

Biosorption of Cadmium from Water using Surfactant Impregnated Eggshell

Elvis Fosso-Kankeu*, Gerhard Steenekamp, Enoch Akinpelu and Frans Waanders

Abstract—Cadmium (Cd) is regarded as one of the most toxic elements and it causes water pollution when it exceeds the permissible limit. It is currently present in the Mooi River above these permitted levels, and must therefore be removed using effective methods. In this study, egg shell adsorbent was prepared through physico-chemical treatment including impregnation with an anionic surfactant, namely sodium dodecyl sulphate; the resulting adsorbents were characterized using FTIR spectroscopy. The prepared adsorbents were then used for the removal of cadmium from solution. The results showed that there were additional binding groups on the egg shell following impregnation, this contributed to increase the adsorption capacity of the adsorbent. The adsorption of cadmium by the pristine egg shell membrane (ESM) and the impregnated egg shell membrane (SDS-ESM) fitted the pseudo second order kinetic model which allowed to predict adsorption capacities of 16.56 mg/g and 20.4 mg/g, respectively. In can therefore be concluded that the impregnation of the ESM is a suitable approach to improve its adsorption capacity.

Keywords— Water pollution, cadmium, adsorption, eggshell, surfactant impregnation

I. INTRODUCTION

A significant amount of pressure is currently exerted on arid and semi-arid countries due to a lack of safe, usable water resources [1]. Water is an essential necessity for all humans, vegetation and animals throughout the world [2]. However, water scarcity and pollution have become worldwide concerns due to an increase in human population, industrial and agricultural production as well as domestic activities [3].

Water scarcity is expanding in too many regions throughout the world due to the impact of water usage on available resources and the existing water quality [4-12]. Water scarcity poses a threat on biodiversity and human welfare [5].

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Humans, animals and vegetation can be harmed when exposed to an excess amount of heavy metals, such as cadmium, chromium arsenic, mercury, zinc and lead [13]. Serious and painful symptoms can be experienced when exposed to a critical amount of heavy metals. These symptoms include fatigue, high blood pressure, speech disorders, vascular occlusion, sleep disabilities, poor concentration, increased allergic reactions, autoimmune diseases and even memory loss [1, 14].

The contamination of drinking water by means of heavy metals is a significant problem that can be seen in South Africa as well as many other countries throughout the world [4, 15]. South Africa has a substantial amount of semi-arid regions and experiences an uneven distribution of seasonal rainfall. Floods, droughts and high rates of evaporation are also serious problems experienced in South Africa [16]. Therefore, South Africa faces water scarcity on a regular basis and thus, the quality and quantity of every possible water resource must be maintained.

Conventional methods used for the removal of heavy metals from surface water, consist of different physico-chemical processes [17-22]. These methods include mechanical screening, hydrodynamic classification, gravity concentration, filtration, reverse osmosis, scrubbing, chemical precipitation, coagulation and flocculation, ion exchange, electro dialysis as well as adsorption [18, 23, 24].

However, toxic heavy metals can also be removed by the use of agricultural by-products as adsorbents through the unconventional biosorption process [4,25-29]. Organic material such as eggshell membranes, hazelnut shell, orange peels, pecan shells, maize cob and coconut shell can be used as adsorbents in the biosorption process [18, 30]. These adsorbents can be pretreated with chemical or physical methods to improve their adsorption capacities.

The scope of this study is to prepare an eggshell-based biosorbent for the removal of cadmium from the surface water of the Mooi River. Eggshell membranes have a very high surface area with functional groups such as hydroxyl, thiol, carboxyl, amino and amide. These functional groups have the ability to bind effectively to heavy metal ions [31]. Therefore, these membranes have the ability to be utilized as biosorbents for the removal of toxic heavy metal ions [32].

The adsorption capacity of natural eggshell membranes as well as eggshell membranes modified with surfactant impregnation will be investigated. Sodium dodecyl sulphate (SDS) is an anionic surfactant that will be used for modification to increase the membranes' affinity to the adsorption of cations, such as heavy metal ions [33, 34].

The characteristics of the two biosorbents will be investigated to establish the behavior and adsorption mechanism of each biosorbent. This investigation will then also be used to make educated conclusions regarding the adsorption process. Lastly, the kinetics and isotherms of each biosorbent will be utilized to quantify the method of biosorption. Thus, the purpose of this study is to analyze and evaluate the adsorption mechanism of cadmium (II) ions onto natural and SDS impregnated eggshell membranes.

II. EXPERIMENTAL

A. Materials

Egg shell membranes which are abundant biological waste materials in the study area were used as biosorbents. The eggshells were collected at a hatchery near Potchefstroom, South Africa. Two different pathways were followed when using the biosorption materials during the experimental procedure, namely the use of non-treated eggshell membranes and surfactant impregnated eggshell membranes.

Sodium dodecyl sulphate (SDS) was used as surfactant to impregnate the eggshell membranes. It was bought from ACE chemicals. By impregnating the membranes, the structure of the membranes is altered, which will cause an increase in its adsorption efficiency.

Cadmium chloride ($\text{CdCl}_2 \cdot 2, 5 \text{H}_2\text{O}$) was used as the initial salt for cadmium adsorbate and it was supplied by the School of Chemical- and Mineral Engineering of the North-West University. Deionized water was used for the preparation of the heavy metal solutions.

B. Preparation of eggshell membranes

To prepare the adsorbent, the membranes were then separated from the eggshells, then rinsed with tap water to remove any unwanted materials. Thereafter, the eggshell membranes were dried in an oven at 105°C for 45 minutes. After drying has taken place, eggshell membranes were grinded in the mortar. The grinded membranes were then pulverized by using a mechanical pulverizer. The membranes were pulverized for 5 minutes at 200 rpm and then for 5 minutes at 250 rpm. The powdered materials were then sieved to obtain particles at the desired uniform particle size, namely $100 \mu\text{m}$. Finally, the materials were stored in the refrigerator at 4°C .

C. Surfactant impregnation of eggshell membranes

Firstly, 10 grams of sodium dodecyl sulphate (SDS) was dissolved in 1 liter of deionized water. Then, 10 grams of the biosorbent was added to the SDS solution. The solution was then shaken at 160 rpm and 50°C for 24 hours by using a shaking incubator. After shaking has taken place, the solution was filtered by using a Buchner funnel (vacuum filter). Finally, the residue on the filter paper was oven dried at 60°C for 12 hours. The surfactant impregnated adsorbent was then used for the removal of cadmium from aqueous solutions.

D. Characteristics of biosorbents

The non-impregnated and impregnated biosorbents were characterise using Fourier Transform Infrared Spectroscopy (FTIR). Characterization was done before and after biosorption has taken place.

FTIR analyses were carried out to identify the functional groups of the non-impregnated and impregnated biosorbents. Thus, the purpose of these FTIR analyses was to establish the degree of surface activation and to determine which functional groups were available to bind to the cadmium ions.

E. Adsorption studies

To determine the equilibrium concentration of each cadmium solution, an Inductively Coupled Plasma Optical Emission Spectroscopy (ICP-OES) was used.

The effects of adsorbent dosage (0.01–0.5 g), initial metal concentration (20–100 mg/L) and contact time (15–300 min) were tested, keeping constant the following parameters: temperature at 25°C , adsorbent particle size at $100 \mu\text{m}$, rotation speed at 150 rpm and a cadmium solution volume of 50 mL. The adsorption of cadmium from Mooi River surface water was done at a temperature of 25°C , rotation speed of 150 rpm and aliquots of 50 mL by using the optimal adsorbent dosages of each biosorbent.

F. Isotherm and kinetic models

The cadmium adsorption of natural eggshell membranes (ESM) and SDS impregnated eggshell membranes (SDS-ESM) were investigated by using the Langmuir and Freundlich isotherm models [35].

The Langmuir isotherm assumes a monolayer surface coverage and homogeneous adsorption [36–38]. The linear form of the Langmuir model is given as follow:

$$\frac{C_e}{q_e} = \frac{1}{k \cdot q_m} + \frac{C_e}{q_m} \quad (1)$$

The Freundlich model is not limited to monolayer adsorption and assumes a heterogeneous system [38]. The linear form of the Freundlich model is given as follow:

$$\log q_e = \log k_f + \frac{1}{n} \cdot \log C_e \quad (2)$$

In the two isotherm models given, C_e is the cadmium concentration at equilibrium in (mg/L), q_e is the amount of cadmium adsorbed per unit mass of adsorbent at equilibrium in (mg/g), q_m is a Langmuir constant associated with the adsorption capacity in (mg/g), k is a Langmuir constant related to energy released during adsorption in (L/mg), k_f is the Freundlich adsorption capacity parameter in (mg/g) and n indicates the deviation of adsorption from linearity.

The pseudo-kinetic models were also used to identify the adsorption affinity of the different biosorbents [35, 39]. The pseudo-first-order kinetic model is as follow:

$$\ln(q_e - q_t) = q_e - \frac{k_1}{2.303} t \quad (3)$$

The pseudo-second-order kinetic model is as follow [40, 41]:

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} t \quad (4)$$

In the two kinetic models given, q_e is the amount of cadmium adsorbed per unit mass of adsorbent at equilibrium in (mg/g), q_t is the amount of cadmium adsorbed at time t in (mg/g), k_1 is the first-order rate constant in (min^{-1}), k_2 is the second-order rate

constant in (g/mg/min) and t is the time in (min).

III. RESULTS AND DISCUSSION

3.1 FTIR analyses

3.1.1 Natural eggshell membranes

A peak could be observed between 3500 cm^{-1} and 3100 cm^{-1} , indicating a vibrational stretch of N-H groups and the presence of amines and amides.

Between 3300 cm^{-1} and 3200 cm^{-1} a peak emerged in the spectral region, implying symmetric and asymmetric vibrational stretch of C-H functional groups.

Between 3000 cm^{-1} and 2800 cm^{-1} a peak was observed in the spectral region, which may represent a symmetric and asymmetric stretch of H-C-H groups.

Between 1750 cm^{-1} and 1625 cm^{-1} there is a vibrational stretch of the C=O functional groups as depicted by the peak formed in this spectral region. This can be related to aldehyde and ketone.

The presence of peaks between 1300 cm^{-1} and 1000 cm^{-1} indicates the presence of a vibrational stretch of the C-O functional groups.

3.1.2 SDS impregnated eggshell membranes

A peak was also observed between 3500 cm^{-1} and 3100 cm^{-1} , and could represent a vibrational stretch of N-H groups.

Between 3300 cm^{-1} and 3200 cm^{-1} a peak was present in the spectral region, and could be ascribed to a symmetric and asymmetric vibrational stretch of $\equiv\text{C-H}$ functional groups.

The presence of a peak between 3100 cm^{-1} and 3000 cm^{-1} indicates the presence of a vibrational asymmetric stretch of the C=C-H functional groups of alkenes and aromatic rings.

Hydrogen-bonded O-H groups could be responsible of the broad peak between 3600 cm^{-1} and 3100 cm^{-1} . Between 3000 cm^{-1} and 2800 cm^{-1} , there are peaks in the spectral region, which represent a symmetric and asymmetric stretch of H-C-H groups.

Between 1750 cm^{-1} and 1625 cm^{-1} there is a vibrational stretch of the C=O functional groups as depicted by the peak formed in this spectral region.

Between 1600 cm^{-1} and 1500 cm^{-1} , there is a peak in the spectral region representing a vibrational stretch of N=O groups.

The presence of peaks between 1300 cm^{-1} and 1000 cm^{-1} relates to a vibrational stretch of the C-O functional groups.

3.2 Adsorption behavior of ESM and SDS-ESM

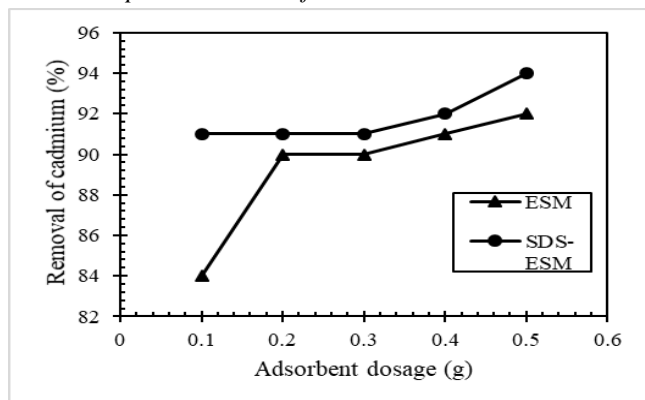


Fig. 1. Adsorption of cadmium at various adsorbents dosages.

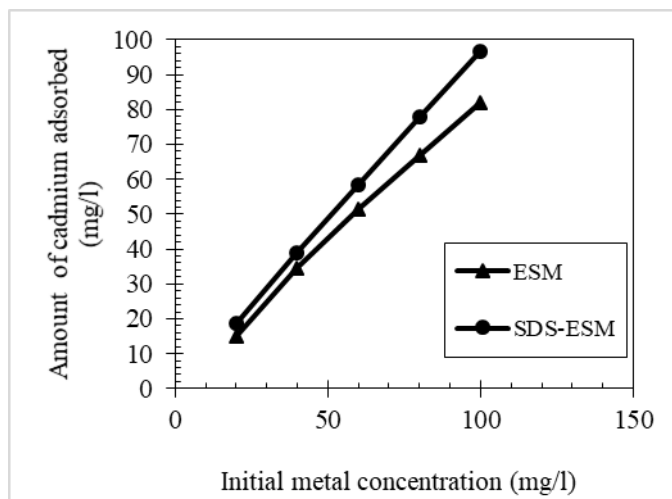


Fig. 2. Adsorption pattern at various initial cadmium concentrations

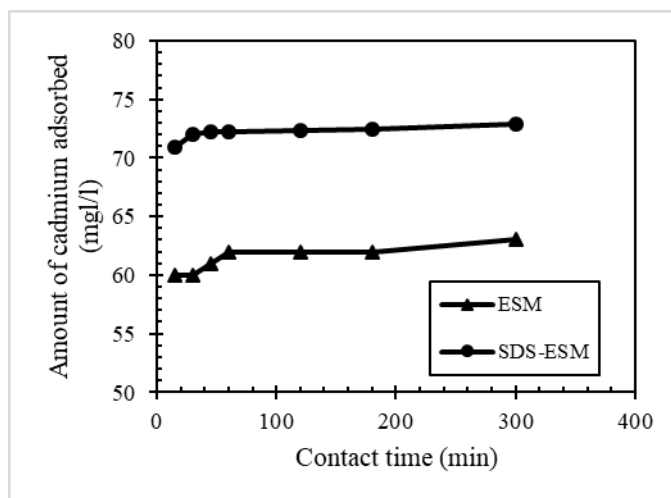


Fig. 3. Adsorption of cadmium at various exposure times

The percentage of cadmium adsorbed is plotted as a function of the amount of adsorbent dosage, as shown in Figure 1. When the biosorbent was impregnated with SDS, less biosorbent was required to remove an efficient amount of cadmium. The optimal adsorbent dosage of ESM was therefore chosen as 0.2 g and the optimal dosage of SDS-ESM was chosen as 0.1 g.

The amount of cadmium adsorbed is plotted against the initial cadmium concentration of the metal solutions, as shown in Figure 2. There exists a linear increase in the adsorption of cadmium ions when the initial cadmium concentration was increased. The maximum amount of cadmium was adsorbed at 100 mg/L using ESM and SDS-ESM. Furthermore, the SDS-ESM delivered a higher adsorption affinity to cadmium ions than the ESM, as shown in Figure 2.

The amount of cadmium adsorbed is plotted against the contact time of adsorption in the metal solutions, as shown in Figure 3. Impregnated eggshell membranes displayed a higher adsorption affinity than that of the non-treated eggshell membranes. The equilibrium concentration of cadmium ions

was reached at 62 mg/L after 60 minutes, when ESM was used as the biosorbent. The equilibrium concentration was therefore reached after 60 min ESM dosage of 0.2 g. Furthermore, the equilibrium concentration of cadmium ions was reached at 72 mg/L after 30 minutes, when SDS-ESM was used as the biosorbent. The equilibrium concentration was reached faster using SDS-ESM.

The coefficient of determination (R^2) of the different SDS-ESM models, as shown in Tables 1 and 2, indicates that the Freundlich isotherm model describes the adsorption behaviour the best. The Freundlich model is not limited to monolayer adsorption and assumes a heterogeneous system [37, 42-49].

After comparing the correlation coefficient of the different ESM models, it was found that the Langmuir isotherm model describes the adsorption behaviour of ESM the best. The Langmuir isotherm assumes a monolayer surface coverage and homogeneous adsorption [38, 42-49]. Whereas the Freundlich adsorption capacity constant (k_f) is significantly higher for the SDS-ESM impregnated eggshell membranes.

The coefficient of determination (R^2) of the different kinetic models, as shown in Tables 3 and 4, indicates that the pseudo-second-order model describes the kinetic data of both biosorbents the best.

TABLE IV PARAMETERS OF THE PSEUDO-SECOND-ORDER KINETIC MODEL

Kinetic Model	Parameters	Cd(II)	
		ESM	SDS-ESM
Pseudo Second Order	q_e (mg/g)	16.55629	20.40816
	k_2 (g/mg.min-1)	0.010131	0.022932
	R^2	0.9997	0.9999

The surface water of the Mooi River contains many essential trace elements. As shown in Figure 4, both ESM and SDS-ESM had the ability to adsorb cadmium in the presence of other heavy metals. Again, SDS-ESM delivered a higher adsorption efficiency for cadmium than ESM. SDS-ESM removed 83.45% of cadmium, while ESM removed 62.10% of cadmium.

TABLE I
PARAMETERS OF THE LANGMUIR ISOTHERM MODEL

Isotherm Model	Parameters	Cd(II)	
		ESM	SDS-ESM
Langmuir	q_m (mg/g)	50.25126	250
	k_L (l/mg)	0.038521	0.075901
	R^2	0.9946	0.2092

TABLE II
PARAMETERS OF THE FREUNDLICH ISOTHERM MODEL

Isotherm Model	Parameters	Cd(II)	
		ESM	SDS-ESM
Freundlich	k_f (mg/g)	2.583449639	17.74598
	n	1.380834024	1.126761
	R^2	0.9943	0.9058

TABLE III PARAMETERS OF THE PSEUDO-FIRST-ORDER KINETIC MODEL

Kinetic Model	Parameters	Cd(II)	
		ESM	SDS-ESM
Pseudo First Order	q_e (mg/g)	15.5	36
	k_1 (min^{-1})	0.0006909	0.0006909
	R^2	0.7241	0.5677

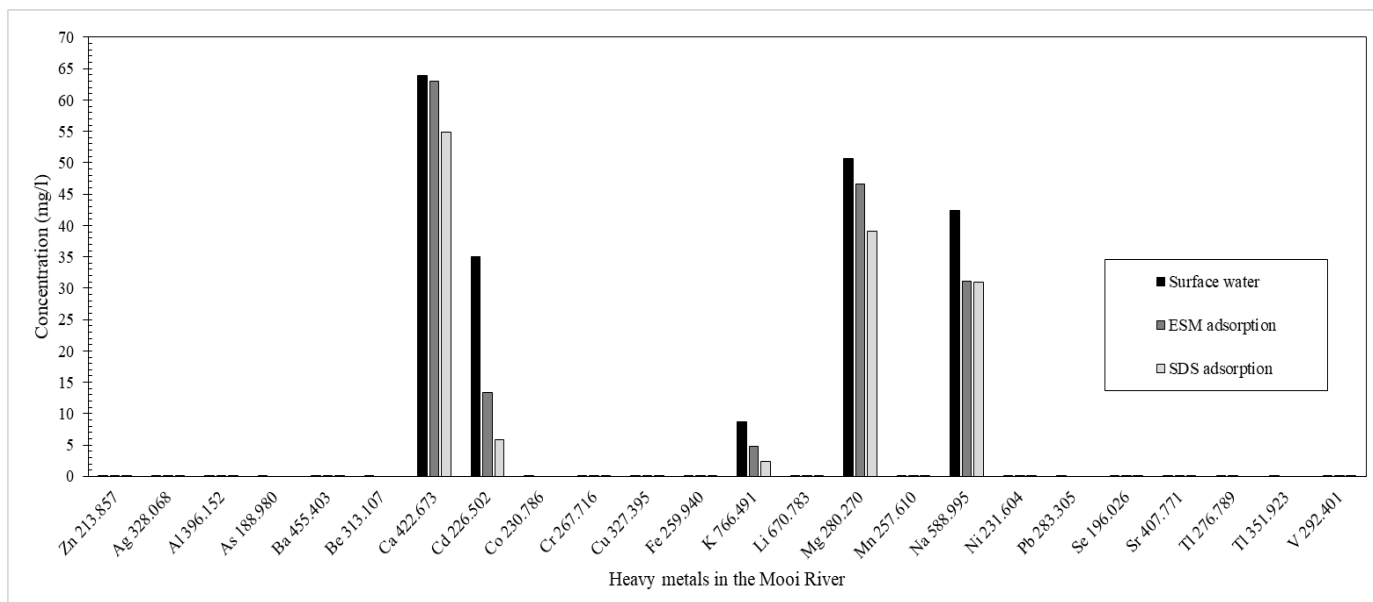


Fig. 4. Trace element concentrations in the Mooi River

IV. CONCLUSION

Successful surfactant impregnation of the eggshell membranes was established after comparing the FTIR results and characteristics of the different biosorbents.

After comparing the variation in adsorbent dosage, it was evident that the SDS-ESM delivered a higher cadmium removal efficiency than ESM. The optimal dosage for ESM and SDS-ESM was 0.2 g and 0.1 g, respectively.

The equilibrium concentration of cadmium (II) ions was reached at 62 mg/L after 60 minutes, when ESM was used as biosorbent, and the equilibrium concentration of cadmium ions was reached at 72 mg/L after 30 minutes, when SDS-ESM was used as the biosorbent. Thus, the equilibrium concentration was reached faster when SDS-ESM was used.

Both ESM and SDS-ESM had the ability to adsorb cadmium, in the presence of other heavy metals, from the surface water of the Mooi River. SDS-ESM had a higher adsorption efficiency than ESM.

The Langmuir isotherm model describes the adsorption behaviour of ESM and therefore, ESM has a monolayer surface coverage and applied homogeneous adsorption.

The Freundlich isotherm model describes the adsorption behaviour of SDS-ESM and therefore, SDS-ESM applied multilayer adsorption that occurred on a heterogeneous surface and it was not limited to monolayer adsorption.

The pseudo-second order kinetic was found to fit best the adsorption of cadmium by both ESM and SDS-ESM implying that the adsorption occurred mainly through chemisorption process.

It is thus clear that natural and SDS impregnated eggshell membranes have the ability of serving as adsorbents during the adsorption of heavy metals such as cadmium. These biosorbents, especially SDS-ESM, can therefore be used in the purification of surface water of essential water resources.

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REFERNECES

- [1] Abbas, A., Al-Amer, A.M., Laoui, T., Al-Marri, M.J., Nasser, M.S., Khraisheh, M. & Atieh, M.A., 2016. Heavy metal removal from aqueous solution by advanced carbon nanotubes: critical review of adsorption applications. *Separation and Purification Technology*, 157: 141-161. <https://doi.org/10.1016/j.seppur.2015.11.039>
- [2] John, O.O. 2017. Evaluation of the effect of Benzene-contaminated water on the Gravimetry, Serum Cholesterol and Triglyceride concentration of Africa Catfish (*Clarias Gariepinus*). *Mintage Journal of Pharmaceutical and Medical Sciences*: 12-14.
- [3] Voulvoulis, N. & Georges, K. 2016. Industrial and agricultural sources and pathways of aquatic pollution. In *Impact of water pollution on human health and environmental sustainability*: 29-54.
- [4] Al-Ghouti, M.A. & Salih, N.R. 2018. Application of eggshell wastes for boron remediation from water. *Journal of Molecular Liquids*, 256: 599-610. <https://doi.org/10.1016/j.molliq.2018.02.074>
- [5] E. Fosso-Kankeu, A. Mulaba-Bafubiandi, B.B. Mamba, T.G. Barnard, Mitigation of Ca, Fe, and Mg loads in surface waters around mining areas using indigenous microorganism strains. *Journal of Physics and Chemistry of the Earth*, Vol. 34, pp. 825-829, 2009. <https://doi.org/10.1016/j.pce.2009.07.005>
- [6] E. Fosso-Kankeu, A. Mulaba-Bafubiandi, B.B. Mamba, L. Marjanovic, T.G. Barnard, A comprehensive study of physical and physiological parameters that affect biosorption of metal pollutants from aqueous solutions. *Journal of Physics and Chemistry of the Earth*, Vol. 35, pp. 672-678, 2010. <https://doi.org/10.1016/j.pce.2010.07.008>
- [7] E. Fosso-Kankeu, F. Waanders, M. Reitz, Selective adsorption of heavy and light metals by natural zeolites. 6th International Conference on Green Technology, Renewable Energy and Environmental Engineering (ICGTREEE'2014). 27-28 November 2014, Cape Town-South Africa. (**Award Winning Paper**). Editors: Muzenda E. and Sandhu S. ISBN: 978-93-84468-08-8. Pp 271-274. 2014.
- [8] E. Fosso-Kankeu, F. Waanders, C. Fraser, Bentonite clay adsorption affinity for anionic and cationic dyes. 6th International Conference on Green Technology, Renewable Energy and Environmental Engineering (ICGTREEE'2014). 27-28 November 2014, Cape Town-South Africa.

- Editors: Muzenda E. and Sandhu S. ISBN: 978-93-84468-08-8. Pp 257-260. Pp 257-260. 2014.
- [9] E. Fosso-Kankeu, C.M. Van der Berg, F.B. Waanders, Physico-chemical activation of South African bentonite clay and impact on metal adsorption capacity. 6th International Conference on Green Technology, Renewable Energy and Environmental Engineering (ICGTREEE'2014). 27-28 November 2014, Cape Town-South Africa. Editors: Muzenda E. and Sandhu S. ISBN: 978-93-84468-08-8. Pp 247-252. 2014.
- [10] E. Fosso-Kankeu; O. Ntwampe, F. Waanders, and A. Webster, The Performance of Polyaluminium Chloride and Bentonite clay Coagulant in the Removal of Cationic and Anionic Dyes. 7th International Conference on Latest Trends in Engineering and Technology (ICLTET' 2015), November 26-27, 2015 Irene, Pretoria (South Africa). Editors: E. Muzenda and T Yingthawornsuk. ISBN: 978-93-84422-58-5. 2015.
- [11] E. Fosso-Kankeu, A. Manyatshe, D. van der Berg, N. Lemmer, F. Waanders, and H. Tutu, Contaminants in Sediments across the Mooi and Vaal Rivers Network in The Vicinity of Potchefstroom. 7th International Conference on Latest Trends in Engineering and Technology (ICLTET' 2015), November 26-27, 2015 Irene, Pretoria (South Africa). Editors: E. Muzenda and T Yingthawornsuk. ISBN: 978-93-84422-58-5. 2015.
- [12] E. Fosso-Kankeu, F. Waanders, and C. Laurette Fourie, Surfactant Impregnated Bentonite Clay for the Adsorption of Anionic Dyes. 7th International Conference on Latest Trends in Engineering and Technology (ICLTET' 2015), November 26-27, 2015 Irene, Pretoria (South Africa). Editors: E. Muzenda and T Yingthawornsuk. ISBN: 978-93-84422-58-5. 2015.
- [13] Liu, J., Yang, H., Gosling, S.N., Kumm, M., Flörke, M., Pfister, S., Hanasaki, N., Wada, Y., Zhang, X., Zheng, C. & Alcamo, J. 2017. Water scarcity assessments in the past, present, and future. *Earth's Future*, 5(6): 545-559. <https://doi.org/10.1002/2016EF000518>
- [14] Chowdhury, S., Mazumder, M.J., Al-Attas, O. & Husain, T. 2016. Heavy metals in drinking water: Occurrences, implications, and future needs in developing countries. *Science of the total Environment*, 569: 476-488. <https://doi.org/10.1016/j.scitotenv.2016.06.166>
- [15] Zeitoun, M.M. & Mehana, E.E. 2014. Impact of water pollution with heavy metals on fish health: overview and updates. *Global Veterinaria*, 12(2): 219-231.
- [16] Homer-Dixon, T. & Percival, V. 2018. The Case of South Africa. *Environmental Conflict*, 1: 13-35).
- [17] Manyatshe, A., Fosso-Kankeu, E., Van der Berg, D., Lemmer, N., Waanders, F. & Tutu, H. 2017. Dispersion of inorganic contaminants in surface water in the vicinity of Potchefstroom. *Physics and Chemistry of the Earth, Parts A/B/C*, 100: 86-93. <https://doi.org/10.1016/j.pce.2017.04.008>
- [18] Kurniawan, T.A., Chan, G.Y., Lo, W.H. & Babel, S. 2006. Physico-chemical treatment techniques for wastewater laden with heavy metals. *Chemical engineering journal*, 118(1-2): 83-98. <https://doi.org/10.1016/j.cej.2006.01.015>
- [19] E. Fosso-Kankeu, H. Mittal, F. Waanders, S.S. Ray, Performance of synthesized hybrid hydrogel nanocomposite applied for the removal of metal ions from aqueous solutions. In: Drebenstedt, C. & Paul, M.: IMWA 2016 – Mining Meets Water – Conflicts and Solutions. – p. 850 – 857; Freiberg/Germany (TU Bergakademie Freiberg). 2016.
- [20] E. Fosso-Kankeu, F. Waanders, E. Maloy, B. Steyn, Reduction of salinity and hardness of water using copolymerized biopolymers. In: Drebenstedt, C. & Paul, M.: IMWA 2016 – Mining Meets Water – Conflicts and Solutions. – p. 994 – 1001; Freiberg/Germany (TU Bergakademie Freiberg). 2016.
- [21] E. Fosso-Kankeu, F. Waanders, J. Potgieter, Enhanced adsorption capacity of sweet sorghum derived biochar towards malachite green dye using bentonite clay. International Conference on Advances in Science, Engineering, Technology and Natural Resources (ICASETNR-16) Nov. 24-25, 2016, Parys – South Africa. ISBN: 978-93-84468-79-8. 2016.
- [22] C. Mampho, S. Pandey, J. Ramontja, E. Fosso-Kankeu, F. Waanders, Synthesis and characterization of superadsorbent hydrogels based on natural polymers kappa carrageenan. International Conference on Advances in Science, Engineering, Technology and Natural Resources (ICASETNR-16) Nov. 24-25, 2016, Parys – South Africa. ISBN: 978-93-84468-79-8. 2016.
- [23] Gunatilake, S.K., 2015. Methods of removing heavy metals from industrial wastewater. *Methods*, 1(1).
- [24] Sounthararajah, D.P., Loganathan, P., Kandasamy, J. & Vigneswaran, S. 2015. Adsorptive removal of heavy metals from water using sodium titanate nanofibres loaded onto GAC in fixed-bed columns. *Journal of hazardous materials*, 287: 306-316. <https://doi.org/10.1016/j.jhazmat.2015.01.067>
- [25] E. Fosso-Kankeu, A.F. Mulaba-Bafubiandi, B.B. Mamba and T.G. Barnard, Prediction of metal-adsorption behaviour in the remediation of water contamination using indigenous microorganisms. *Journal of Environmental Management*. Vol. 92, no. 10, pp. 2786-2793, 2011. <https://doi.org/10.1016/j.jenvman.2011.06.025>
- [26] H. Mittal, E. Fosso-Kankeu, Shivani B. Mishra, Ajay K. Mishra, Biosorption potential of Gum ghatti-g-poly (acrylic acid) and susceptibility to biodegradation by *B. subtilis*. *International Journal of Biological Macromolecules*. Vol. 62, pp. 370-378, 2013. <https://doi.org/10.1016/j.ijbiomac.2013.09.023>
- [27] E. Fosso-Kankeu, A.F. Mulaba-Bafubiandi, T.G. Barnard, Establishing suitable conditions for metals recovery from metal saturated Bacillaceae bacterium using experimental design. *International Biodeterioration and Biodegradation*. Vol. 86, pp. 218-224, 2014. <https://doi.org/10.1016/j.ibiod.2013.09.022>
- [28] E. Fosso-Kankeu, A.F. Mulaba-Bafubiandi, Challenges in the escalation of metal-biosorbing processes for water treatment: applied and commercialized technologies. *African Journal of Biotechnology*. Vol. 13, no. 17, pp. 1756-1771, 2014. <https://doi.org/10.5897/AJB2013.13311>
- [29] E. Fosso-Kankeu, A.F. Mulaba-Bafubiandi, Implication of plants and microbial metalloproteins in the bioremediation of polluted waters. *Journal of Physics and Chemistry of the Earth*. Vol. 67-69, 242-252, 2014. <https://doi.org/10.1016/j.pce.2013.09.018>
- [30] Feng, D., Aldrich, C. & Tan, H. 2000. Treatment of acid mine water by use of heavy metal precipitation and ion exchange. *Minerals Engineering*, 13(6): 623-642. [https://doi.org/10.1016/S0892-6875\(00\)00045-5](https://doi.org/10.1016/S0892-6875(00)00045-5)
- [31] Fernandez, M.E., Nunell, G.V., Bonelli, P.R. & Cukierman, A.L. 2014. Activated carbon developed from orange peels: Batch and dynamic competitive adsorption of basic dyes. *Industrial Crops and Products*, 62: 437-445. <https://doi.org/10.1016/j.indcrop.2014.09.015>
- [32] Mittal, A., Teotia, M., Soni, R.K. & Mittal, J. 2016. Applications of egg shell and egg shell membrane as adsorbents: a review. *Journal of Molecular Liquids*, 223: 376-387. <https://doi.org/10.1016/j.molliq.2016.08.065>
- [33] Baláz, M. 2014. Eggshell membrane biomaterial as a platform for applications in materials science. *Acta biomaterialia*, 10(9): 3827-3843. <https://doi.org/10.1016/j.actbio.2014.03.020>
- [34] Anbia, M. & Amirmahmoodi, S. 2016. Removal of Hg (II) and Mn (II) from aqueous solution using nanoporous carbon impregnated with surfactants. *Arabian Journal of Chemistry*, 9: 319-325. <https://doi.org/10.1016/j.arabjc.2011.04.004>
- [35] Chatterjee, S., Lee, D.S., Lee, M.W. & Woo, S.H. 2009. Congo red adsorption from aqueous solutions by using chitosan hydrogel beads impregnated with nonionic or anionic surfactant. *Bioresource Technology*, 100(17): 3862-3868. <https://doi.org/10.1016/j.biortech.2009.03.023>
- [36] Atkins, P. & De Paula, J. 2013. Elements of physical chemistry. Oxford University Press, USA
- [37] Langmuir, I. 1916. The constitution and fundamental properties of solids and liquids. *Journal of the American Chemical Society*, 38: 2221-2295. <https://doi.org/10.1021/ja02268a002>
- [38] Fosso-Kankeu, E., Waanders, F. & Fourie, C.L. 2016. Adsorption of Congo Red by surfactant-impregnated bentonite clay. *Desalination and Water Treatment*, 57(57): 27663-27671. <https://doi.org/10.1080/19443994.2016.1177599>
- [39] Ghosal, P.S. & Gupta, A.K., 2017. Determination of thermodynamic parameters from Langmuir isotherm constant-revisited. *Journal of Molecular Liquids*, 225: 137-146. <https://doi.org/10.1016/j.molliq.2016.11.058>
- [40] Lagergren S. 1898. Zur theorie der sogenannten adsorption gelöster stoffe. *Kungliga Svenska Vetenskapsakademiens Handlingar*, 24(4):1-39.

- [41] Gosset, T., Trancart, J.L. & Thevenot, D.R. 1986. Batch metal removal by peat kinetics and thermodynamics. *Water Res.* 20: 21–6.
[https://doi.org/10.1016/0043-1354\(86\)90209-5](https://doi.org/10.1016/0043-1354(86)90209-5)
- [42] E. Fosso-Kankeu, H. Mittal, S.B. Mishra, A.K. Mishra, Gum ghatti and acrylic acid based biodegradable hydrogels for the effective adsorption of cationic dyes. *Journal of Industrial and Engineering Chemistry*. Vol. 22, pp. 171-178, 2015.
<https://doi.org/10.1016/j.jiec.2014.07.007>
- [43] E. Fosso-Kankeu, H. Mittal, F. Waanders, I.O. Ntwampe, S.S. Ray, Preparation and characterization of gum karaya hydrogel nanocomposite flocculant for metal ions removal from mine effluents. *International Journal of Environmental Science and Technology*. Vol. 13, pp. 711-724, 2016.
<https://doi.org/10.1007/s13762-015-0915-x>
- [44] E. Fosso-Kankeu, F. Waanders, E. Maloy, Copolymerization of ethyl acrylate onto guar gum for the adsorption of Mg(II) and Ca(II) ions. *Desalination and Water Treatment*. doi: 10.1080/19443994.2016.1165147: pp. 1-10, 2016.
- [45] E. Fosso-Kankeu, F. Waanders, C.L. Fourie, Adsorption of Congo Red by surfactant-impregnated bentonite clay. *Desalination and Water Treatment*. doi: 10.1080/19443994.2016.1177599: pp. 1-9, 2016.
- [46] E. Fosso-Kankeu, A.F. Mulaba-Bafubiandi, L.A. Piater, M.G. Tlou, Cloning of the *cnr* operon into a strain of Bacillaceae bacterium for the development of a suitable biosorbent. *World Journal of Microbiology and Biotechnology*. DOI 10.1007/s11274-016-2069-5. 2016.
- [47] E. Fosso-Kankeu, H. Mittal, F. Waanders, S.S. Ray, Thermodynamic properties and adsorption behaviour of hydrogel nanocomposites for cadmium removal from mine effluents. *Journal of Industrial and Engineering Chemistry*. Vol. 48, pp. 151-161, 2017.
<https://doi.org/10.1016/j.jiec.2016.12.033>
- [48] E. Fosso-Kankeu, F.B. Waanders, F.W. Steyn, Removal of Cr(VI) and Zn(II) from an aqueous solution using an organic-inorganic composite of bentonite-biochar-hematite. *Desalination and Water Treatment*. Vol. 59, pp. 144-153, 2017.
- [49] E. Fosso-Kankeu, 2018. Synthesized af-PFCl and GG-g-P(AN)/TEOS hydrogel composite used in hybridized technique applied for AMD treatment. *Journal of Physics and Chemistry of the Earth*. 2018.