

Title page

New insights into the role of microbial community composition in driving soil respiration rates

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1 **Abstract**

2 Microbial community plays critical roles in driving soil carbon (C) cycling in terrestrial
3 ecosystems. However, we lack empirical evidence to demonstrate the role of microbial
4 community in driving soil respiration -a key ecosystem process for global sustainability and
5 climate regulation. Here, we used a long-term field experiment including multiple
6 management practices, to identify, via statistical modeling, the role of microbial community
7 composition in influencing soil respiration. We analyzed major soil properties and microbial
8 (both bacterial and fungal) abundance, diversity and community composition. We found that
9 different management regimes led to different soil respiration rates. Most importantly,
10 microbial community composition explained a unique portion of the variation in soil
11 respiration, which cannot be accounted for by key respiration drivers such as soil properties
12 and other microbial attributes (richness and total abundance). Microbial biomass and fungal
13 richness were also identified as key drivers of soil respiration. Our results indicate that
14 inclusions of microbial compositional data in Earth system models can be potentially used to
15 improve our capacity to predict changes in soil C balance under changing environments.

16
17 **Key words**

18 Bacteria, fungi, carbon cycling, ecosystem processes, microbial community, global change

19 **1. Introduction**

20 Soils store four times as much carbon (C) as plant and atmospheric pools (Singh et al., 2010;
21 Karhu et al., 2014), and soil respiration releases about 60 Pg C annually from the land surface
22 (Shao et al., 2013). Both C sequestration and soil respiration are critical processes controlling
23 key ecosystem functions such as climate regulation, nutrient cycling and plant productivity
24 (Singh et al., 2010; Victoria et al., 2012; Trivedi et al., 2017). Global climate change and
25 human disturbances including intensive agricultural practices are increasing the amount of C
26 emitted to the atmosphere with important implications for the climate regulation of Earth
27 (Nazaries et al., 2015; Spohn et al., 2016). Because of this, predictions of soil C balance in
28 terrestrial ecosystems have become a global priority during the last decades with the
29 development of Earth system models as primary tools (Luo et al., 2016).

30 Soil respiration is driven by both biotic and abiotic factors (Walker et al., 2004; Monson
31 et al., 2006; Orwin et al., 2015). Previous studies have demonstrated the importance of
32 geographic location (Campbell et al., 2004; Whitaker et al., 2014), climate (temperature and
33 rainfall) (García-Palacios et al., 2012; Karhu et al., 2014), soil properties (Delgado-Baquerizo
34 et al., 2016a) and plant features (Raich and Tufekciogul, 2000; Knowles et al., 2015) as key
35 predictors of soil respiration. However, current models are not able to accurately predict the
36 variation in soil stocks and respirations, leading to a high level of uncertainty for these
37 predictions. Identifying new major predictors of soil respiration that allow the improvement
38 of predictive models is one of the major challenges that we are facing today. Most recently,
39 the inclusion of microbial processes (reflected by microbial biomass and enzyme activities)
40 has been reported to improve the prediction of soil C fluxes at the global scales (Allison et al.,
41 2010; Wieder et al., 2013). Similarly, microbial diversity has been reported to drive multiple
42 soil functions including soil respiration (Delgado-Baquerizo et al., 2016b; Liu et al., 2017).
43 Further, in addition to microbial biomass and diversity, other microbial parameters such as
44 community composition (relative abundance of phylotypes) may greatly improve our
45 prediction for soil respiration (Whitaker et al., 2014), given the strong positive relationships
46 between microbial composition and the functional genes that regulate soil respiration (Trivedi
47 et al., 2016). However, the importance of soil microbial community composition in driving
48 soil respiration remains poorly understood and largely unexplored; and no experimental

49 approach has been used to address this important gap of knowledge.

50 A growing number of studies had emphasized the significance of microbial community
51 composition in driving soil processes and functions including gaseous emission,
52 decomposition, nutrient cycling and plant production (Fierer et al., 2007; Peter et al., 2011;
53 Trivedi et al., 2013; Trivedi et al., 2016). We argue that current knowledge on microbial life
54 strategies and functional attributes can be used to markedly improve our prediction of soil
55 respiration rates. For example, previous studies suggested that ecological functional
56 categories of copiotrophs and oligotrophs have specific roles in utilizing soil organic C (SOC)
57 for respiration (Fierer et al., 2007; Ramirez et al., 2012; Trivedi et al., 2013). Thus,
58 copiotrophs such as Bacteroidetes, Alphaproteobacteria and Gammaproteobacteria are
59 expected to have higher respiration rates compared to oligotrophs including Actinobacteria,
60 Acidobacteria and Deltaproteobacteria (Fierer et al., 2007; Bastian et al., 2009; Singh et al.,
61 2010; Trivedi et al., 2013). In addition, bacteria are considered to have lower C use efficiency
62 compared to fungi (Austin et al., 2004; Waring et al., 2013). All these suggest that differences
63 in microbial community composition could potentially explain a unique portion of the
64 variation in soil respiration.

65 Herein, we posit that community composition of fungi and bacteria can potentially help
66 explain unique portions of the variation in soil respiration which cannot be accounted for by
67 other key drivers of soil respiration including soil properties, land management practices (i.e.
68 inorganic and organic fertilization) and other key microbial attributes such as microbial
69 biomass or community richness. To test our hypothesis, we collected soils from a long-term
70 field experiment including multiple combined applications of inorganic and organic fertilizers
71 including nitrogen (N), phosphorus (P) and potassium (K) (NPK), livestock manure, wheat
72 straw, and commercial organic fertilizers. These agricultural practices are well-known to
73 simultaneously modify soil abiotic properties, microbial biomass, and community richness
74 and composition (Sun et al., 2015; 2016), which in turn provides a unique opportunity to
75 empirically identify the role of relative importance of soil microbial community in driving
76 soil respiration responses to the combined fertilization after accounting for key soil
77 properties.

78 2. Materials and Methods

79 2.1 Experimental design and soil sampling

80 The long-term experiment was established in Linquan county, Anhui province, China
81 (33°04'58N, 115°13'42E), in October 2010. Mean annual temperature in this region is 15.3°C
82 and mean annual precipitation is 892 mm. The experimental plots (10 × 5 m in size) were
83 subject to wheat-corn rotation, and the locations were selected using a randomization
84 approach. The soil in this site belongs to a lime concretion black soil (Eutric Acrisols) (Zhang
85 et al., 2016), with 23% clay and 48% silt content. The initial pH of the soil was 5.72, which
86 had 0.73% organic C, 80.16 mg kg⁻¹ available N, 16.92 mg kg⁻¹ available P and 116.7 mg kg⁻¹
87 available K. This experiment included nine treatments with three replicate plots for each: (1)
88 control, no fertilization; (2) chemical NPK fertilizers application (NPK); (3) 50% NPK
89 fertilizers plus 6000 kg fresh cow manure ha⁻¹ y⁻¹ (NPK+CM); (4) 50% NPK fertilizers plus
90 6000 kg fresh pig manure ha⁻¹ y⁻¹ (NPK+PM); (5) NPK fertilizers plus all of preceding crop
91 wheat straw (NPK+ST); (6) 50% NPK fertilizers plus 6000 kg pig manure and wheat straw
92 from all of preceding crop (NPK+PM+ST); (7) 50% NPK fertilizers plus 6000 kg cow
93 manure and wheat straw from all of preceding crop (NPK+CM+ST); (8) 30% NPK fertilizers
94 plus 3600 kg commercial organic fertilizer ha⁻¹ y⁻¹(NPK+OCM), which is made of cow
95 manure; (9) 30% NPK fertilizers plus 3600 kg commercial organic fertilizer (NPK+OPM),
96 which is made of pig manure. The NPK fertilizer comprised urea (300 kg N ha⁻¹ y⁻¹),
97 superphosphate (120 kg P₂O₅ ha⁻¹ y⁻¹) and potassium chloride (100 kg K₂O ha⁻¹ y⁻¹). These
98 different proportions of NPK fertilizers with corresponding organic matter were applied to
99 manipulate the relative balance of nutrients in the soil for crop growing. All chemical
100 fertilizers and organic matter were applied once before sowing of the winter wheat in October,
101 and the quantities of nutrient yearly added to the plots is showed in Table S1. The wheat
102 straws were cut into small pieces of ~10 cm in length before use. Surface soil (0-15 cm) from
103 each plot was collected in June 2016 after the harvest of wheat (*Triticum* spp.). Soil samples
104 were sealed in plastic bags, and shipped back to the laboratory in an iced cooler. All the soil
105 samples were mixed homogenously, passed through a 2.0 mm sieve, followed by dividing
106 into two sub-samples. One sub-sample was stored at -20 °C for microbial analysis, and

107 another sub-sample was stored at 4°C for the analysis of soil properties.

108 2.2 Soil properties and respiration rate analysis

109 Soil water content was determined by oven-drying the samples at 105°C, and soil texture was
110 analyzed using the pipette method (Gee and Bauder,1986). Soil pH was measured using a
111 fresh soil to water ratio of 1: 2.5 with a Delta pH-meter, and soil organic carbon (SOC) was
112 determined using the K₂CrO₇ oxidation titration method (Walkley & Black 1934). Total
113 carbon (TC) and total nitrogen (TN) were measured on a LECO macro-CN analyzer (LECO,
114 St. Joseph, MI, USA). Inorganic N and labile carbon in the soils were extracted with 0.5 M
115 K₂SO₄ in a ratio of 1:5 by shaking at 200 rpm for 1 h and filtered through 0.45-µm Millipore
116 filter paper. Total C and N concentrations in the extracts were analyzed by TOC analyzer
117 with total nitrogen unit (TOC-L Analyzer, Shimadzu, Japan). In parallel, the carbon in
118 microbial biomass (MBC) was determined using the fumigation-extraction method (Vance et
119 al., 1987). For each measurement of respiration rate, approximately 10 g of fresh soil (within
120 48 h after sampling) was incubated in a 120 ml container at 25 °C for 24 h. At the end of this
121 period, CO₂ concentrations in headspace were measured using an Agilent-7890a gas
122 chromatograph equipped with a flame ionization detector (FID) and an electron capture
123 detector (ECD) (Agilent Technologies, Wilmington, DE, USA). Soil respiration rates were
124 calculated from the net accumulation of CO₂ over time.

125 2.3 Soil microbial community characterization

126 The total genomic DNA was extracted from 0.30 g of soil using the MoBio PowerSoil DNA
127 Isolation Kit (MoBio Laboratories, Carlsbad, CA, USA) following the manufacturer
128 instructions. The concentration and quality of isolated DNA was checked using a NanoDrop®
129 ND-2000c UV-Vis spectrophotometer (NanoDrop Technologies, Wilmington, DE, USA). The
130 abundance of bacteria and fungi were evaluated by quantifying *16S rRNA* and *ITS* gene copy
131 number on an iCycler iQ5 thermocycler (Biorad, USA) using the primer pairs
132 Eub338F/Eub518R (Cregger et al., 2012) and ITS1-5.8S (Fierer et al., 2005), respectively.

133 To evaluate the microbial community composition, the V4 region of the bacterial *16S*
134 *rRNA* gene and *ITS* of fungi were amplified using the primer pairs of 338F/806R (Liu et al.,

135 2016) and ITS1F /2043R (Zhao et al., 2016b), respectively. The 50 µl PCR reaction mixtures
136 consisted of 25 µl PremixTaq™ (Takara Biotechnology, Dalian, China), 1 µl of each primer
137 (10 µM), 3 µl of template DNA, and 20 µl of sterilized ddH₂O. The resultant PCR products
138 were purified using the Wizard® SV Gel and PCR Clean-Up System (Promega, San Luis
139 Obispo, USA). The purified amplicons were equimolarly mixed, and 2 × 250 bp paired-end
140 sequencing was carried out on an Illumina Miseq sequencer (Illumina Inc., San Diego, USA).
141 Raw reads generated from the MiSeq paired-end sequencing were merged together using the
142 Fast Length Adjustment of Short reads (FLASH). A chimera filtering approach UPARSE was
143 employed as the Operational Taxonomic Unit (OTU) picking strategy at 97% sequence
144 similarity. Representative sequences from individual OTUs generated in UPARSE were
145 processed using the Quantitative Insights into Microbial Ecology (QIIME) pipeline. The
146 resultant OTU map file was converted into a biom file for diversity calculation in QIIME. A
147 resampling procedure was conducted at a depth of 51,438 sequences for bacteria and 52,561
148 for fungi per sample before diversity calculation. Bacterial and fungal diversity index were
149 calculated based on 97 % OTU similarity of obtained bacterial and fungal sequences,
150 respectively. The taxonomic identity of each phylotype was determined using
151 the SILVA ribosomal RNA gene database project (Quast et al., 2013).

152 *2.4 Statistical analyses*

153 ANOVA was used to evaluate the effect of different fertilization treatments on the respiration
154 rate. Spearman's correlation analyses were performed to assess the relationships between soil
155 properties, respiration and bacterial and fungal community including abundance, diversity
156 and composition. The *16S rRNA* and *ITS* gene copy numbers were log-transformed prior to
157 statistical analysis to meet normality assumptions.

158 We conducted a classification random forest analysis (Breiman, 2001), as done in
159 Trivedi et al. (2016), to identify the major statistically significant microbial predictors of the
160 composition (relative abundance: number of sequences of major phyla/class level) of bacteria
161 and fungi. These analyses were conducted using the rfPermute package (Archer, 2016) of the
162 R statistical software (<http://cran.r-project.org/>). The random forest model determined the
163 importance of each predictor variable via evaluating the decrease in prediction accuracy (i.e.

164 increase in the mean square error between observations and OOB predictions) when the data
165 for that predictor are randomly permuted, as previously described (Delgado-Baquerizo et al.,
166 2015).

167 To examine the importance of microbial community attributes for soil respiration
168 compared with soil properties, we used a multi-model inference approach based on
169 information theory and non-parametric distance based linear regressions (DISTLM;
170 (McArdle and Anderson, 2001). In particular, this analysis provided insights on whether the
171 microbial community composition in our models provides additional predictive strength for
172 soil respiration after accounting for other important soil factors. The phyla/classes included in
173 the models were statistically significant predictors of the respiration rate from classification
174 random forest analysis (Table 1). The Euclidean distance was used as the measure of
175 dissimilarity in soil respiration between pairs of samples. We ranked best-fitting models that
176 could be generated with our independent variables according to the second-order Akaike
177 information criterion (AICc). The lower the AICc index the better the model. Here, we
178 consider a $\Delta AICc > 2$ threshold to differentiate between two substantially different models
179 and then select the best of those models (Burnham and Anderson, 2003). Then, we compared
180 the AICc of the best model which presumably includes soil properties and microbial
181 community to that of the corresponding models without microbial community composition.
182 Differences < 2.0 in AICc between alternative models indicate that they are approximately
183 equivalent in explanatory power (Burnham and Anderson, 2003). We conducted
184 distance-based multimodel inference with the PRIMER V6 statistical package for Windows
185 (PRIMER-E Ltd., Plymouth Marine Laboratory, UK).

186 We used structural equation model (SEM) to evaluate the direct and indirect effect of the
187 microbial community and soil properties on soil respiration in response to different
188 fertilization treatments. The SEM allows us to partition causal influences among multiple
189 variables, and to separate the direct and indirect effects of the predictors included in the
190 model; ultimately providing mechanistic information on the drivers of soil respiration (Grace,
191 2006). We established an *a priori* model according to our current knowledge of abiotic and
192 biotic impact on soil respiration (Fig. S1). We only included those variables that were
193 identified as statistically significant predictors of the respiration rate from random forest

194 analysis and distance-based best-fitting model. The data matrix was fitted to the model using
195 the maximum-likelihood estimation method. There is no single universally accepted test of
196 overall goodness of fit for SEM. Thus, we used the Chi-square test (χ^2 ; the model has a good
197 fit when $0 \leq \chi^2/\text{d.o.f} \leq 2$ and $0.05 < P \leq 1.00$) and the root mean square error of
198 approximation (RMSEA; the model has a good fit when $0 \leq \text{RMSEA} \leq 0.05$ and
199 $0.10 < P \leq 1.00$ (Schermele-Engel et al., 2003). We also calculated the standardized total
200 effects of microbial predictors, soil properties, and fertilization treatments on the soil
201 respiration. The net influence that one variable has upon another is calculated by summing all
202 direct and indirect pathways between the two variables. If the SEM fits the data well, the total
203 effect should approximate to the bivariate correlation coefficient for that pair of variables
204 (Grace 2006). The SEM analyses were performed using AMOS 21.0 (SPSS Inc., Chicago, IL,
205 USA).

206 **3. Results**

207 *3.1 Soil respiration rates and its relationship with soil properties*

208 As expected, different fertilization regimes led to different levels of soil respiration rate (Fig.
209 1, $P = 0.001$). The highest soil respiration rate was found in the treatment with
210 NPK+PM+STR (Fig. 1). Across all treatments, soil respiration rate was related to soil pH and
211 SOC (Table 2, $P < 0.05$). These soil properties also varied in response to the different
212 fertilization treatments (Table 3).

213 *3.2 Unique role of soil microbial community composition in predicting soil respiration rates*

214 Using random forest modeling, we identified the major bacterial and fungal phyla/classes for
215 predicting soil respiration (Fig. 2, $P < 0.05$). These taxa include several bacterial phyla such
216 as Alphaproteobacteria, WPS.2 and Deltaproteobacteria. Soil respiration rate was highly
217 correlated to other microbial attributes too (Table 4, $P < 0.05$). Particularly, there were
218 significant relationships between the respiration and microbial biomass, bacterial abundance,
219 bacterial community and richness, and fungal community and richness.

220 The best-fitting distance-based model accounted for 78% of the variation in soil
221 respiration and included both soil properties (C and C: N ratio) and microbial attributes
222 (community composition and fungal richness, Table 5). Most importantly, our model

223 provided evidence that microbial community composition accounted for a unique portion of
224 the variation in soil respiration that cannot be accounted for by soil properties, or other key
225 microbial attributes such as microbial biomass, abundance, and diversity. Thus, all models
226 excluding microbial composition showed higher AICc values ($AIC > 6.26$). The best-fitting
227 model first identified Alphaproteobacteria and Bacteroidetes as major microbial taxa
228 predicting soil respiration (Table 5). Our models further suggest that, besides microbial
229 composition, other microbial attributes such as microbial biomass carbon and fungal richness
230 were also important factors determining soil respiration, as they were included in the
231 best-fitting models.

232 *3.3 Mechanistic understanding on the role of microbial composition in driving soil* 233 *respiration rates*

234 We used the SEM to account for direct and indirect effects of management practices, soil
235 properties and microbial attributes on driving soil respiration, and therefore, to obtain a
236 system-level mechanistic understanding on the drivers of soil respiration. Our SEM explained
237 92% of the variance in the soil respiration, parameterized using predictors from the
238 best-fitting distance-based model. Soil properties, microbial composition, and fungal richness
239 have strong direct effects on soil respiration (Fig. 3a). However, fertilization treatment mostly
240 had indirect impacts on soil respiration through soil properties, microbial community
241 composition (i.e. Alphaproteobacteria and Bacteroidetes) and fungal richness. Overall, the
242 most important microbial attributes controlling soil respiration rates were the relative
243 abundances of Alphaproteobacteria and Bacteroidetes, and fungal richness (Fig. 3b).

244 **4. Discussion**

245 Our study provides direct experimental evidence that soil microbial community composition
246 plays a unique role in predicting soil respiration rates after accounting for key soil properties,
247 multiple fertilization treatments and other key microbial attributes such as microbial biomass
248 and richness. In particular, taxa from proposed copiotrophic Alphaproteobacteria and
249 Bacteroidetes (Fierer et al., 2007; Trivedi et al., 2013) were selected as the major microbial
250 predictors of soil respiration according to the best-fitting models. These fast-growing
251 copiotrophic organisms had been suggested to have higher C use efficiency than other taxa
252 belonging to slow-growing oligotrophic organisms (Fierer et al., 2007; Ramirez et al., 2012).

253 We thus identified their relative importance in driving soil respiration compared to other key
254 drivers of respiration. Our findings also provide evidence from an experimental approach that
255 the inclusion of microbial community composition in Earth system model could potentially
256 improve our capacity to predict C feedbacks (e.g soil C respiration) in terrestrial ecosystems.

257 Here, we demonstrated that changes in microbial community composition can help
258 explain a unique proportion of the variation in soil respiration. This interesting result implies
259 that major soil properties such as pH, C and C: N ratio, which are well-known to drive both
260 soil respiration rates and microbial community composition (Table 2, Table 6) (Fierer and
261 Jackson, 2006; Lauber et al., 2009; Maestre et al., 2015), were not able to account for the
262 variation in soil respiration explained by microbial community composition. The major
263 microbial predictors of soil respiration were identified through the distance-based
264 multi-model inference, a statistical approach which is specially recommended to identify key
265 predictors explaining a unique portion of the variation in a particular response variable that
266 cannot be explained by other predictors (DISTLM; McArdle & Anderson 2001). Our results
267 suggest a direct link between microbial community composition and soil respiration rates
268 potentially linked to particular microbial life strategies and functional capabilities (Trivedi et
269 al., 2016).

270 We identified major microbial taxa predicting soil respiration including several bacterial
271 phyla and fungal Ascomycota phylum based on the random forest model analysis (Fig. 2).
272 Further, both proposed taxa (Alphaproteobacteria and Bacteroidetes) from the best-fitting
273 model belong to fast-growing copiotrophs, which have been reported to be important drivers
274 of soil respiration rates (Fierer et al., 2007). We also used SEM to achieve a system-level
275 understanding on the role of microbial community composition in driving soil respiration
276 when considering human management, soil properties and other key microbial drivers
277 selected by our best-fitting distance-based models. Our SEM provided further evidence that
278 class Alphaproteobacteria had the strongest direct effect on soil respiration, which was also
279 modulated directly or indirectly via changes in fungal richness and land management. These
280 results suggest that drawing upon life-strategies, the relative abundance of key microbial taxa
281 at the highest taxonomic rank can be used to improve predictions in soil C respiration and
282 balance in terrestrial ecosystems. These results suggest that these taxa could potentially

283 influence soil C respiration rates under climate change and land use intensification scenarios
284 through shifting soil microbial community composition. In addition, significant positive
285 relationship of fungal richness with soil respiration rates (Fig. 3b) further suggests fungal role
286 in driving C cycling. These results support the growing literature that demonstrates the
287 important roles of fungal diversity in driving soil functions (Wagg et al., 2014;
288 Delgado-Baquerizo et al., 2016b). Our observations are also in accord with previous studies
289 suggesting that fungi have stronger potential in degrading available and complex forms of C
290 and N, attributing to their relatively strong hyphal growth form and enzymatic capacities (de
291 Boer et al., 2005). All of these observations are in line with recent studies suggesting the
292 significance of soil microbial community for controlling ecosystem sustainability and
293 multiple functions (Delgado-Baquerizo et al., 2016b; Liu et al., 2017).

294 Finally, our SEM demonstrates that fertilization treatments can indirectly drive soil
295 respiration through changing microbial community composition. Thus, any alteration of
296 dominant microbial taxa may result in subsequent changes in soil respiration, with
297 implications for the balance of terrestrial ecosystems. For example in this study, as one of
298 important microbial drivers of soil respiration, Bacteroidetes were suppressed by NPK
299 fertilization (Fig. 3a), which is a traditional agricultural practice for improving crop
300 productivity. Recently, application of organic matter has become a prevailing strategy for
301 improving soil fertility and mitigating soil degradation resulting from single chemical
302 fertilization (Reeves, 1997; Yadvinder-Singh et al., 2004; Zhao et al., 2016a). Chemical NPK
303 combined with commercial organic fertilizers (NPK+OCM) were also found to strongly
304 influence the richness of soil fungal communities, another major drivers of soil respiration in
305 our study (Fig. 3a). Overall, our results provide important information for land managers to
306 maintain C balance in terrestrial ecosystems.

307 Taken together, our study provides experimental evidence that microbial community
308 composition explained a unique portion of the variation in soil respiration which cannot be
309 accounted for by soil properties and human land management - the major drivers of soil
310 respiration. We further provide new insights into the role of key microbial taxa in driving soil
311 respiration, with implications for the prediction of this key soil process under global change
312 scenarios. These taxa could be potentially used to modulate soil respiration directly via soil

313 inoculum or indirectly via land management. Our findings have important implications for
314 improving our ability to predict soil C balance using Earth system models, and future studies
315 will need to test our hypothesis in observational datasets at the global scale.

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Table 1. Complete list of predictors used for the distance-based multi-modeling approach.

Predictor	Group	Variable	Abbreviature	Units
1	Soil properties	Soil water content	SWC	%
2		pH	pH	Unitless
3		Soil total carbon	TC	%
4		ratio soil C:N	C:N	Unitless
5		Soil organic carbon	SOC	g kg ⁻¹
6		Labile carbon	LC	mg kg ⁻¹
7	Biomass and diversity	Microbial biomass carbon	MBC	mg kg ⁻¹
8		Abundance of fungi (qPCR)	Fungi abundance	gene copies g ⁻¹
9		Abundance of bacteria (qPCR)	Bacteria abundance	gene copies g ⁻¹
10		Richness of fungi	Richness fungi	Number of OTUs
11		Richness of bacteria	Richness bacteria	Number of OTUs
12	Community composition	Fibrobacteres	Fibrobacteres	%
13		Bacteroidetes	Bacteroidetes	%
14		Acidobacteria	Acidobacteria	%
15		Cyanobacteria	Cyanobacteria	%
16		Actinobacteria	Actinobacteria	%
17		Betaproteobacteria	Betaproteobacteria	%
18		WS6	WS6	%
19		Latescibacteria (WS3)	Latescibacteria (WS3)	%
20		Deltaproteobacteria	Deltaproteobacteria	%
21		WPS.2	WPS.2	%
22		Alphaproteobacteria	Alphaproteobacteria	%

Table 2 Correlation coefficients (Spearman's ρ) between soil properties and soil respiration. SWC, soil water content; SOC, soil organic carbon; LC, labile organic carbon; *P* values below 0.05 are in bold.

Soil properties	ρ	<i>P</i>
SWC	0.083	0.681
pH	0.567	0.002
TC	-0.050	0.804
C:N	-0.026	0.899
SOC	0.388	0.046
Inorganic N	0.279	0.158
LC	0.166	0.408

Table 3 Soil basic properties from the different fertilization treatments. TC and TN, total carbon and total nitrogen, respectively; SOC, soil organic carbon; LC, labile carbon. Control, no fertilization; NPK, chemical NPK fertilizers; NPK+PM, NPK fertilizers plus fresh pig manure; NPK+CM, NPK fertilizers plus fresh cow manure; NPK+ST, NPK fertilizers plus wheat straw; NPK+PM+ST, NPK fertilizers plus fresh pig manure and wheat straw; NPK+CM+ST, NPK fertilizers plus fresh cow manure and wheat straw; NPK+OPM, NPK fertilizers plus commercial organic fertilizer that is made of pig manure; NPK+OCM, NPK fertilizers plus commercial organic fertilizer that is made of cow manure. Variables do not share the same letter are significantly different ($P < 0.05$)

Treatments	pH	TN (%)	TC (%)	C:N	SOC (g kg ⁻¹)	Inorganic N (mg kg ⁻¹)	LC (mg kg ⁻¹)
Control	5.86 ± 0.03a	0.13 ± 0.01a	1.23 ± 0.03a	9.73 ± 0.02a	7.31 ± 0.65a	1.99 ± 2.21a	65.63 ± 3.46a
NPK	5.03 ± 0.02b	0.16 ± 0.01b	1.68 ± 0.02b	10.65 ± 0.06b	7.30 ± 0.40a	30.98 ± 1.67b	78.21 ± 2.10b
NPK+PM	5.74 ± 0.04c	0.17 ± 0.02b	1.79 ± 0.01c	10.38 ± 0.07b	10.20 ± 1.20b	57.69 ± 17.90c	73.80 ± 3.82b
NPK+CM	5.85 ± 0.02a	0.16 ± 0.03b	1.46 ± 0.02d	9.16 ± 0.05c	14.17 ± 2.21b	25.35 ± 7.50b	82.42 ± 7.31b
NPK+STR	5.07 ± 0.01d	0.15 ± 0.03b	1.58 ± 0.01e	10.36 ± 0.10b	10.97 ± 1.30b	17.00 ± 2.38d	87.74 ± 5.68b
NPK+PM+STR	5.31 ± 0.03e	0.16 ± 0.02b	2.22 ± 0.02f	14.31 ± 0.09d	12.40 ± 1.31b	27.44 ± 3.00bc	87.59 ± 2.42b
NPK+CM+STR	5.21 ± 0.05e	0.19 ± 0.01bc	1.74 ± 0.03c	9.22 ± 0.08c	10.78 ± 1.03ab	25.63 ± 1.09bc	87.28 ± 6.74b
NPK+OPM	4.75 ± 0.02f	0.24 ± 0.02c	2.78 ± 0.02g	11.77 ± 0.12e	8.43 ± 5.29ab	9.17 ± 1.41e	81.20 ± 3.99b
NPK+OCM	4.74 ± 0.02f	0.16 ± 0.03b	1.50 ± 0.03d	9.61 ± 0.04a	8.57 ± 3.46ab	14.13 ± 0.22d	86.09 ± 2.86b

Table 4 Correlation coefficients (Spearman's ρ) between soil microbial traits and soil respiration. MBC, microbial biomass carbon ($\mu\text{g g}^{-1}$). Bacterial and fungal abundance were calculated from 16S rRNA and ITS gene copies g^{-1} soil, respectively. NMDS, non-metric multidimensional scaling ordination derived from the on Bray-Curtis similarities depicting the bacterial and fungal community compositions. *P* values below 0.05 are in bold.

Microbial community	ρ	<i>P</i>
MBC	0.710	<0.001
Bacterial abundance	0.400	0.039
Fungal abundance	0.163	0.417
Bacterial NMDS 1	-0.543	0.003
Bacterial NMDS 2	-0.030	0.881
Fungal NMDS 1	-0.556	0.003
Fungal NMDS 2	-0.311	0.115
Bacterial richness	0.521	0.005
Fungal richness	0.470	0.013

Table 5 Best-fitting models including/excluding microbial community composition as predictors of soil respiration. Model A is the best-fitting model including microbial community composition. Model B is the same model as the best model A, but excludes microbial composition. Model C is the best model when microbial community is taken out of the equation, but microbial diversity and soil properties are freely included. Models are ranked by AICc. AICc measures the relative goodness of fit of a given model; the lower its value, the more likely the model to be correct. Δ AICc are difference between the AICc of each model and that of the best model, and Δ AICc >2 threshold indicates two different models.

Models	Microbial Composition	Microbial diversity	Soil properties	R ²	AICc	Δ AICc
A	Alphaproteobacteria + Bacteroidetes	Fungal richness	TC + C:N	0.780	-23.47	
B	Excluded	Fungal richness	TC + C:N	0.649	-17.21	6.26
C	Excluded	Fungal richness	pH + TC + C:N	0.714	-19.75	3.72

Table 6 Correlation coefficients (Spearman's ρ) between main microbial predictors and soil properties. *P* values below 0.05 are in bold.

	SWC	pH	TC	C:N	SOC	Inorganic N	LC
Fungal richness	-0.508	0.694	0.16	-0.21	0.322	0.436	-0.18
	.007	.000	0.425	0.304	0.102	0.023	0.376
Alphaproteobacteria	0.483	-0.499	.327	0.681	-0.18	-0.27	0.189
	.011	.008	0.096	0	0.377	0.174	0.346
Bacteroidetes	-0.35	0.459	-.030	-0.32	0.681	0.269	0.458
	0.077	0.016	0.882	0.101	0	0.174	0.016

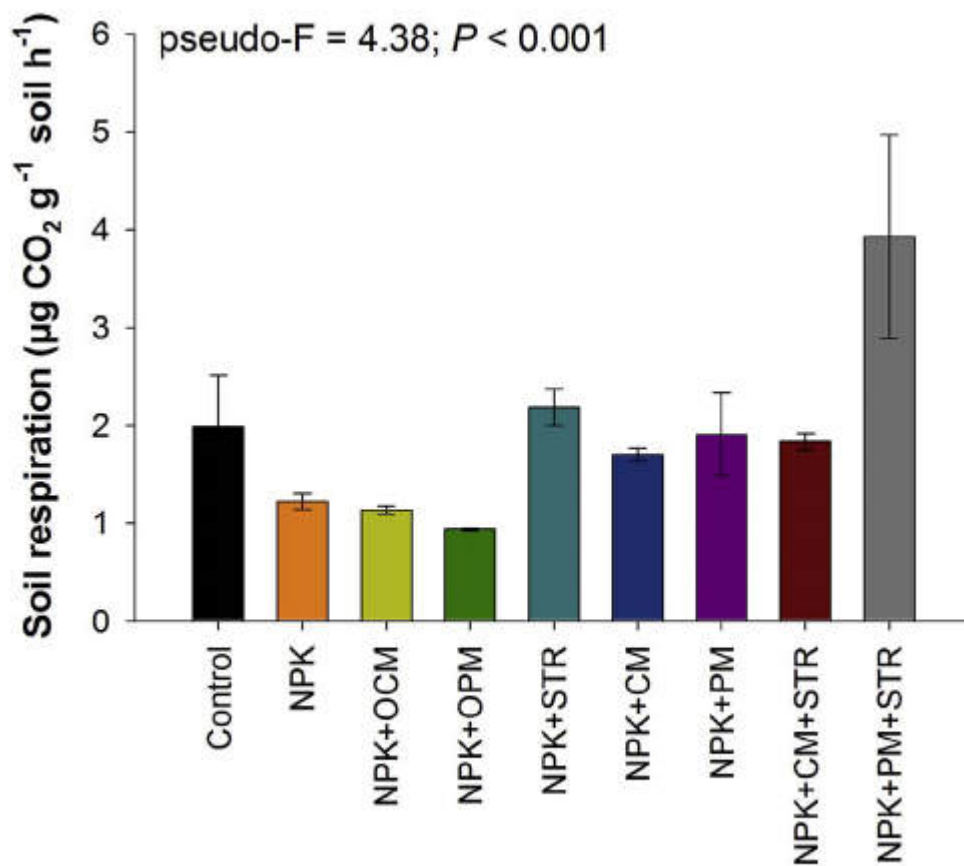


Fig. 1 Soil respiration rates under different treatment regimes. Control, no fertilization; NPK, chemical NPK fertilizers; NPK+PM, NPK fertilizers plus fresh pig manure; NPK+CM, NPK fertilizers plus fresh cow manure; NPK+ST, NPK fertilizers plus wheat straw; NPK+PM+ST, NPK fertilizers plus fresh pig manure and wheat straw; NPK+CM+ST, NPK fertilizers plus fresh cow manure and wheat straw; NPK+OPM, NPK fertilizers plus commercial organic fertilizer that is made of pig manure; NPK+OCM, NPK fertilizers plus commercial organic fertilizer that is made of cow manure.

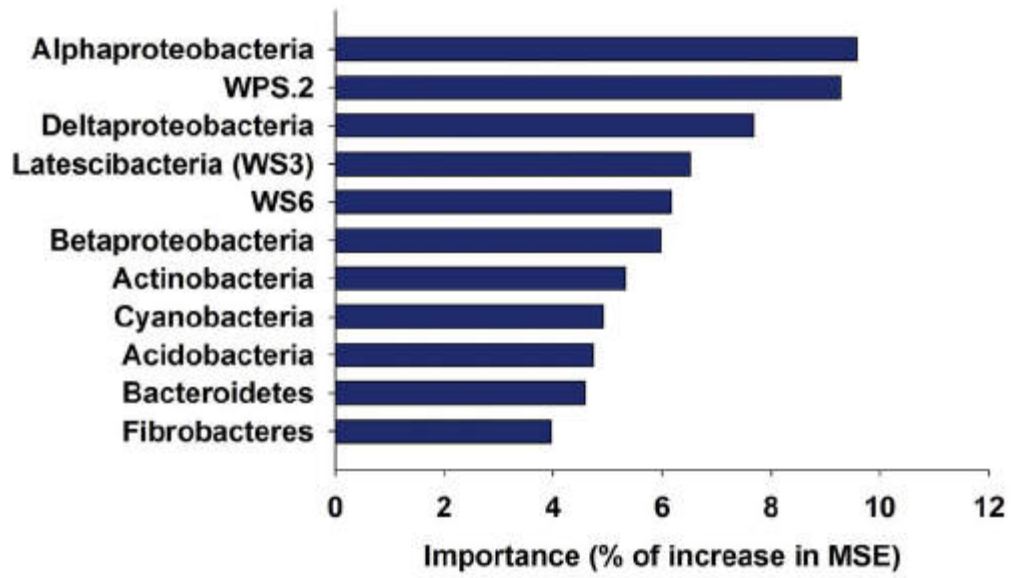


Fig. 2 Predictor importance (percentage of increase of mean square error, MSE) of major bacterial and fungal phyla/classes as drivers of soil respiration based on random forest model.

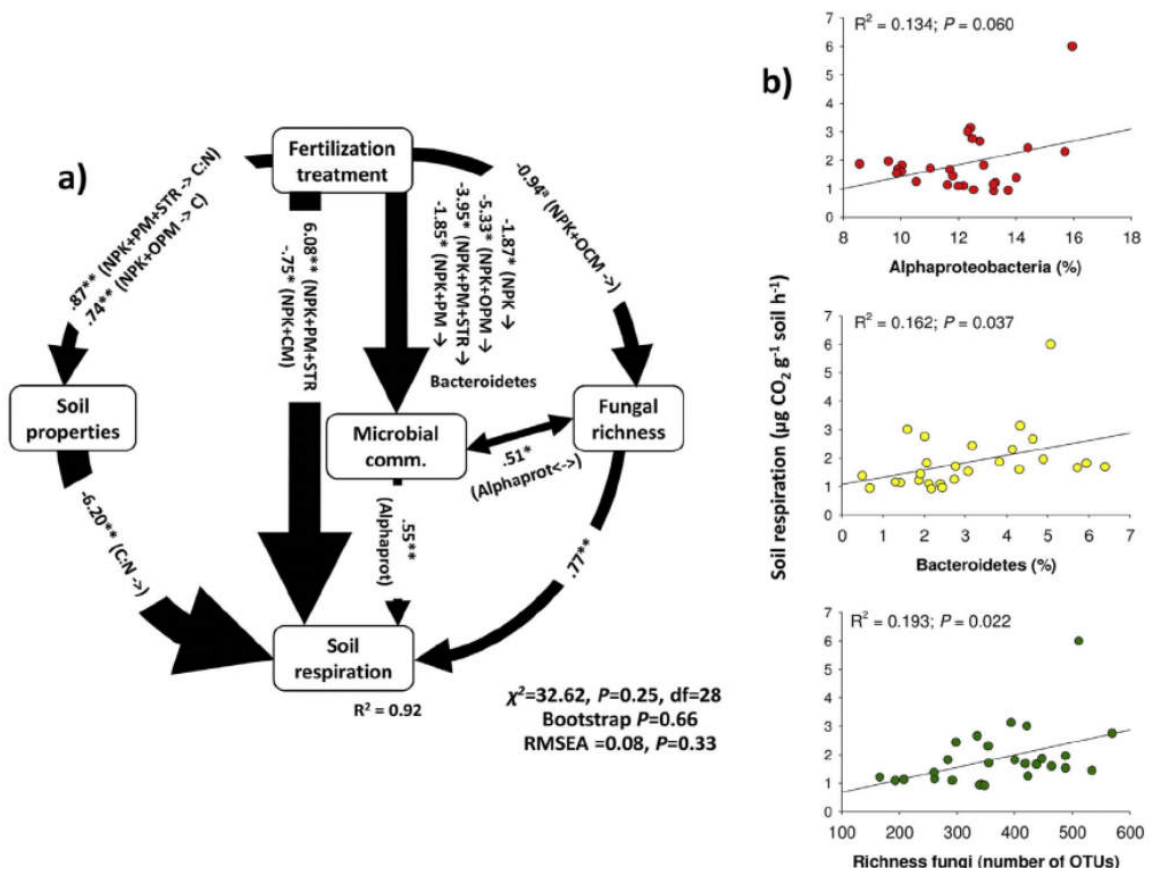


Fig. 3 (a) Structural equation modeling (SEM) showing effects of soil abiotic and biotic properties on soil respiration. Black lines and arrows indicate significant positive effect and negative effect, respectively. Numbers adjacent to arrows are path directions and coefficients, and width of the arrows is proportional to the strength of path coefficients. For simplicity, only the largest direct effects of fertilization on soil properties are shown. Minus represents negative effect of factors on soil respiration. Significance levels are as follows: ^a $P = 0.06$, $*P < 0.05$, $**P < 0.01$; (b) Relationships between the major microbial parameters and soil respiration rates

Supplementary Materials

New insights into the role of microbial community composition in driving soil respiration rates

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Table S1. The quantities of nutrients yearly added to the plots (10 × 5 m in size) with different treatments. Control, no fertilization; NPK, chemical NPK fertilizers; NPK+PM, NPK fertilizers plus fresh pig manure; NPK+CM, NPK fertilizers plus fresh cow manure; NPK+ST, NPK fertilizers plus wheat straw; NPK+PM+ST, NPK fertilizers plus fresh pig manure and wheat straw; NPK+CM+ST, NPK fertilizers plus fresh cow manure and wheat straw; NPK+OPM, NPK fertilizers plus commercial organic fertilizer that is made of pig manure; NPK+OCM, NPK fertilizers plus commercial organic fertilizer that is made of cow manure.

	Control	NPK	NPK+PM	NPK+CM	NPK+STR	NPK+PM+STR	NPK+CM+STR	NPK+OPM	NPK+OCM
Total N (kg)	0.00	2.75	1.90	1.61	2.90	2.05	1.76	1.65	1.65
Total P (kg)	0.00	0.48	0.35	0.29	0.48	0.35	0.29	0.36	0.29
Total K (kg)	0.00	0.46	0.30	0.26	0.90	0.74	0.70	0.29	0.44
Total C (kg)	0.00	0.00	10.95	11.19	16.11	27.06	27.31	7.87	8.81

Figure S1. *a priori* model showing effects of soil abiotic and biotic on soil respiration

