Soil Sci. 59, 674692.
- 711 Torn, M.S., Trumbore, S.E., Chadwick, O.A., Vitousek, P.M., Hendricks, D.M., 1997.

712 Mineral control of soil organic carbon storage and turnover. Nature 389, 170-173.

- 713 Velde, B., Meunier, A., 2008. The origin of clay minerals in soils and weathered rocks.
- 714 Springer-Verlag Berlin Heidelberg.
- Wagai, R., Mayer, L.M., Kitayama, K., Shirato, Y., 2011. Association of organic matter
 with iron and aluminium across a range of soils determined via selective
 dissolution techniques coupled with dissolved nitrogen analysis. Biogeochemistry
 112, 95–109.
- Wiesmeier, M, Urbanski, L., Hobley, E., Lang, B., con Lützow, M., Marin-Spiotta, E.,
 van Wesemael, B., Rabot, E., Lie, M., Garcia-Franco, N., Wollschläger, U.,
 Vogel, H-J., Kögel-Knabner,, I. 2019. Soil organic carbon storage as a key
 function of soils A review of drivers and indicators at various scales. Geoderma

- 723 333: 149–162.Wilson, M.J., Bain, D.C., Duthie, D.M.L., 1984. The soil clays of
- 724 Great Britain. II. Scotland. Clay Miner. 19, 709–735.

725

- 727 Figure caption
- Figure 1: Location of the 9 forest stands under study and details from the lithological
 map of the Basque Country (EVE, 1984) for mount Oiz (left) and mount Karakate
 (right).
- Figure 2: (a) Composition of C functional groups (% of the total signal intensity) in
 soils (0-20 cm; weighed means) developed on different parent materials, as
 characterized by CPMAS 13C NMR spectra: composition of C functional groups
 in (b) 0-5 cm and (c) 5-20 cm layers.



% of total SOC

Table 1: Main chemical and physical properties of the soil samples. SOC stock (Mg C ha⁻¹), SOC, soil organic carbon (g C kg⁻¹), POM, particulate organic matter (g C kg⁻¹), MOM, mineral-bound organic matter (g C kg⁻¹), texture (%), eCEC, effective cation exchange capacity (cmol+ kg⁻¹), CEC pH7, cation exchange capacity at pH =7 (cmol+ kg⁻¹) and exchangeable calcium, magnesium, sodium and potassium (cmol+ kg⁻¹). Mean values are shown with 95% confidence interval.

	Sandstone		Bas	salt	Trachyte	
	0-5 cm	5-20cm	0-5 cm	5-20cm	0-5 cm	5-20cm
Organic matter						
SOC stock	29.2 ± 1.0	32.4 ± 3.4	36.7 ± 9.7	82.8 ± 25.3	48.3 ± 12.8	104.6 ± 42.8
SOC	54.8 ± 2.7	20.3 ± 1.7	120 ± 29.1	90.4 ± 25.6	151.0 ± 38.0	109.2 ± 40.6
POM	29.9 ± 3.8	11.2 ± 6.4	53.6 ± 19.3	41.4 ± 10.8	92.0 ± 27.6	59.1 ± 30.3
MOM	19.5 ± 3.0	10.8 ± 2.7	55.8 ± 21.4	50.4 ± 29.3	49.4 ± 7.4	46.3 ± 5.6
Texture						
Clay	23 ± 3.9	27 ± 4.3	23 ± 1.3	24 ± 1.3	19 ± 0.3	20 ± 1.8
Silt	37 ± 6.0	34 ± 4.4	50 ± 4.0	53 ± 1.9	53 ± 2.3	55 ± 2.2
Sand	41 ± 10.3	40 ± 8.6	27 ± 4.7	23 ± 3.0	28 ± 2.6	25 ± 1.8

Chemical properties						
pH (H2O)	4.5 ± 0.4	4.8 ± 0.3	4.6 ± 0.2	4.7 ± 0.1	4.5 ± 0.0	4.6 ± 0.1
pH (KCl)	3.4 ± 0.2	3.7 ± 0.2	3.9 ± 0.1	3.9 ± 0.1	4.3 ± 0.7	3.8 ± 0.3
pH (NaF)	8.1 ± 0.5	9.3 ± 0.3	8.9 ± 0.2	9.6 ± 0.2	9.3 ± 1.1	9.7 ± 0.9
eCEC	10.2 ± 0.7	7.8 ± 0.3	14.8 ± 2.1	11.7 ± 1.8	16.1 ± 4.1	12.3 ± 3.7
CEC pH7	19.4 ± 2.4	13.0 ± 1.1	37.3 ± 5.6	33.1 ± 5.7	44.9 ± 8.9	36.5 ± 7.4
Ca	3.3 ± 1.4	1.7 ± 0.8	4.5 ± 1.5	2.3 ± 0.9	5.8 ± 2.1	2.6 ± 1.3
Mg	0.7 ± 0.2	0.3 ± 0.1	1.1 ± 0.4	0.7 ± 0.3	1.3 ± 0.7	0.7 ± 0.4
К	0.02 ± 0.01	0.01 ± 0.0	0.02 ± 0.0	0.01 ± 0.0	0.02 ± 0.01	0.01 ± 0.0
Na	0.26 ± 0.1	0.14 ± 0.05	0.19 ± 0.01	0.17 ± 0.02	0.23 ± 0.07	0.18 ± 0.05

	Sandstone		Basa	alt	Trachyte	
	0-5 cm	5-20cm	0-5 cm	5-20cm	0-5 cm	5-20cm
Air dried soil < 2 mm						
Quartz	72.0 ± 8.5	75.3 ± 2.4	9.3 ± 0.7	19.0 ± 7.4	22.0 ± 9.3	19.3 ± 2.8
Phyllosilicates	28.0 ± 8.5	24.7 ± 2.4	59.3 ± 5.2	48.0 ± 5.2	57.7 ± 13.5	48.7 ± 9.2
Potassium feldspar	0.0	0.0	2.3 ± 4.6	6.4 ± 19	5.7 ± 5.6	13.3 ± 4.6
Plagioclase	0.0	0.0	16.0 ± 4.9	19.0 ± 2.0	14.7 ± 1.7	18.7 ± 6.9
Haematite	0.0	0.0	13.0 ± 4.1	7.7 ± 2.4	0.0	0.0
Phyllosilicates in the clay-						
size fraction (< 2 μm)						
Illite	60.7 ± 7.3	52.3 ± 1.7	21.0 ± 41.2	15.7 ± 30.7	12.7 ± 24.8	9.7 ± 18.9
Hydrobiotite	8.3 ± 4.7	10.7 ± 3.3	0.0	0.0	0.0	0.0
Hydroxyvermiculite	19.0 ± 3.0	24.7 ± 2.4	0.0	0.0	0.0	0.0
Kaolinite	12.0 ± 2.0	12.7 ± 2.4	0.0	0.0	0.0	0.0
	l					

Table 2: Mineralogical analysis (XRD) of the soil samples and clay fraction in %. Mean values are shown with 95% confidence interval.

Chlorite/vermiculite	0.0	0.0	79.0 ± 41.2	84.3 ± 30.7	87.3 ± 24.8	90.3 ± 18.9

	Sandstone		Ba	salt	Trachyte	
	0-5 cm	5-20cm	0-5 cm	5-20cm	0-5 cm	5-20cm
Alox	1.39 ± 0.42	1.53 ± 0.42	9.24 ± 1.15	7.49 ± 2.2	8.07 ± 1.21	6.14 ± 2.8
Ald	2.55 ± 0.31	2.33 ± 0.45	8.33 ± 0.25	7.48 ± 2.18	7.37 ± 0.69	5.44 ± 2.93
Al_p	1.45 ± 0.39	1.65 ± 0.40	7.55 ± 0.46	6.39 ± 2.81	7.05 ± 1.41	4.01 ± 3.75
Alsro	nd	nd	1.69 ± 0.72	1.09 ± 0.61	1.01 ± 0.28	2.13 ± 1.05
Feox	3.47 ± 1.61	3.66 ± 1.57	11.79 ± 3.01	11.06 ± 2.03	7.31 ± 2.21	8.48 ± 3.56
Fed	10.7 ± 3.68	13.9 ± 6.87	25.8 ± 1.66	27.0 ± 3.17	19.0 ± 2.12	24.4 ± 1.15
Fe _p	3.85 ± 1.24	4.64 ± 1.72	8.12 ± 0.87	5.86 ± 2.22	5.78 ± 1.39	4.25 ± 4.68
Fesro	nd	nd	3.67 ± 2.15	5.19 ± 3.94	1.53 ± 0.82	4.23 ± 1.20
Si _{ox}	0.46 ± 0.08	0.53 ± 0.07	1.10 ± 0.29	1.02 ± 0.13	0.54 ± 0.11	0.85 ± 0.07

Table 3: Chemical selective dissolution of Al, Fe and Si in g kg⁻¹. Al_{SRO} and Fe_{SRO} were calculated as oxalate-extractable Al and Fe minus pyrophosphate Al and Fe. Mean values are shown with 95% confidence interval.